

Manuscript Details

Manuscript number	ENVPOL_2020_3590_R3
Title	Comparison of the impact of ships to size-segregated particle concentrations in two harbour cities of northern Adriatic Sea
Article type	Research Paper

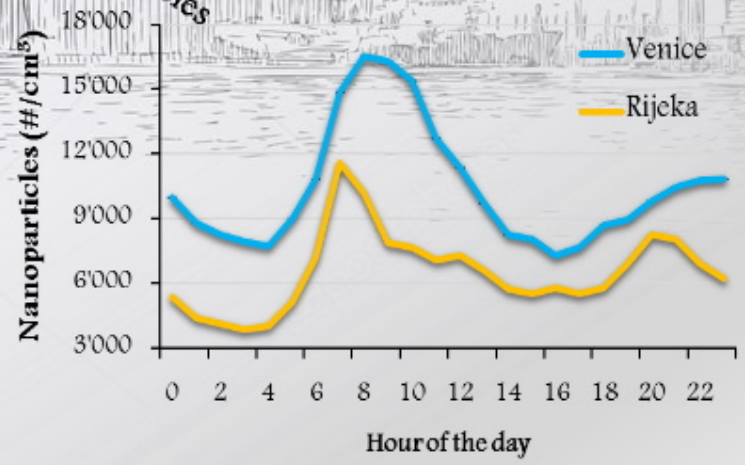
Abstract

Detailed information on in-harbour shipping contribution to size segregated particles in coastal cities are scarce, especially in the busy Mediterranean basin. This poses issues for human exposure and air quality in urban harbour agglomerates, where only criteria pollutants (i.e. PM10 and/or PM2.5) are usually monitored. In this work, particle number and mass size distributions, in a large size range (0.01-31 μm), were obtained in two coastal cities of northern Adriatic Sea: Venice (Italy) and Rijeka (Croatia). Three size ranges were investigated: nanoparticles (diameter $D < 0.25 \mu\text{m}$); fine particles ($0.25 < D < 1 \mu\text{m}$), and coarse particles ($D > 1 \mu\text{m}$). Absolute concentrations were larger in Venice for all size ranges showing, using analysis of daily trends, a large influence of local meteorology and boundary-layer dynamics. Contribution of road transport was larger (in relative terms) in Rijeka compared to Venice. The highest contributions of shipping were in Venice, mainly because of the larger ship traffic. Maximum impact was on nanoparticles 7.4% (Venice) and 1.8% (Rijeka), the minimum was on fine range 1.9% (Venice) and $< 0.2\%$ (Rijeka) and intermediate values were found in the coarse fraction 1.8% (Venice) and 0.5% (Rijeka). Contribution of shipping to mass concentration was not distinguishable from uncertainty in Rijeka ($< 0.2\%$ for PM1, PM2.5, and PM10) and was about 2% in Venice. Relative contributions as function of particles size show remarkable similitudes: a maximum for nanoparticles, a quick decrease and a successive secondary maximum (2-3 times lower than the first) in the fine range. For larger diameters, the relative contributions reach a minimum at 1-1.5 μm and there is a successive increase in the coarse range.

Keywords	particle size distributions; nanoparticles; shipping impacts; ship traffic; harbour pollution
Corresponding Author	EVA MERICO
Corresponding Author's Institution	Institute of Atmospheric Sciences and Climate, National Research Council of Italy (ISAC-CNR)
Order of Authors	EVA MERICO, Marianna Conte, Fabio Massimo Grasso, Daniela Cesari, ANDREA GAMBARO, Elisa Morabito, Elena Gregoris, Salvatore Orlando, Ana Alebic-Juretic, Velimir Zubak, Boris Mifka, Daniele Contini

Highlights

- High temporal resolution aerosol data were collected in two Adriatic port cities
- Shipping contribution to particle concentration of different sizes was investigated
- Contributions to nanoparticles were significantly larger compared to other sizes
- Relative contributions to nanoparticles were 7.4% in Venice and 1.8% in Rijeka
- Contributions to PM₁, PM_{2.5} and PM₁₀ were about 2% in Venice and <0.2% in Rijeka



1 Comparison of the impact of ships to size-segregated particle concentrations in 2 two harbour cities of northern Adriatic Sea

3 Merico E.^{1*}, Conte M.¹, Grasso F.M.¹, Cesari D.¹, Gambaro A.², Morabito E.², Gregoris E.^{2,3},
4 Orlando S.², Alebić-Juretić A.⁴, Zubak V.⁵, Mifka B.⁶, Contini D.¹

5 ¹Institute of Atmospheric Sciences and Climate, National Research Council of Italy (ISAC-CNR), Str. Prv.
6 Lecce-Monteroni km 1.2, 73100 Lecce, Italy

7 ²Department of Environmental Sciences, Informatics and Statistics, Ca' Foscari University of Venice, Via
8 Torino 155, 30172, Venice, Italy

9 ³Institute of Polar Sciences, National Research Council of Italy (ISP-CNR), Via Torino 155, 30172, Venice,
10 Italy

11 ⁴Faculty of Medicine, University of Rijeka, Braće Branchetta 20, Rijeka, Croatia

12 ⁵Teaching Institute of Public Health, Krešimirova 52a, Rijeka, Croatia

13 ⁶Department of Physics, University of Rijeka, Braće Branchetta 20, Rijeka, Croatia

14
15 *Corresponding author: e.merico@isac.cnr.it
16
17
18

19 Abstract

20 Detailed information on in-harbour shipping contribution to size segregated particles in coastal cities
21 are scarce, especially in the busy Mediterranean basin. This poses issues for human exposure and air
22 quality in urban harbour agglomerates, where only criteria pollutants (i.e. PM₁₀ and/or PM_{2.5}) are
23 usually monitored. In this work, particle number and mass size distributions in a large size range
24 (0.01-31 µm) were obtained in two coastal cities of northern Adriatic Sea: Venice (Italy) and Rijeka
25 (Croatia). Three size ranges were investigated: nanoparticles (diameter D<0.25 µm); fine particles
26 (0.25<D<1 µm), and coarse particles (D>1 µm). Absolute concentrations were larger in Venice for
27 all size ranges showing, using analysis of daily trends, a large influence of local meteorology and
28 boundary-layer dynamics. Contribution of road transport was larger (in relative terms) in Rijeka
29 compared to Venice. The highest contributions of shipping were in Venice, mainly because of the
30 larger ship traffic. Maximum impact was on nanoparticles 7.4% (Venice) and 1.8% (Rijeka), the
31 minimum was on fine range 1.9% (Venice) and <0.2% (Rijeka) and intermediate values were found
32 in the coarse fraction 1.8% (Venice) and 0.5% (Rijeka). Contribution of shipping to mass
33 concentration was not distinguishable from uncertainty in Rijeka (<0.2% for PM₁, PM_{2.5}, and PM₁₀)
34 and was about 2% in Venice. Relative contributions as function of particles size show remarkable
35 similitudes: a maximum for nanoparticles, a quick decrease and a successive secondary maximum (2-
36 3 times lower than the first) in the fine range. For larger diameters, the relative contributions reach a
37 minimum at 1-1.5 µm and there is a successive increase in the coarse range.

38 Size distributions showed a not negligible contribution of harbour emissions to nanoparticle and fine
39 particle number concentrations, compared to PM_{2.5} or PM₁₀, indicating them as a better metric to
40 monitor shipping impacts compared to mass concentrations (PM_{2.5} or PM₁₀).

41

42 **Keywords:** particle size distributions; nanoparticles; shipping impacts; ship traffic; harbour pollution

43

44

45

Highlights

46

- 47 • High temporal resolution aerosol data were collected in two Adriatic port cities
- 48 • Shipping contribution to particle concentration of different sizes was investigated
- 49 • Contributions to nanoparticles were significantly larger compared to other sizes
- 50 • Relative contributions to nanoparticles were 7.4% in Venice and 1.8% in Rijeka
- 51 • Contributions to PM₁, PM_{2.5} and PM₁₀ were about 2% in Venice and <0.2% in Rijeka

52

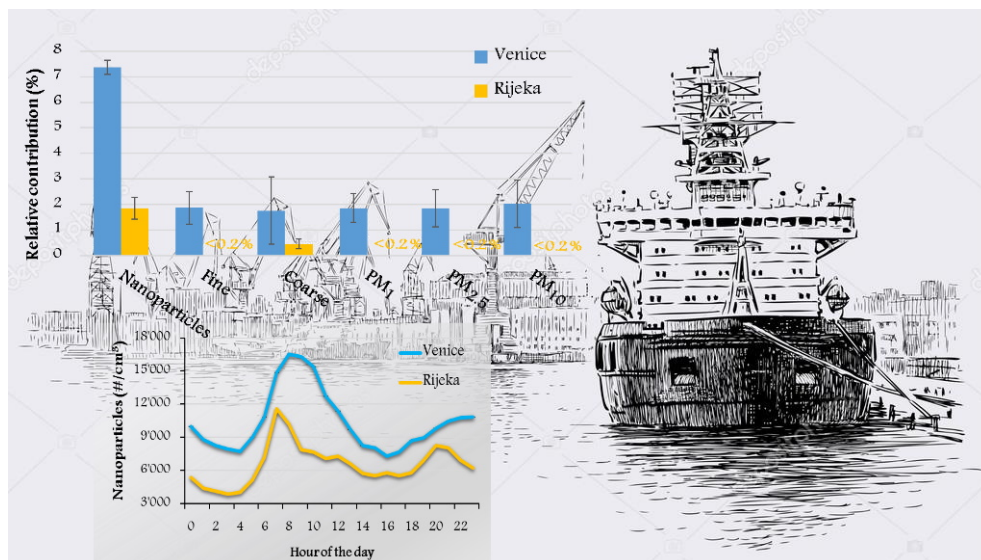
53

54

Graphical abstract

55

56



57

58

59

60

61 1. Introduction

62 International maritime trade **is** expected to expand at an average annual growth rate of 3.5%
63 over the 2019–2024 period (driven by containers, dry bulk and gas cargoes), faster than other
64 transportation modes whose emissions are decreasing because of stricter regulations (UNCTAD,
65 2019). **At the same time, the scientific community and policy makers, especially in harbour areas,**
66 **should address environmental pressures and potential health effects of shipping.**

67 In Europe shipping emissions represent 16%, 11%, and 5% of total anthropogenic NO_x, SO_x,
68 and PM₁₀ emissions, respectively, with a certain variability associated to the emission databases used
69 (Russo et al., 2018). Although local (in harbour) emissions represent a small share compared to those
70 at global scale (Sorte et al., 2019), shipping emissions (mainly PM, NO_x and SO_x) can have important
71 effects on air quality and on exposure of coastal communities (Ramacher et al., 2019; Viana et al.,
72 2020). Hotelling phase, when auxiliary engines are used, is **usually** the largest contributor to local
73 emissions of PM and NO_x, considering that this phase lasts generally more than manoeuvring phase
74 (Jahangiri et al., 2018; Merico et al., 2016; Nunes et al., 2017). Local SO₂ emissions from shipping
75 are generally larger than **those due to other transport sectors, because of the different standards of the**
76 **sulphur content in fuels** (Merico et al., 2017).

77 However, since 01/01/2020 **it** has been enforced the new IMO regulation that sets the
78 maximum sulphur content of fuels used in ships to 0.5% (IMO, 2019), that will lead to a reduction
79 not only of SO_x emission but also of PM (Contini et al., 2015; Liu et al., 2018; Merico et al., 2017;
80 Tao et al., 2013). Environmental and health benefits are expected (Lack et al., 2011; Rouil et al.,
81 2019; Sofiev et al., 2018; Viana et al., 2020; Winebrake et al., 2009). Globally, projections indicate
82 an expected reduction of ship-related premature mortality and morbidity by 34% and 54%,
83 respectively, compared to 2020 scenario without mitigation actions (Sofiev et al., 2018).

84 Different approaches have been developed worldwide to assess shipping contributions to
85 atmospheric pollutants. Source-oriented modelling consisting in transport and dispersion simulations
86 on the basis of shipping emissions have been used at both large (global, continental and/or regional)
87 (Chen et al., 2017; Feng et al., 2019; Jeong et al., 2017; Lang et al., 2017; Monteiro et al., 2018;
88 Murena et al., 2018; Tao et al., 2017) and local scale (Merico et al., 2019). Receptor-oriented
89 approaches have been also widely used, based on high temporal resolution measurements correlated
90 with wind conditions and ship traffic (Contini et al., 2011; Ledoux et al., 2018) or on chemical
91 composition of PM looking for oil combustion tracer (Cesari et al., 2014; Gregoris et al., 2016; Saraga
92 et al., 2019; Scerri et al., 2018; Viana et al., 2009). Average contribution of shipping to PM_{2.5} ranges
93 between 0.2% and 14% in Europe, and similar percentages have also been observed for PM₁₀ (Sorte

94 et al., 2020; Saraga et al., 2019; Sarigiannis et al., 2017; Viana et al., 2014; 2020). In Europe, a clear
95 gradient was observed, with larger contributions in the Mediterranean Sea sites compared to Northern
96 Europe sites (Viana et al., 2014). This is likely due to several factors including intense shipping traffic
97 and unfavourable dispersion conditions.

98 Most of the available studies investigate the impact of in-port shipping to criteria pollutants
99 (i.e. PM_{2.5} or PM₁₀ for particles). In contrast, other studies regarding non-criteria pollutants like
100 particle number concentration (PNC) or regarding impacts to particles of different sizes (including
101 nanoparticles) are relatively scarce (Contini et al., 2015; Donateo et al., 2014; Gobbi et al., 2020;
102 Ledoux et al., 2018; Merico et al., 2017). This is a gap on current knowledge because size and
103 chemical characterisation of ship-emitted particles should be considered for their health and
104 environmental implications (Gwinn and Vallyathan, 2006; Viana et al., 2020). In particular, UFP
105 (ultrafine particles) can act as carriers for transition metals (i.e., vanadium) in the human body with
106 possible adverse influence on respiratory diseases. The implementation of the new IMO regulation
107 for use of low-sulphur content fuel, since 2020, is expected to reduce mortality and morbidity related
108 to PM_{2.5} shipping emissions (Sofiev et al., 2018). Available results (Merico et al., 2016) show that
109 relative contribution on ultrafine particles (diameter <0.3 µm) could be up to 3-4 times larger than
110 those to mass concentration (either PM_{2.5} or PM₁₀). Few studies investigate size-segregated
111 contribution of shipping to particles considering number size distributions (PNSD) or mass size
112 distributions (PMSD). High temporal resolution measurements of ship plumes at the stack or inside
113 harbour area, show a reduction of mass, but not number, of emitted particles in cleaner fuels (from
114 HFO to distillate fuels), with the size distribution moving towards smaller particles (Anderson et al.,
115 2015; Zetterdahl et al., 2017). Typically, ship emissions are characterised by a bimodal size
116 distribution in number (PNSD) and in mass (PMSD). PNSD shows two modes at around 0.04-0.06
117 µm and 0.1-0.2 µm (Kivekäs et al., 2014; Pirjola et al., 2014), but also other modes in nucleation
118 range (at about 0.01 µm) were also observed (Diesch et al., 2013). In terms of PMSD, the bimodal
119 shape of distribution has a first mode in accumulation range (0.4-0.5 µm) and a second one in coarse
120 range (> 1 µm), thus influencing differently the different PM fractions (Merico et al., 2016; 2017;
121 Moldanová et al., 2013).

122 This work aims to contribute to fill the gap in knowledge on the impact of shipping traffic and
123 related-harbour activities on particulate matter of different sizes (ranging from nanoparticles to PM₁₀),
124 both in number and in mass concentrations. The sampling campaigns were performed in two Adriatic
125 port-cities (Rijeka in Croatia and Venice in Italy) by using the same instrumental set-up, integrating
126 high temporal resolution data of size distributions of particles with meteorological measurements and

127 ship **traffic information. Shipping** contributions to particle concentrations as function of particle size
128 were compared at the two sites and with the only results previously available in the Mediterranean
129 basin (for the harbour town of Brindisi in Italy).

130

131 **2. Methodological approach**

132 **2.1 Sampling sites**

133 Sampling campaigns were carried out in two port cities of the Northern Adriatic Sea (Fig. S1): Venice
134 (Italy), and Rijeka (Croatia). The sites are diametrically opposed, separated by the Istrian peninsula,
135 in the northernmost region of the Mediterranean Sea, bounded by the Italian territory westward and
136 the Balkans eastward. Here, intense surface (wind stress, heat and water fluxes) and lateral (river
137 runoffs and open southern boundary transports) fluxes occur. The dominant winds are the Bora, a
138 north-easterly cold, dry and gusty wind, mostly prevailing in winter, and the sirocco, a warm and
139 humid wind blowing from the Southeast along the axis of the Adriatic basin. The Bora winds are
140 strongly sheared due to the orography along the Croatian coasts while events of sirocco, together with
141 other processes like low atmospheric pressure and high astronomical tides, cause flooding in the
142 shallow lagoons (including the Venice Lagoon).

143 The two measurement sites are important logistic hubs for commercial (Rijeka) and tourist
144 (Venice) sea traffic, being both core seaports included in the Mediterranean corridor of TEN-T
145 network (https://ec.europa.eu/transport/themes/infrastructure/ten-t_en). The inclusion in the Baltic-
146 Adriatic corridor is completed for Venice but underway for Rijeka. Both port authorities are strongly
147 enhancing their efforts to upgrade and modernise their infrastructures and capabilities for intermodal
148 connectivity of the ports within wider areas.

149 Rijeka is the third largest city in Croatia (128,624 inhabitants) and its seaport, located on the
150 shore of the Kvarner Gulf at the bottom of the Rijeka bay, is the largest Croatian port. In numbers, a
151 cargo throughput of 17.8 million tonnes in 2018 (liquid+dry+bulk+general cargo) and 260,375
152 containers (in TEUs) were recorded for year 2018; also, a cruisers' flow of 151,983 passengers
153 (15.2% of the total passengers) and Ro/Ro and ferries of 128,882 were accounted
154 (<https://www.portauthority.hr/>).

155 The city of Venice (Veneto Region) is **located in** the Venetian Lagoon, an extremely fragile
156 ecosystem because of its environmental and cultural heritage capital. The area includes a highly
157 populated urban territory (260,520 inhabitants, including Mestre and islands), the largest European
158 coastal industrial settlement of Porto Marghera as well as agricultural and artisan activities. The port
159 is organised in two operative areas with their own separated access, with commercial terminals and

160 passenger piers at Porto Marghera zone and Marittima basin, respectively. The Venice Terminal
161 Passenger (VTP) includes the Marittima **station**, located near the 4-km causeway that links the historic
162 city with the mainland that hosts the largest cruise ships, and the San Basilio pier, just around the
163 corner in the Giudecca Canal, which is devoted to local ferries and catamarans
164 (<https://www.port.venice.it/>). Venice is designated as one of the best Mediterranean homeports with
165 about 1.6 million of cruise passengers in 2018 (<https://www.port.venice.it/>). The tourist harbour at
166 Marittima can hosts several cruise ships at 5 km of quayside of 10 multifunctional passenger terminals
167 and recently it was re-newed for berthing mega yachts too.

168

169 **2.2 Measurement campaigns and instruments used**

170 Two different measurement campaigns were performed in Venice and Rijeka, with the same
171 instrumental set-up (described thereafter) and close to each harbour area (Fig 1a-c). The
172 measurement periods were 06/09/2018-27/11/2018 and 28/03/2019-13/05/2019 for Venice and
173 Rijeka, respectively.

174 The site chosen in Rijeka was on the roof of the Public Health building (45°19'56'' N,
175 14°25'33'' E, 34 m a.s.l.) in front of the **harbour** entrance (approximately 500 m from the main sailing
176 routes and at about 200 m from the closest quay, in a straight line, handling bulk cargo, approximately
177 1 km from the passenger area, and at 2.5-3.0 km from the new container area). It was a background
178 urban site, separated from the port commercial area with an intense cranes activity, by a busy seaside
179 street (named Kresimirova).

180 In Venice, the site was located on the Sacca Fisola island (45°25'42'' N, 12°18'46'' E, 3 m
181 a.s.l.), in front of the Stazione Marittima tourist harbour and beside a fixed environmental monitoring
182 station of the Protection and Prevention Agency of Veneto region (ARPAV). It faces the Giudecca
183 channel that includes the main ship routes, being at about 500 m from the location of ships at berths.

184 The equipment was **set** inside an outdoor two-modular air-conditioned cabinet, as shown in
185 Fig. S1. High-temporal resolution data were taken to collect real-time measurements of main
186 meteorological parameters and concentration of particles of different sizes ranging from 0.01 to 31
187 μm using instruments remotely-controlled via PC. Specifically the setup included:

- 188 • an ultrasonic anemometer (Gill R3 at 100 Hz) coupled with a thermo-hygrometer (Rotronic
189 MP100A, Campbell Scientific) placed on the roof of the cabinet (about 3m above the ground),
190 measuring wind velocity, wind direction, temperature, and relative humidity at 1-min
191 resolution;

- 192 • A CPC (Grimm 5.403) able to measure the total number of sub-micrometric particles, with 1-
193 min resolution. Aerosol was sampled through a 70 cm-long sampling inlet and a portion of
194 the main flow was injected into the CPC through a 50 cm-long conductive silicon tube and a
195 diffusion dryer (silica gel cartridges) to reduce water vapour concentration before the CPC
196 (Merico et al., 2016). The total counting efficiency was evaluated as the product of the
197 penetration factor and the **counting efficiency** of the CPC obtained from Heim et al. (2004).
198 The cut-off diameter (50% efficiency) was 9 nm, thereby the system was measuring particles
199 in the size range 0.009-1 μm (the latter is the upper limit of the CPC).
- 200 • An OPC (Grimm 11-A) able to measure particle number size distributions in the size range
201 0.25-31 μm in 31 size channels, operating at controlled flow of 1.2 L/min. It used the same
202 inlet as the CPC and it operated with 1-min time resolution. The internal software was also
203 able to reconstruct mass size distributions as well as PM_{10} , $\text{PM}_{2.5}$, and PM_{10} mass
204 concentration.
- 205 • A video camera operating at two frames per minute, used to synchronise data of ship
206 movements provided by the port authorities with concentrations and meteorological
207 measurements.

208

209 The OPC and the CPC measured at the same height above the ground (approximately 3 m) and
210 underwent periodic zero tests, on average once per week, during the campaigns.

211

212 **2.3 Statistical approach for evaluation of the impact of shipping**

213 Data of particle concentration, ship traffic (manoeuvring/hotelling) and wind direction were
214 statistically processed on 30-min averages. The methodological approach used in this study for
215 estimating primary ship contribution was originally introduced by Contini et al. (2011) for the Venice
216 harbour, successively applied to the Brindisi harbour (Donateo et al., 2014; Merico et al., 2016) and
217 to other sites (Gregoris et al., 2016; Ledoux et al., 2018; Wang et al., 2019). The contribution was
218 estimated **using the** differences between measured concentrations in cases influenced and not
219 influenced by emissions of ships, selecting wind direction favourable to measure ship plumes
220 (measurement site downwind of the emissions).

221 In Venice, the site was downwind in the range $315^\circ - 360^\circ$ during hotelling and between 315°
222 - 45° during manoeuvring of ships (Fig. S1b). Similarly, the wind direction intervals defined for
223 Rijeka were $122.5^\circ - 180^\circ$ (for hotelling) and $122.5^\circ - 247.5^\circ$ (for manoeuvring) (Fig. S1c).

224 The relative contribution of in-port ship activities to average atmospheric concentration was
225 estimated for each size range by the Eq. (1):

226

$$227 \quad \varepsilon_C = \frac{(C_{DP} - C_{DSP})F_P}{C_D} = \frac{\Delta_C F_P}{C_D} \quad . \quad (\text{Eq. 1})$$

228

229 Where $(C_{DP} - C_{DSP}) = \Delta_C$ is the difference between average concentration in periods potentially
230 influenced (C_{DP}) and not influenced (C_{DSP}) by ship when the site is downwind; C_D is the average
231 concentration in the downwind sector; F_P is the fraction of cases (i.e. 30-min averages) influenced by
232 ship.

233 Uncertainties have been evaluated looking at the variability of ε_C calculated in elaborations
234 done with and without wind calm (velocities <0.2 m/s) and with small changes by $\pm 10^\circ$ in wind
235 direction intervals definition. It should be said that this method could have other uncertainties due to
236 some specific factors (Ausmeel et al., 2019; Wang et al., 2019): choices of wind directions; distance
237 from the docks; choice of cases influenced and not by ship from traffic database; temporal resolution
238 of measurements; non stationary meteorological conditions; collinearity with other surrounding
239 sources present upwind of the measurement site in the same sector where ships are located.

240

241 3. Results and discussion

242 3.1 Meteorological conditions and ship traffic data

243 Local meteorology of each site should be carefully investigated due to its influence on measurements.
244 As briefly described in Section 2.1, the climate in Northern Adriatic (and therefore at the site) is
245 extremely influenced by the orography of Gorski Kotar and the Dinarides. In summer, there are north-
246 western winds (etesians) in the open sea, and, at the same time, local daily periodic circulation is
247 developed between the larger islands and the coast, generating a sea breeze regime. In winter (and at
248 night), local conditions are dominant. Dominant wind (especially in the coastal area of Istria) is Bora,
249 reaching up to several tens of kilometers per hour, thus creating problems to road and maritime traffic
250 (Poje, 1992). In Venice, the daily cycle of the wind direction is recognized within the general air
251 circulation pattern of the Venice lagoon (Contini et al., 2011; Prodi et al., 2009). It can be described
252 as having two prevalent wind directions: a nocturnal prevailing wind direction from N-NE and a
253 diurnal one from S-SE.

254 The wind roses for the two measurements campaigns are shown in Fig. S2. During the
255 sampling campaign in Rijeka, there was a dominant wind direction from ESE and a second wind
256 direction sector from S to NE with slightly stronger winds from E-ENE. This indicated that the site

257 was influenced mainly by Sirocco. Instead, for the Venice site, the wind rose showed a dominant
258 direction from NE (mainly during night, coming from Alps mountains) and, a second direction from
259 SE (from the Adriatic Sea) during daytime. This is the typical circulation of Venice lagoon, also
260 observed in other measurement campaigns in the same area, especially in spring and summer seasons
261 (Contini et al., 2015).

262 Furthermore, starting from high-temporal resolution hourly averaged data, daily patterns of
263 temperature and relative humidity were obtained for both sites (Fig. S2). Both variables were lower,
264 on average, in Rijeka compared to Venice, as a consequence of the different measurement periods as
265 well as of local circulation conditions. Average temperatures of about 15° and 17°C were measured
266 in Rijeka and Venice, respectively. Relative humidity was between 50% and 66% in Rijeka, instead,
267 a higher value, about 80%, was observed in Venice with 70% reached only in diurnal hours between
268 10:00 and 17:00).

269 In the period between 24 and 26 April 2019 (during Rijeka campaign), an intense event of
270 Saharan dust occurred on a large scale interesting also the measurement site. Back-trajectories of air
271 masses calculated by Hysplit model (Fig. S3a) and the simulations of the Dust REgional Atmospheric
272 Model (BSC-DREAM8b) (Fig. S3b) confirmed the phenomenon. The event lead to a significant
273 increase in the number of coarse ($D > 1 \mu\text{m}$) particles, while a limited contribution on the
274 concentration of sub-micrometric particles was observed (Fig. S3c). For this reason, corresponding
275 data were excluded by the analysis of the ship contribution in order to avoid their influence on average
276 concentrations.

277 As described in Section 2.3, measurements when the site was downwind (and during
278 manoeuvring and/or hoteling phases) need to be selected. Ship traffic (arrivals/departures) in both
279 harbours (using data provided by Rijeka and Venice Port Authority synchronised with concentration
280 measurements) were used to evaluate the daily pattern. These are compared with the daily patterns of
281 the percentage of time in which the site is downwind of the harbour areas at the two sites (Fig. 1). In
282 total, 92 and 240 ships in Rijeka and Venice, respectively, were recorded during the entire sampling
283 campaigns, with vessel traffic in Rijeka harbour of about 8.6% in gross tonnage (about 820,000 tons)
284 of that in Venice (around 9,500,000 tons). Both at arrival and departure, gross tonnage and number
285 of vessels showed a gradual decrease (25-30%) of the total number going from September to October
286 in Venice, however, a rapid reduction was present in November (about 70%) both in gross tonnage
287 and number, due to the end of the cruise period in the area. Contrarily, in Rijeka, many smaller ships

288 (i.e. ferries) were recorded since mid-April compared to the first days of the same month and in May
289 (46% less than April but with larger ships such as cargoes and bulk carrier).

290 A clear daily trend is present for Venice with arrival of ships mainly concentrated in the first
291 hours of the morning (6:00-8:00) and ship departure in the afternoon (16:00-18:00). In this case,
292 considering that the percentage of downwind cases was significantly larger (>70%) during the night
293 and the morning (time interval 2:00-10:00) and decreased by half (down to <30%) during afternoon,
294 it is reasonable that ship arrival will give the most relevant contribution at the site studied (Fig. 1a).
295 This is also found in previous works (Contini et al., 2015) for evaluation of the impact of ship traffic
296 to air quality in a nearby site (in Sacca San Biagio, 45° 25' 38.50'' N – 12° 18' 33.86'' E at 1 km
297 south of the Stazione Marittima of Venice). For Rijeka, a clear daily trend of ship traffic was not
298 observed (Fig. 1b), even if there is a greater traffic volume in the central hours of the day compared
299 to the night. The site in Rijeka was potentially influenced (>60% of cases) by ships during the whole
300 day (time interval 8:00-18:00) and this means by the majority of ship traffic (both in arrival and
301 departure).

302

303 **3.2 Particle mass and number concentrations**

304 Combining the CPC data with measurements of the OPC allowed to obtain the average size
305 distribution in number (Fig. 2a) and in mass (Fig. 2b) for both sites. Particles associated with high
306 mass are those in the coarse fraction, usually associated to dust from breaks or road surfaces, bio-
307 aerosol, sea spray; while high number concentration could be due to combustion emissions of
308 ultrafine particles of soot, sulphates, primary organic aerosol (POA), and secondary organic aerosol
309 (SOA). Looking at size distributions, three size ranges, likely influenced by different sources and
310 processes, were identified and used for further post-processing: nanoparticles (or ultrafine particles
311 $D < 0.25 \mu\text{m}$); fine particles ($0.25 < D < 1 \mu\text{m}$); coarse particles ($D > 1 \mu\text{m}$).

312 Average number concentrations in the different size ranges are reported in Table S1, showing
313 lower concentrations in Rijeka compared to Venice: ranging from 47.5% of the concentration
314 observed in Venice for fine particles up to about 77.6% for coarse particles. Nanoparticles in Rijeka
315 are 64.4% of those observed in Venice. The number size distributions have very similar shape for
316 Venice and Rijeka, with the highest value of about 10,000 $\#/\text{cm}^3$ in the nanoparticles range,
317 decreasing up to a few particles per cm^3 at diameters of 0.6 μm . Size distributions in mass are similar
318 at the two sites showing a bimodal shape, even if concentrations in Venice are larger than those in
319 Rijeka. The first mode is centred at diameters around 0.3-0.4 μm (at both sites) and it is likely

320 influenced by combustion sources including shipping; instead the second mode is broad (size range
321 2-5 μm) in Venice and slightly narrow in Rijeka (2-3 μm), being influenced also by mechanical
322 processes and natural sources like soil dust and sea spray (Fridell et al., 2008; Merico et al., 2016;
323 Moldanová et al., 2013). It should be noted that local road vehicles could influence particle
324 concentrations mainly in Rijeka, taking into account the location site near a traffic-loaded road and
325 logistic activities in the harbour area (i.e. loading/unloading of ships). Instead, this influence is likely
326 more limited in Venice, being the site located on an island directly facing the passenger terminal.

327 Average mass concentrations of PM_{10} , $\text{PM}_{2.5}$ and PM_1 measured at both sites were
328 significantly different, more than 50% higher values were observed in Venice compared to Rijeka
329 (Tab. 1). The same trend was found for accumulation particles that differ significantly at both sites
330 with number particle concentration in Rijeka accounting for 48% compared to that in Venice. Instead,
331 nanoparticles and coarse particles had a relative lower difference with 22% and 35% between the two
332 locations (with larger values in Venice).

333 Daily patterns of number concentration of nanoparticles and larger particles (i.e. sum of fine
334 and coarse fractions) at both sites were compared in Fig. 3. The daily trends at the two sites are quite
335 different. Looking at nanoparticles, in Rijeka two evident peaks are visible associated to typical rush
336 hours in the morning (up to about 12,000 $\#/\text{cm}^3$) and in the evening (up to about 8000 $\#/\text{cm}^3$), followed
337 by a low decrease in the night. In Venice a much broader morning peak is observed between 7:00 and
338 10:00 up to about 17,000 $\#/\text{cm}^3$, instead, in the evening it is not visible a peak, rather there is a slow
339 increase likely related to the development of the shallow stable nocturnal boundary-layer. This slow
340 increase, related to the boundary-layer dynamics, starting in late afternoon and continuing up to late
341 night, was also observed for $\text{PM}_{2.5}$ concentrations in other sites of the Venice lagoon (Donateo et al.,
342 2012). The decrease after 10:00 is related to the change of wind direction, that typically happens at
343 that time (Fig. 1a), in which sea breeze starts to bring air masses from the SSE-SE direction cleaner
344 compared to the nocturnal and early morning air masses coming from the NNW sector that travel
345 above the urban area and the harbour of Venice.

346 Looking at particles with $D > 0.25 \mu\text{m}$, larger concentrations are observed in Venice with a
347 complete different daily trend compared to Rijeka. Concentrations in Rijeka exhibited a small peak
348 in the morning between 7:00 and 9:00, likely influenced by emissions of specific urban sources (i.e.
349 road traffic) being correlated with the analogous peak in nanoparticles. Furthermore, there is a second
350 peak in the evening rush hours at around 20:00-21:00. In Venice, the trend is completely different,
351 having a shape typical of urban background and rural sites (Dinoi et al., 2017) with a modulation due
352 to the atmospheric stability and boundary-layer height. Specifically, it shows a decrease starting early

353 in the morning (at about 5:00) reaching a minimum (about -28% lower than nocturnal values) in the
354 early afternoon and a slow increase starting late in the afternoon and during the night.

355 Looking at mass concentrations, daily trends of PM_{10-1} concentration have the same behaviour
356 (almost superimposable) of particles with $D > 0.25 \mu m$ at both sites. A completely different pattern
357 was observed for PM_{10-1} (coarse mode). At both sites, larger concentrations are observed during the
358 day. In Venice, two maxima are individuated: in the morning and the late afternoon, with
359 concentration peaks of about $9 \mu g/m^3$ and $7.5 \mu g/m^3$, respectively. Instead, in Rijeka a broad increase
360 in diurnal hours was evident, with a maximum value in the morning peaks of approximately $4 \mu g/m^3$.

361 This analysis suggests that, at both sites, local meteorology has played a role in determining
362 concentrations and its influence was more evident in the Venice site. Even if the observed
363 concentrations are lower compared to Venice, it appears that road traffic has a large relative impact
364 on the Rijeka site.

365

366 **3.3 Primary contribution of shipping emissions to particles of different sizes**

367 For Venice, in order to limit the influence of the boundary layer dynamics maintaining almost all ship
368 traffic (as reported in Fig. 1a), the Eq. (1) was applied on a subset of data selecting only hours between
369 5:00 and 23:00. Instead, from Rijeka dataset, the period 24-26 April 2019 was removed,
370 corresponding to the intense African dust advection event, as previously described (§3.1).

371 The absolute contributions of ship traffic at the two sites are reported in Table S2 for particles
372 in the different size ranges, the relative contributions are reported in Fig. 4. In Venice, the contribution
373 to nanoparticles was about $1000 \#/cm^3$, about eight times that observed for Rijeka (around $130 \#/cm^3$).
374 Contributions to fine and coarse particles are obviously much smaller than that to nanoparticles and
375 they are larger in Venice compared to Rijeka.

376 The relative contribution of shipping to nanoparticles was $7.4 \pm 0.3\%$ in Venice and $1.8 \pm 0.4\%$
377 in Rijeka and smaller contributions were found for number particle concentrations in the other two
378 size ranges, between 1.7% and 2.0% in Venice and between 0.2% and 0.5% in Rijeka. The
379 contribution of shipping to measured concentrations is larger in Venice for all size ranges, as
380 consequence of the larger traffic of ships (section 3.1) and of the smaller distance of the site from the
381 emissions (i.e. the harbour area). It is known that contribution of shipping emissions to air quality
382 peaks near the harbour area and it quickly decreases with distance from the harbour (Merico et al.,
383 2019). The general trend, in relative terms, is the same at both sites: larger contribution to
384 nanoparticles, lower contribution to fine particles and a slight increase in the coarse range. Usual
385 metrics for mass concentrations have comparable or only slightly variable contributions in both

386 absolute and relative terms. This is because primary emissions from ships are due to exhaust plumes
387 and are characterised by ultrafine particles as observed in several studies (Diesch et al. 2013; Kivekäs
388 et al., 2014; Pirjola et al., 2014; Merico et al., 2016; Ledoux et al., 2018; Ausmeel et al., 2019). In
389 some studies, relatively fresh ship exhaust particle size distributions revealed either unimodal or
390 bimodal structures, however, a typical bimodal size distribution was observed with the two modes
391 centred at around 40-60 nm and 100-200 nm (Kivekäs et al., 2014; Pirjola et al., 2014). In the harbour
392 area of Calais it has been observed that when wind was blowing from the harbour the number of
393 particles was ten times higher compared to background level, with the highest differences in the 30-
394 67 nm and the 109-167 nm size ranges (Ledoux et al., 2018). A contribution of shipping in the
395 nucleation range (at about 10 nm) was found at the banks of the Elbe in Northern Germany in Diesch
396 et al. (2013).

397 In terms of mass concentrations, the absolute contributions to PM₁, PM_{2.5}, and PM₁₀ were
398 comparable and in the range 0.4-0.5 μg/m³ in Venice. In Rijeka, these were not clearly distinguishable
399 above the uncertainties. The relative contributions were about 2% in Venice, similar for the different
400 size ranges, and <0.2% in Rijeka. Looking at Venice, it is interesting to observe that absolute
401 contribution to PM₁ was essentially comparable with that to PM_{2.5} and was about 80% of the
402 contribution to PM₁₀. This is very similar to the results obtained in another port city (Brindisi) of the
403 Adriatic Sea in which the contribution to PM₁ was about 80% of that to PM₁₀ and the contribution of
404 PM_{2.5} was about 84% of that to PM₁₀ (Merico et al., 2016). This happens because the vast majority
405 of the exhaust emissions from ships are in the ultrafine range (Kasper et al., 2007). The results
406 obtained here support the idea that particle number concentrations, in the nanoparticle or ultrafine
407 size ranges could be a better metric, compared to PM₁, PM_{2.5}, or PM₁₀, to investigate the impact of
408 shipping to local air quality as suggested also in other studies (Merico et al., 2016; Muntean et al.,
409 2019; Gobbi et al., 2020).

410

411 3.4 Comparison with other studies

412 Relatively few works are focused on impact of shipping to nanoparticles and fine particles number
413 concentrations, however, several studies were performed on the impacts to mass concentrations,
414 mainly PM₁, PM_{2.5}, and PM₁₀ (Viana et al., 2014; 2020; Merico et al., 2017; Sorte et al., 2020). In
415 Europe, the contributions to PM_{2.5} or PM₁₀ ranges between 0.2% and 14% and there is a clear gradient
416 with larger contribution in Mediterranean area compare to northern Europe. The contribution to total
417 particle number concentrations (PNC) are expected to be 3-4 times larger than that to PM_{2.5} (Merico

418 [et al., 2016](#)). The values found here are essentially comparable with the previous observations in other
419 Mediterranean cities.

420 Previous estimates for Venice (Contini et al., 2015, Gregoris et al., 2016) and Rijeka (Merico
421 et al., 2017), done with similar methodological approaches but in different sites, could be compared
422 with the results found here. In Rijeka (Merico et al., 2017), the contributions of shipping in the period
423 2013-2014 estimated for a site located at the harbour entrance were 0.5% ($\pm 0.2\%$) for $PM_{2.5}$ and 0.3%
424 ($\pm 0.1\%$) for PM_{10} , with a decreasing trend moving from 2013 to 2014. These values are comparable
425 with those observed in this work, however, information on contributions of shipping to nanoparticles
426 or fine particles in number were not previously available. In Venice, the relative contributions of
427 shipping to PM_{10} were found in the range between 1.9% and 2.5% at three different sites (Gregoris
428 et al., 2016). Contributions of ships to total particle number concentrations (PNC) in the size range
429 0.005-3 μm and to $PM_{2.5}$ in Venice were estimated for the summer 2012 at the Sacca San Biagio site,
430 located near that studied here (less than 200 m), in Contini et al. (2015). The comparison of absolute
431 and relative contributions found in 2012 and 2018 is reported in Fig. 4. Looking at absolute
432 contributions, that to $PM_{2.5}$ was very similar in the two years (approximately 0.4 $\mu g/m^3$), however,
433 there was an increase of the contribution to PNC from 800 $\#/cm^3$ in 2012 to over 1000 $\#/cm^3$ in 2018.
434 The relative contributions depends on the average concentrations observed that were larger in 2018
435 for both PNC and $PM_{2.5}$. This leads, when relative contributions are considered (Fig. 5), to
436 comparable impacts to PNC, taking into account uncertainty, and lower relative contribution to $PM_{2.5}$.

437 Several measurements of the contributions of ships to atmospheric particle concentrations are
438 available for the port city of Brindisi, located in South Italy facing the Adriatic Sea (Cesari et al.,
439 2014; Donateo et al., 2014; Merico et al., 2016). These refer to two sites: one located inside the
440 harbour area near the docks of ferries, and the other one located in the urban area at about 1.4 km
441 from the harbour. Contributions to $PM_{2.5}$ ranged from 2.8% (urban area) to 7.8% (inside harbour
442 area). Contributions to PNC, measured only inside the harbour area ranged between 23% and 26% in
443 different years. In the year 2014, a characterisation of the size distributions of shipping impact was
444 done for the Brindisi harbour area (Merico et al., 2016) using the same instruments and the same
445 methodological approach used in this work. The size distributions of relative shipping contributions
446 in Venice and Rijeka are compared with the results obtained in Brindisi in Fig. 6. Results obtained in
447 the three harbour cities show different details but also remarkable similarities in the general shape.
448 Relative contributions show a maximum for nanoparticles a quick decrease and a successive
449 secondary maximum in the fine range. The secondary maximum is not distinguishable in Rijeka
450 above the uncertainty, but it is present in Brindisi (in the range 0.3-0.45 μm) and in Venice (in the

451 range 0.4-0.7 μm) being 2-3 times lower than the absolute maximum. For larger diameters, the
452 relative contributions reach a minimum in the size range between 1 μm and 1.5 μm and successively,
453 in the coarse size range, there is growth of the relative contribution for all sites.

454

455 **4. Conclusions**

456 This study represents an attempt to estimate the local impact of harbour activities on particulate matter
457 concentration of different sizes, and, therefore, with different environmental and health issues. High
458 temporal resolution measurements of size distribution and local meteorology were collected in two
459 Adriatic coastal cities with the same instruments and processed with the same methodology. This
460 allowed a direct comparison of results between sites and, in addition, with previous studies conducted
461 with the same approach in other harbours of the Adriatic region.

462 Analysis of size distributions in number and mass allowed focus the results in three size
463 ranges: nanoparticles (diameter $D < 0.25 \mu\text{m}$); fine particles ($0.25 \mu\text{m} < D < 1 \mu\text{m}$), and coarse particles
464 ($D > 1 \mu\text{m}$). Results show that absolute concentrations in number were larger in Venice (from 28% to
465 100% larger according to the size range) and the same happens for mass concentrations (PM_{10} , $\text{PM}_{2.5}$,
466 and PM_{10}) that were approximately twice compared to Rijeka.

467 Daily trends of particles in the different size ranges showed significant differences when the
468 two sites were compared. In Venice there was a larger influence of local meteorology and boundary-
469 layer dynamics, and a clear influence of anthropogenic sources was observed mainly in the
470 nanoparticle range. In Rijeka, the contribution of road transport was instead evident and larger (in
471 relative terms) compared to Venice.

472 The contributions of shipping to measured particle concentrations were significantly larger in
473 Venice compared to Rijeka as consequence mainly of the larger ship traffic and partly because of the
474 largest distance of the measurement site from the docks. However, a similar trend for the different
475 particle sizes was observed. The maximum impact was found on nanoparticles $7.4 \pm 0.3\%$ in Venice
476 and $1.8 \pm 0.4\%$ in Rijeka, the minimum was observed in the fine range $1.9 \pm 0.6\%$ (Venice) and $< 0.2\%$
477 (Rijeka) and intermediate values were found for the coarse fraction $1.8 \pm 1.3\%$ (Venice) and $0.5 \pm 0.2\%$
478 (Rijeka). Contribution of shipping to mass concentration was not distinguishable from uncertainty in
479 Rijeka ($< 0.2\%$ for PM_{10} , $\text{PM}_{2.5}$, and PM_{10}) and was approximately 2% ($\pm 0.7\%$) in Venice. These
480 values correspond to absolute contributions ranging from $0.4 \mu\text{g}/\text{m}^3$ for PM_{10} and $\text{PM}_{2.5}$ to $0.5 \mu\text{g}/\text{m}^3$
481 for PM_{10} . It is interesting to observe that the absolute contribution to $\text{PM}_{2.5}$ is about 80% of that to
482 PM_{10} . This suggests that primary shipping emissions are mainly composed by ultrafine particles and

483 number concentrations, especially in the nanoparticles size range, that could be a better metric to
484 investigate this source compared to air quality standards (PM_{2.5} or PM₁₀).

485 Detailed analysis of the relative contribution as function of particle size was extended
486 comparing results with those previously obtained in another harbour of the Adriatic Sea (Brindisi).
487 Results obtained in the three harbour cities show different details but also remarkable similarities in
488 the general shape. Relative contributions show a maximum for nanoparticles a quick decrease and a
489 successive secondary maximum in the fine range. The secondary maximum is not distinguishable in
490 Rijeka above the uncertainty but it is present in Brindisi (in the range 0.3-0.45 µm) and in Venice (in
491 the range 0.4-0.7 µm) being 2-3 times lower than the absolute maximum. For larger diameters, the
492 relative contributions reach a minimum in the size range between 1 µm and 1.5 µm and successively,
493 in the coarse size range, there is growth of the relative contribution for all sites. **In conclusion, this**
494 **study points out the significant relevance of harbour activities for human exposure and local air**
495 **quality mainly for nanoparticles that are more harmful for human health. This is of particulate interest**
496 **for harbours of the Mediterranean basin where such studies are scarce and and increase of maritime**
497 **traffic is expected in near future. Future efforts in sustainable harbour management should be focused**
498 **on monitoring and reducing nanoparticles, not currently included in legislation, in order to achieve**
499 **both climate and health benefits.**

500

501 **Acknowledgments**

502 This work was performed within the framework of the project “ECOLOGICAL supporting for traffic
503 Management in cOastal areas By using an IntelLIgenT sYstem” (ECOMOBILITY) co-funded by the
504 European Regional Development Fund and Interreg Italy-Croatia CBC Program, and national
505 resources. Financial contribution is gratefully acknowledged. Authors acknowledge the support of
506 the Port of Rijeka Authority and Port Authority of Venice. Also, the support of the Regional Agency
507 of Environmental Protection and Prevention (ARPA) of the Veneto Region for hosting instruments.

508

509 **References**

- 510 Anderson, M., Salo, K., Fridell, E., 2015. Particle- and Gaseous Emissions from an LNG Powered
511 Ship. *Environ. Sci. Technol.* 49, 12568–12575. <https://doi.org/10.1021/acs.est.5b02678>
- 512 Ausmeel, S., Eriksson, A., Ahlberg, E., Kristensson, A., 2019. Methods for identifying aged ship
513 plumes and estimating contribution to aerosol exposure downwind of shipping lanes. *Atmos.*
514 *Meas. Tech. Discuss.* 1–20. <https://doi.org/10.5194/amt-2018-445>
- 515 Cesari, D., Genga, A., Ielpo, P., Siciliano, M., Mascolo, G., Grasso, F.M., Contini, D., 2014. Source
516 apportionment of PM_{2.5} in the harbour-industrial area of Brindisi (Italy): Identification and
517 estimation of the contribution of in-port ship emissions. *Sci. Total Environ.* 497–498, 392–
518 400. <https://doi.org/10.1016/j.scitotenv.2014.08.007>

- 519 Chen, D., Wang, X., Nelson, P., Li, Y., Zhao, N., Zhao, Y., Lang, J., Zhou, Y., Guo, X., 2017. Ship
520 emission inventory and its impact on the PM_{2.5} air pollution in Qingdao Port, North China.
521 *Atmos. Environ.* 166, 351–361. <https://doi.org/10.1016/j.atmosenv.2017.07.021>
- 522 Contini, D., Gambaro, A., Donato, A., Cescon, P., Cesari, D., Merico, E., Belosi, F., Citron, M.,
523 2015. Inter-annual trend of the primary contribution of ship emissions to PM_{2.5} concentrations
524 in Venice (Italy): Efficiency of emissions mitigation strategies. *Atmos. Environ.* 102, 183–
525 190. <https://doi.org/10.1016/j.atmosenv.2014.11.065>
- 526 Contini, D., Gambaro, A., Belosi, F., De Pieri, S., Cairns, W.R.L., Donato, A., Zanotto, E., Citron,
527 M., 2011. The direct influence of ship traffic on atmospheric PM_{2.5}, PM₁₀ and PAH in Venice.
528 *J. Environ. Manage.* 92, 2119–2129. <https://doi.org/10.1016/j.jenvman.2011.01.016>
- 529 Diesch, J.-M., Drewnick, F., Klimach, T., Borrmann, S., 2013. Investigation of gaseous and
530 particulate emissions from various marine vessel types measured on the banks of the Elbe in
531 Northern Germany. *Atmos. Chem. Phys.* 13, 3603–3618. <https://doi.org/10.5194/acp-13-3603-2013>
- 533 Dinoi, A., Donato, A., Belosi, F., Conte, M., Contini, D., 2017. Comparison of atmospheric particle
534 concentration measurements using different optical detectors: Potentiality and limits for air
535 quality applications. *Meas. J. Int. Meas. Confed.* 106, 274–282.
536 <https://doi.org/10.1016/j.measurement.2016.02.019>
- 537 Donato, A., Contini, D., Belosi, F., Gambaro, A., Santachiara, G., Cesari, D., Prodi, F., 2012.
538 Characterisation of PM_{2.5} concentrations and turbulent fluxes on a island of the Venice lagoon
539 using high temporal resolution measurements. *Meteorol. Zeitschrift* 21, 385–398.
540 <https://doi.org/10.1127/0941-2948/2012/0354>
- 541 Donato, A., Gregoris, E., Gambaro, A., Merico, E., Giua, R., Nocioni, A., Contini, D., 2014.
542 Contribution of harbour activities and ship traffic to PM_{2.5}, particle number concentrations
543 and PAHs in a port city of the Mediterranean Sea (Italy). *Environ. Sci. Pollut. Res.* 21, 9415–
544 9429. <https://doi.org/10.1007/s11356-014-2849-0>
- 545 Feng, J., Zhang, Y., Li, S., Mao, J., Patton, A., Zhou, Y., Ma, W., Liu, C., Kan, H., Huang, C., An,
546 J., Li, L., Shen, Y., Fu, Q., Wang, X., Liu, J., Wang, S., Ding, D., Cheng, J., Ge, W., Zhu, H.,
547 Walker, K., 2019. The influence of spatiality on shipping emissions, air quality and potential
548 human exposure in the Yangtze River Delta/Shanghai, China. *Atmos. Chem. Phys.* 19, 6167–
549 6183. <https://doi.org/10.5194/acp-19-6167-2019>
- 550 Fridell, E., Steen, E., Peterson, K., 2008. Primary particles in ship emissions. *Atmos. Environ.* 42,
551 1160–1168. <https://doi.org/10.1016/j.atmosenv.2007.10.042>
- 552 Gobbi, G.P., Di Liberto, L., Barnaba, F., 2020. Impact of port emissions on EU-regulated and non-
553 regulated air quality indicators: The case of Civitavecchia (Italy). *Sci. Total Environ.* 719,
554 134984. <https://doi.org/10.1016/j.scitotenv.2019.134984>
- 555 Gregoris, E., Barbaro, E., Morabito, E., Toscano, G., Donato, A., Cesari, D., Contini, D., Gambaro,
556 A., 2016. Impact of maritime traffic on polycyclic aromatic hydrocarbons, metals and
557 particulate matter in Venice air. *Environ. Sci. Pollut. Res.* 23 (7), 6951–6959.
558 <https://doi.org/10.1007/s11356-015-5811-x>
- 559 **Gwinn, M.R., Vallyathan, V., 2006. Nanoparticles: Health Effects – Pros and Cons, *Environ Health***
560 ***Perspect* 114 (12), 1818–1825, <https://doi.org/10.1289/ehp.8871>.**
- 561 Heim, M., Kasper, G., Reischl, G.P., Gerhart, C., 2004. Performances of a new commercial electrical
562 mobility spectrometer. *Aerosol. Sci. Tech.* 38 (S2), 3–14.
- 563 IMO, 2019. The 2020 global sulphur limit. International Maritime Organization (IMO).
564 [http://www.imo.org/en/OurWork/Environment/PollutionPrevention/AirPollution/
565 Pages/Sulphur-oxides-\(SOx\)-%E2%80%93Regulation-14.aspx](http://www.imo.org/en/OurWork/Environment/PollutionPrevention/AirPollution/Pages/Sulphur-oxides-(SOx)-%E2%80%93Regulation-14.aspx).
- 566 Jahangiri, S., Nikolova, N., Tenekedjiev, K., 2018. An improved emission inventory method for
567 estimating engine exhaust emissions from ships. *Sustain. Environ. Res.* 28, 374–381.
568 <https://doi.org/10.1016/j.serj.2018.08.005>.

- 569 Jeong, J.H., Shon, Z.H., Kang, M., Song, S.K., Kim, Y.K., Park, J., Kim, H., 2017. Comparison of
570 source apportionment of PM_{2.5} using receptor models in the main hub port city of East Asia:
571 Busan. *Atmos. Environ.* 148, 115–127. <https://doi.org/10.1016/j.atmosenv.2016.10.055>
- 572 Kasper, A., Aufdenblatten, S., Forss, A., Mohr, M., Burtcher, H., 2007. Particulate emissions from a
573 low-speed marine diesel engine. *Aerosol Sci. Technol.* 41 (1), 24-32.
- 574 Kivekäs, N., Massling, A., Grythe, H., Lange, R., Rusnak, V., Carreno, S., Skov, H., Swietlicki, E.,
575 Nguyen, Q.T., Glasius, M., Kristensson, A., 2014. Contribution of ship traffic to aerosol
576 particle concentrations downwind of a major shipping lane. *Atmos. Chem. Phys.* 14, 8255–
577 8267. <https://doi.org/10.5194/acp-14-8255-2014>
- 578 Lack, D. A., Cappa, C. D., Langridge, J., Bahreini, R., Buffaloe, G., Brock, C., Cerully, K., Coffman,
579 D., Hayden, K., Holloway, J., Lerner, B., Massoli, P., Li, S. M., McLaren, R., Middlebrook,
580 A. M., Moore, R., Nenes, A., Nuaaman, I., Onasch, T. B., Peischl, J., Perring, A., Quinn, P.
581 K., Ryerson, T., Schwartz, J. P., Spackman, R., Wofsy, S. C., Worsnop, D., Xiang, B., and
582 Williams, E., 2011. Impact of Fuel Quality Regulation and Speed Reductions on Shipping
583 Emissions: Implications for Climate and Air Quality. *Environ. Sci. Technol.* 45, 9052-9060,
584 doi:10.1021/es2013424.
- 585 Lang, J., Zhou, Y., Chen, D., Xing, X., Wei, L., Wang, X., Zhao, N., Zhang, Y., Guo, X., Han, L.,
586 Cheng, S., 2017. Investigating the contribution of shipping emissions to atmospheric PM_{2.5}
587 using a combined source apportionment approach. *Environ. Pollut.* 229, 557–566.
588 <https://doi.org/10.1016/j.envpol.2017.06.087>
- 589 Ledoux, F., Roche, C., Cazier, F., Beaugard, C., Courcot, D., 2018. Influence of ship emissions on
590 NO_x, SO₂, O₃ and PM concentrations in a North-Sea harbor in France. *J. Environ. Sci.* 71,
591 56–66. <https://doi.org/10.1016/j.jes.2018.03.030>
- 592 Liu, H., Jin, X.X., Wu, L.L., Wang, X.M., Fu, M.L., Lv, Z.F., et al., 2018. The impact of marine
593 shipping and its DECA control on air quality in the Pearl River Delta, China. *Sci. Total*
594 *Environ.* 625, 1476–1485.
- 595 Merico, E., Dinoi, A., Contini, D., 2019. Development of an integrated modelling-measurement
596 system for near-real-time estimates of harbour activity impact to atmospheric pollution in
597 coastal cities. *Transp. Res. Part D Transp. Environ.* 73, 108–119.
598 <https://doi.org/10.1016/j.trd.2019.06.009>
- 599 Merico, E., Gambaro, A., Argiriou, A., Alebic-Juretic, A., Barbaro, E., Cesari, D., Chasapidis, L.,
600 Dimopoulos, S., Dinoi, A., Donateo, A., Giannaros, C., Gregoris, E., Karagiannidis, A.,
601 Konstandopoulos, A.G., Ivošević, T., Liora, N., Melas, D., Mifka, B., Orlić, I., Poupkou, A.,
602 Sarovic, K., Tsakis, A., Giua, R., Pastore, T., Nocioni, A., Contini, D., 2017. Atmospheric
603 impact of ship traffic in four Adriatic-Ionian port-cities: Comparison and harmonization of
604 different approaches. *Transp. Res. Part D Transp. Environ.* 50, 431–445.
605 <https://doi.org/10.1016/j.trd.2016.11.016>
- 606 Merico, E., Donateo, A., Gambaro, A., Cesari, D., Gregoris, E., Barbaro, E., Dinoi, A., Giovanelli,
607 G., Masieri, S., Contini, D., 2016. Influence of in-port ships emissions to gaseous atmospheric
608 pollutants and to particulate matter of different sizes in a Mediterranean harbour in Italy.
609 *Atmos. Environ.* 139, 1–10. <https://doi.org/10.1016/j.atmosenv.2016.05.024>
- 610 Moldanová, J., Fridell, E., Winnes, H., Holmin-Fridell, S., Boman, J., Jedynska, A., Tishkova, V.,
611 Demirdjian, B., Joulie, S., Bladt, H., Ivleva, N.P., Niessner, R., 2013. Physical and chemical
612 characterisation of PM emissions from two ships operating in European emission control
613 areas. *Atmos. Meas. Tech.* 6, 3577–3596. <https://doi.org/10.5194/amt-6-3577-2013>
- 614 Monteiro, A., Russo, M., Gama, C., Borrego, C., 2018. How important are maritime emissions for
615 the air quality: At European and national scale. *Environ. Pollut.* 242, 565–575.
616 <https://doi.org/10.1016/J.ENVPOL.2018.07.011>
- 617 Muntean, M. et al., Identifying key priorities in support to the EU Macro-regional Strategies
618 implementation – An ex-ante assessment for the Adriatic-Ionian and Alpine regions focusing

619 on clean growth in transport and bioenergy, European Commission, Ispra, 2019, JRC110395

620 Murena, F., Mocerino, L., Quaranta, F., Toscano, D., 2018. Impact on air quality of cruise ship
621 emissions in Naples, Italy. *Atmos. Environ.* 187, 70–83.
622 <https://doi.org/10.1016/j.atmosenv.2018.05.056>

623 Nakatsubo, R., Oshita, Y., Aikawa, M., Takimoto, M., Kubo, T., et al., 2020. Influence of marine
624 vessel emissions on the atmospheric PM_{2.5} in Japan's around the congested sea areas. *Sci.*
625 *Tot. Environ.* 702, 134744.

626 Nunes, R.A.O., Alvim-Ferraz, M.C.M., Martins, F.G., Sousa, S.I.V., 2017. Assessment of shipping
627 emissions on four ports of Portugal. *Environ. Pollut.* 231, 1370–1379.
628 <https://doi.org/10.1016/j.envpol.2017.08.112>

629 Pirjola, L., Pajunoja, A., Walden, J., Jalkanen, J.-P., Rönkkö, T., Kousa, A., Koskentalo, T., 2014.
630 Mobile measurements of ship emissions in two harbour areas in Finland. *Atmos. Meas. Tech.*
631 7, 149–161. <https://doi.org/10.5194/amt-7-149-2014>

632 Poje, D., 1992: Wind persistence in Croatia. *Int. J. Climatol* 12, 569-586.

633 Prodi, F., Belosi, F., Contini, D., Santachiara, G., Di Matteo, L., Gambaro, A., Donateo, A., Cesari,
634 D., 2009. Aerosol fine fraction in the Venice Lagoon: Particle composition and sources.
635 *Atmos. Res.* 92, 141–150. <https://doi.org/10.1016/j.atmosres.2008.09.020>

636 Ramacher, M.O.P., Karl, M., Bieser, J., Jalkanen, J.-P., Johansson, L., 2019. Urban population
637 exposure to NO_x emissions from local shipping in three Baltic Sea harbour cities e a generic
638 approach. *Atmos. Chem. Phys.* 19, 9153–9179. <https://doi.org/10.5194/acp-19-9153-2019>.

639 Russo, M.A., Leitão, J., Gama, C., Ferreira, J., Monteiro, A., 2018. Shipping emissions over Europe:
640 A state-of-the-art and comparative analysis. *Atmos. Environ.* 177, 187–194.
641 <https://doi.org/10.1016/J.ATMOSENV.2018.01.025>

642 Saraga, D.E., Tolis, E.I., Maggos, T., Vasilakos, C., Bartzis, J.G., 2019. PM_{2.5} source apportionment
643 for the port city of Thessaloniki, Greece. *Sci. Total Environ.* 650, 2337–2354.
644 <https://doi.org/10.1016/j.scitotenv.2018.09.250>

645 Sarigiannis, D.A., Handakas, E.J., Kermenidou, M., Zarkadas, I., Gotti, A., Charisiadis, P., Makris,
646 K., Manousakas, M., Eleftheriadis, K., Karakitsios, S.P., 2017. Monitoring of air pollution
647 levels related to Charilaos Trikoupis Bridge. *Sci. Total Environ.* 609, 1451-1463.

648 Scerri, M.M., Kandler, K., Weinbruch, S., Yubero, E., Galindo, N., Prati, P., Caponi, L., Massabò,
649 D., 2018. Estimation of the contributions of the sources driving PM_{2.5} levels in a Central
650 Mediterranean coastal town. *Chemosphere* 211, 465–481.
651 <https://doi.org/10.1016/j.chemosphere.2018.07.104>

652 Sofiev, M., Winebrake, J.J., Johansson, L., Carr, E.W., Prank, M., Soares, J., Vira, J., Kouznetsov,
653 R., Jalkanen, J.-P., Corbett, J.J., 2018. Cleaner fuels for ships provide public health benefits
654 with climate tradeoffs. *Nat. Commun.* 9, 406. <https://doi.org/10.1038/s41467-017-02774-9>

655 Sorte, S., Arunachalam, S., Naess, B., Seppanen, C., Rodrigues, V., Valencia, A., Borrego, C.,
656 Monteiro, A., 2019. Assessment of source contribution to air quality in an urban area close to
657 a harbor: Case-study in Porto, Portugal. *Sci. Total Environ.* 662, 347–360.
658 <https://doi.org/10.1016/j.scitotenv.2019.01.185>

659 Sorte, S., Rodrigues, V., Borrego, C., Monteiro, A., 2020. Impact of harbour activities on local air
660 quality: A review. *Environ Poll* 257, 11354. <https://doi.org/10.1016/j.envpol.2019.113542>.

661 Tao, J., Zhang, L., Cao, J., Zhong, L., Chen, Dongsheng, Yang, Y., Chen, Duohong, Chen, L., Zhang,
662 Z., Wu, Y., Xia, Y., Ye, S., Zhang, R., 2017. Source apportionment of PM_{2.5} at urban and
663 suburban areas of the Pearl River Delta region, south China - With emphasis on ship
664 emissions. *Sci. Total Environ.* 574, 1559–1570.
665 <https://doi.org/10.1016/j.scitotenv.2016.08.175>

666 Tao, L., Fairley, D., Kleeman, M.J., Harley, R.A., 2013. Effects of switching to lower sulfur marine
667 fuel oil on air quality in the San Francisco Bay area. *Environ. Sci. Technol.* 47, 10171–10178.

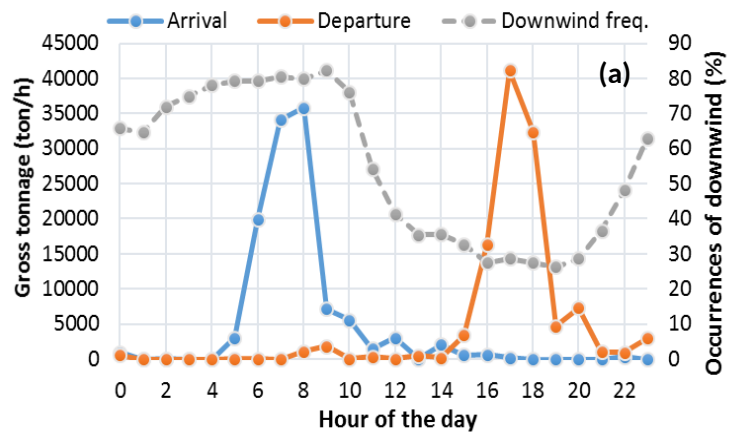
668 UNCTAD, 2019. Review of Maritime Transport 2019, UN, New York,

- 669 <https://doi.org/10.18356/17932789-en>
- 670 Viana, M., Amato, F., Alastuey, A., Querol, X., Moreno, T., Dos Santos, S.G., Herce, M.D.,
671 Fernández-Patier, R., 2009. Chemical tracers of particulate emissions from commercial
672 shipping. *Environ. Sci. Technol.* 43 (19), 7472–7477. <https://doi.org/10.1021/es901558t>.
- 673 Viana, M., Hammings, P., Colette, A., Querol, X., Degraeuwe, B., Vliieger, I. de, van Aardenne, J.,
674 2014. Impact of maritime transport emissions on coastal air quality in Europe. *Atmos.*
675 *Environ.* 90, 96–105. <https://doi.org/10.1016/j.atmosenv.2014.03.046>
- 676 Viana, M., Rizza, V., Tobías, A., Carr, E., Corbett, J., Sofiev, M., Karanasiou, A., Buonanno, G.,
677 Fann, N., 2020. Estimated health impacts from maritime transport in the Mediterranean region
678 and benefits from the use of cleaner fuels. *Environ. Int.* 138, 105670.
679 <https://doi.org/10.1016/j.envint.2020.105670>
- 680 Wang, X., Shen, Y., Lin, Y., Pan, J., Zhang, Y., Louie, K.K.P., Li, M., Fu, Q., 2019. Atmospheric
681 pollution from ships and its impact on local air quality at a port site in Shanghai. *Atmos.*
682 *Chem. Phys.* 19, 6315–6330. <https://doi.org/10.5194/acp-19-6315-2019>
- 683 Winebrake, J.J., Corbett, J.J., Green, E.H., Lauer, A., Eyring, V., 2009. Mitigating the health impacts
684 of pollution from oceangoing shipping: an assessment of low-sulfur fuel mandates. *Environ.*
685 *Sci. Technol.* 43, 4776–4782. <https://doi.org/10.1021/es803224q>
- 686 Winnes, H., Styhre, L., Fridell, E., 2015. Reducing GHG emissions from ships in port areas. *Res.*
687 *Transp. Bus. Manag.* 17, 73-82. <https://doi.org/10.1016/j.rtbm.2015.10.008>
- 688 Zetterdahl, M., Salo, K., Fridell, E., Sjöblom, J., 2017. Impact of aromatic concentration in marine
689 fuels on particle emissions. *J. Mar. Sci. Appl.* 16, 352–361. [https://doi.org/10.1007/s11804-](https://doi.org/10.1007/s11804-017-1417-7)
690 [017-1417-7](https://doi.org/10.1007/s11804-017-1417-7)
- 691

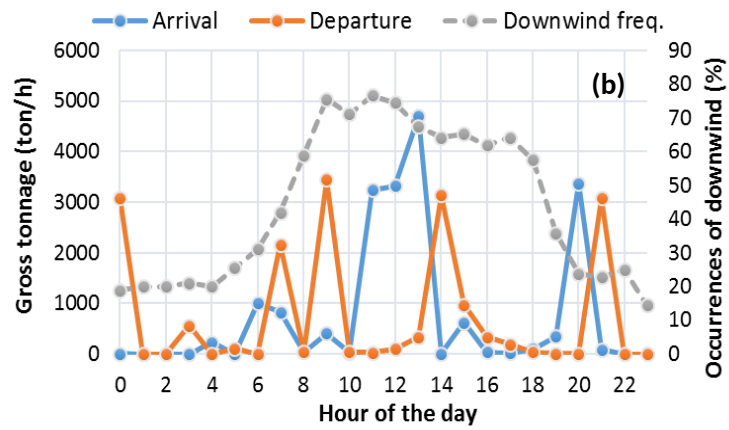
1

Figure 1

2



3



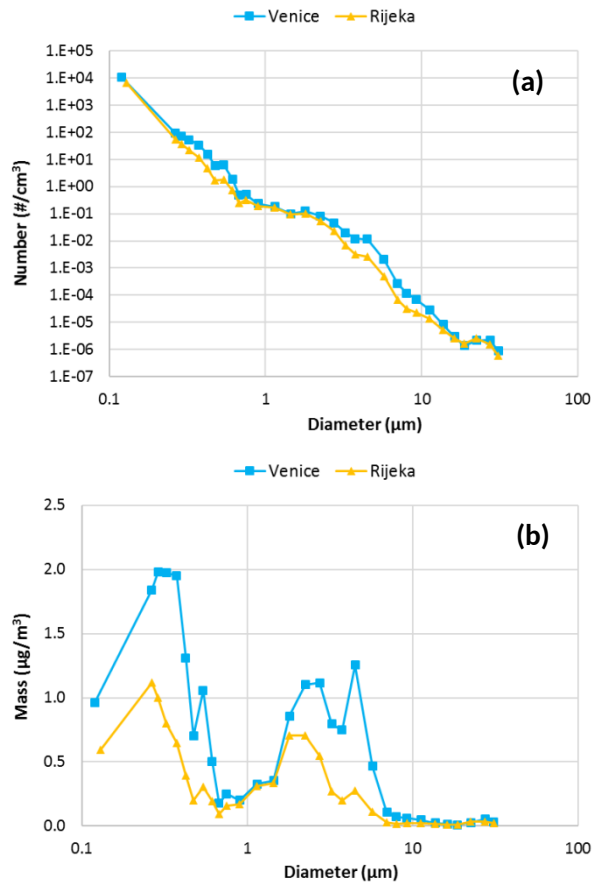
4

5 Figure 1) Daily trend of ship traffic (in terms of gross tonnage per hour) and of the percentage of time
6 when the sites of Venice (a) and Rijeka (b) were downwind of ship emissions.

7

8

Figure 2



9

10

11

12

13

14 Figure 2) Average particle size distribution in number (a) and in mass (b) in the two sites.

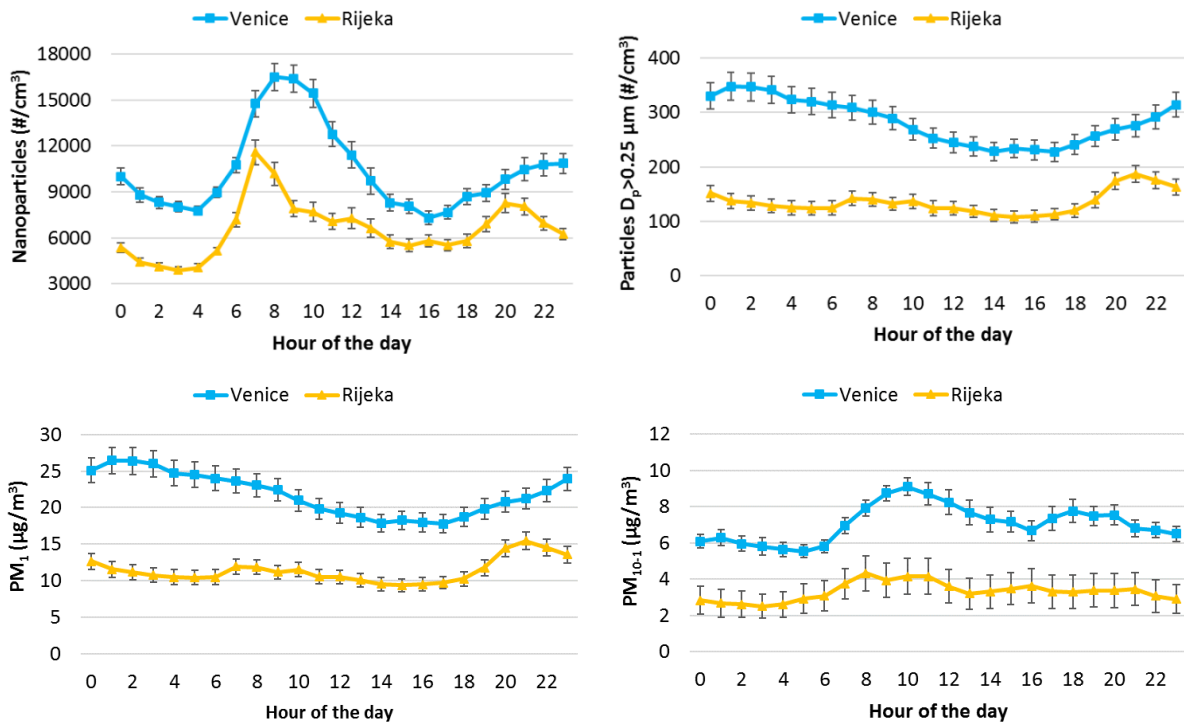
15

16

17

Figure 3

18



19

20

21 Figure 3) Average daily patterns of concentration in number (upper) and mass (lower) in Rijeka and
22 Venice (with error standard indicated by the error bars).

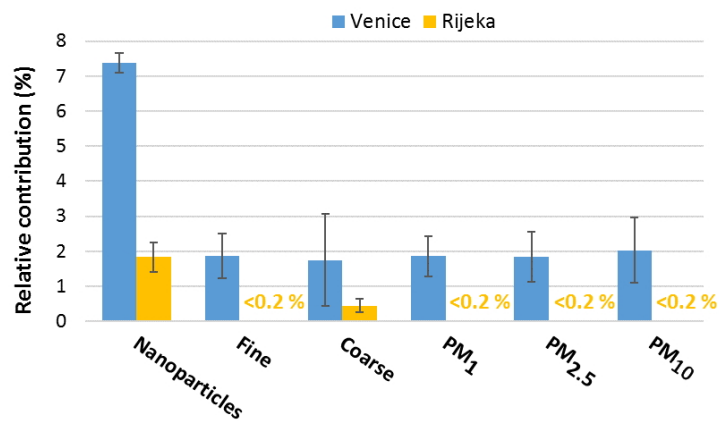
23

24

25

26

Figure 4



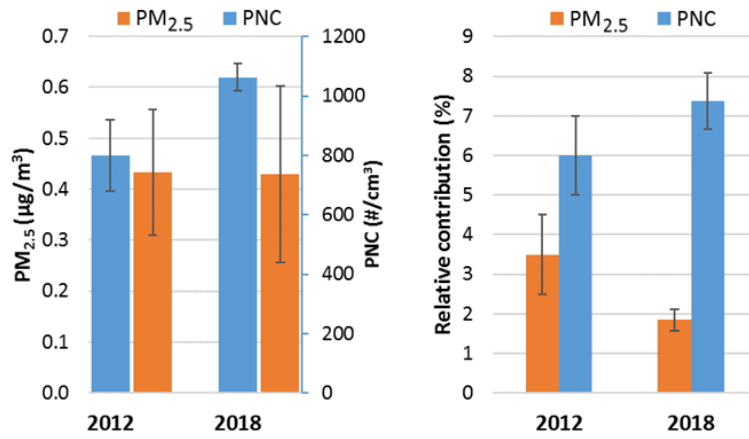
27

28 Figure 4) Relative contribution to particles concentration (in mass and number) in Rijeka and Venice
29 for the different size ranges.

30

31
32
33

Figure 5



34

35 Figure 5) Comparison in terms of absolute (left) and relative (right) contributions of ships to PNC
36 and PM_{2.5} observed in Venice in 2012 and in 2018.

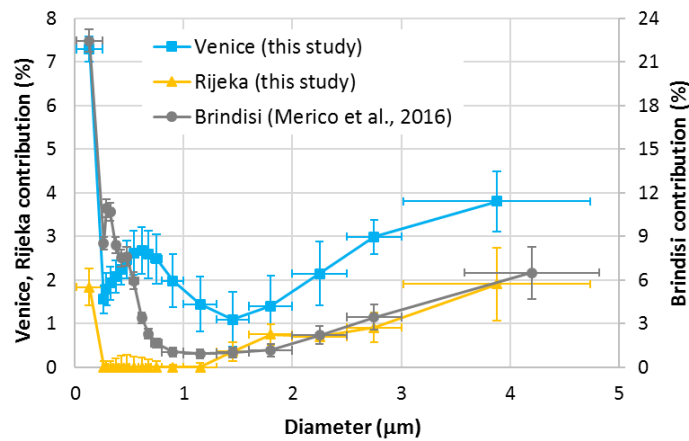
37

38

39

40

Figure 6



41

42 Figure 6) Comparison of relative contributions of shipping to atmospheric particle concentrations as
43 function of size for three harbour towns of the Adriatic Sea. Vertical bars represent the errors and
44 horizontal bars the size of the channel used in the evaluations.

45

Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

CRedit authorship contribution statement

E. Merico: conceptualization, formal analysis, investigation, data curation, writing – original draft, visualisation, writing – review & editing. **M. Conte:** investigation, data curation. **F.M. Grasso:** investigation, data curation. **D. Cesari:** investigation, data curation. **A. Gambaro:** conceptualization, funding acquisition, project administration, resources. **E. Morabito:** investigation, validation, visualization. **E. Gregoris:** investigation, validation, visualization. **S. Orlando:** Funding Acquisition, resources. **A. Alebić-Juretić:** conceptualization, funding acquisition, data curation, formal analysis, investigation, methodology, resources, supervision, validation, visualization, writing – original draft, writing – review & editing. **V. Zubak:** investigation, data curation. **B. Mifka:** investigation, data curation, validation. **D. Contini:** conceptualization, funding acquisition, data curation, formal analysis, investigation, methodology, resources, supervision, validation, visualization, writing – original draft, writing – review & editing

1 **Comparison of the impact of ships to size-segregated particle concentrations in**
2 **two harbour cities of northern Adriatic Sea**

3 Merico E.^{1*}, Conte M.¹, Grasso F.M.¹, Cesari D.¹, Gambaro A.², Morabito E.², Gregoris E.^{2,3},
4 Orlando S.², Alebić-Juretić A.⁴, Zubak V.⁵, Mifka B.⁶, Contini D.¹

5 ¹Institute of Atmospheric Sciences and Climate, National Research Council of Italy (ISAC-CNR), Str. Prv.
6 Lecce-Monteroni km 1.2, 73100 Lecce, Italy

7 ²Department of Environmental Sciences, Informatics and Statistics, Ca' Foscari University of Venice, Via
8 Torino 155, 30172, Venice, Italy

9 ³Institute of Polar Sciences, National Research Council of Italy (ISP-CNR), Via Torino 155, 30172, Venice,
10 Italy

11 ⁴Faculty of Medicine, University of Rijeka, Braće Branchetta 20, Rijeka, Croatia

12 ⁵Teaching Institute of Public Health, Krešimirova 52a, Rijeka, Croatia

13 ⁶Department of Physics, University of Rijeka, Braće Branchetta 20, Rijeka, Croatia

14
15 *Corresponding author: e.merico@isac.cnr.it
16

17
18 **Supporting information**

19 **Total pages – 5**

20 **Figures – 3**

21 **Tables – 2**
22

23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48

Figure S1

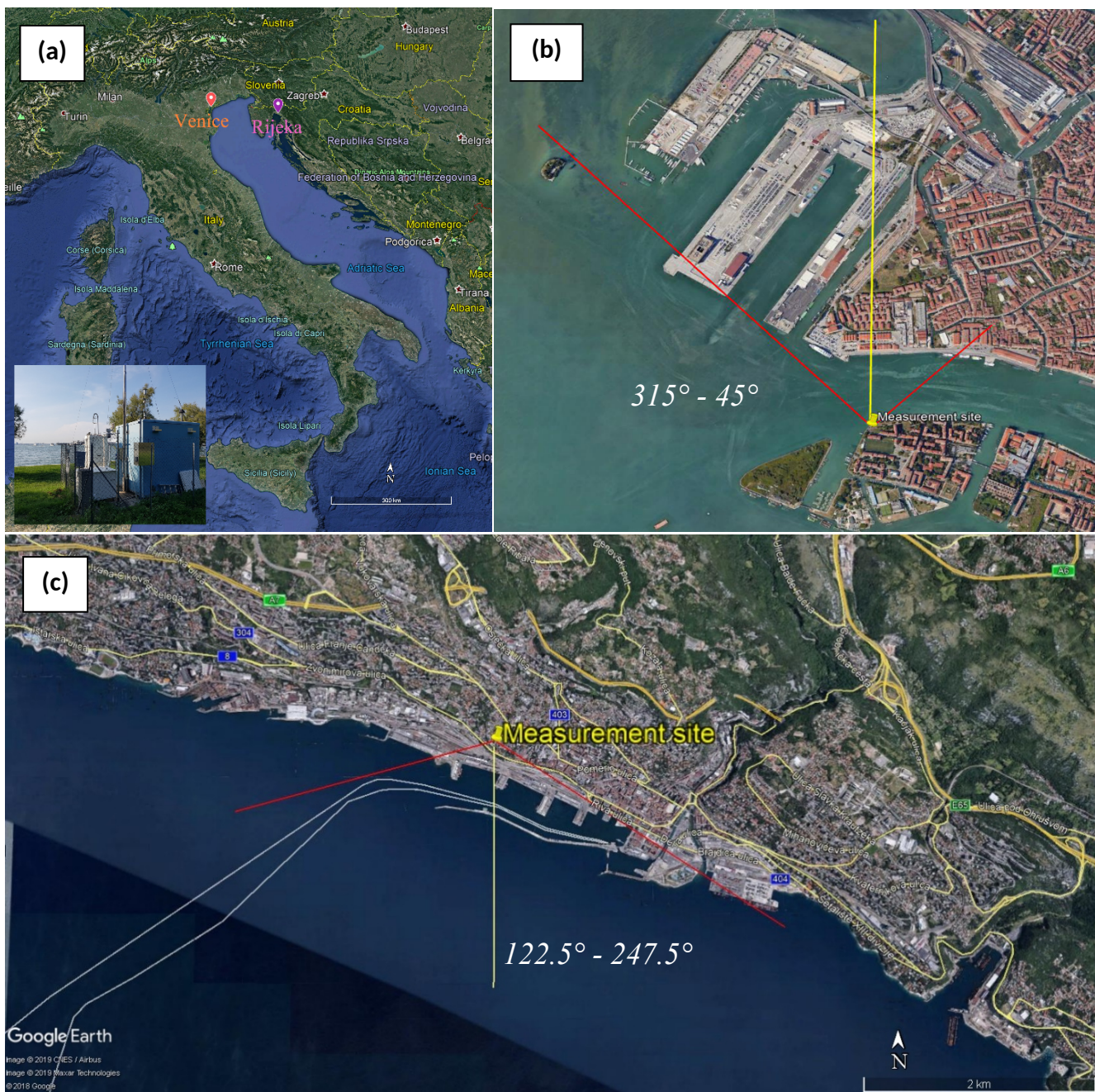
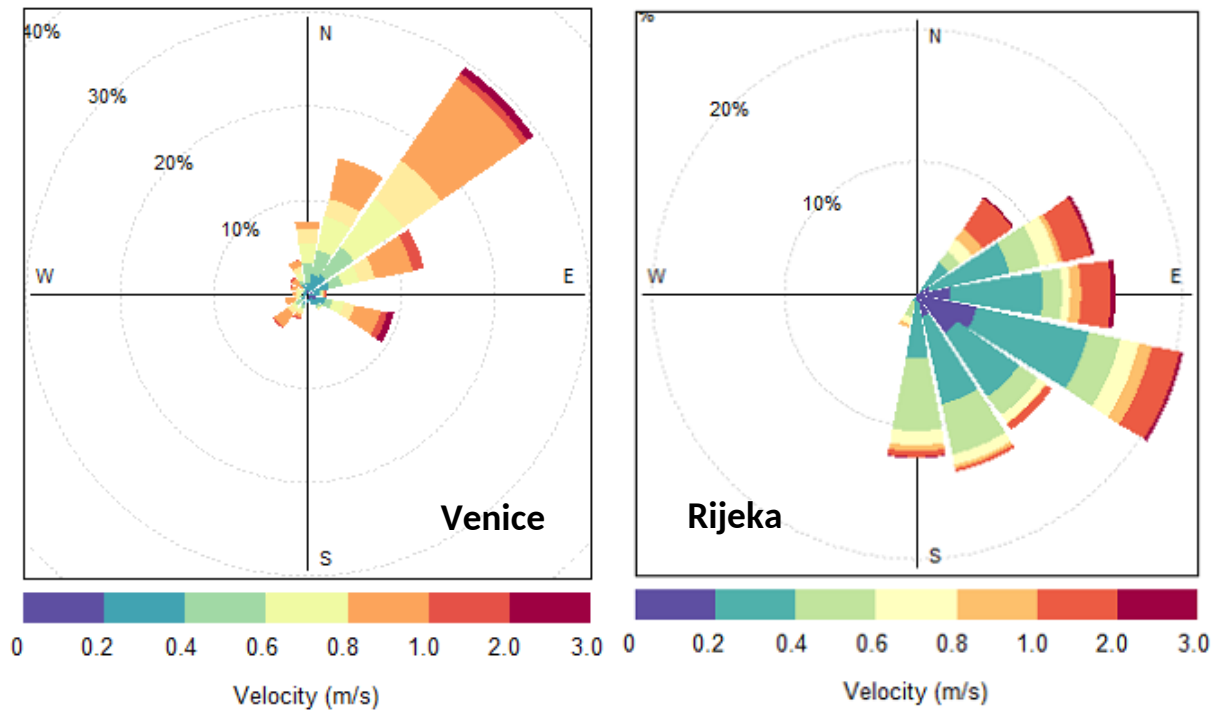


Figure S1) Measurement sites and setup used (a) with indication of the angles used to select data in post-processing in Venice (b) and Rijeka (c).

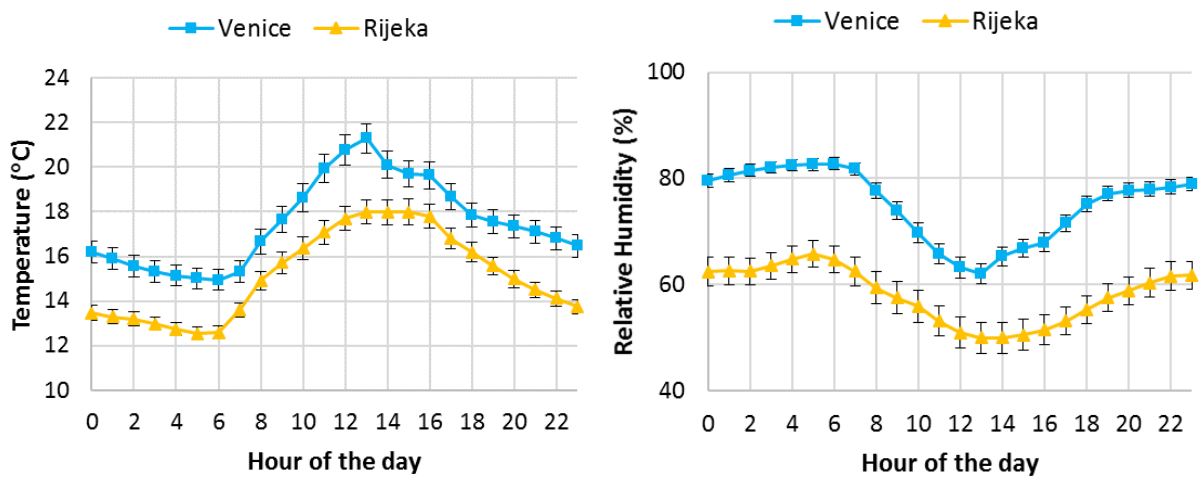
49

Figure S2

50



51



52

53

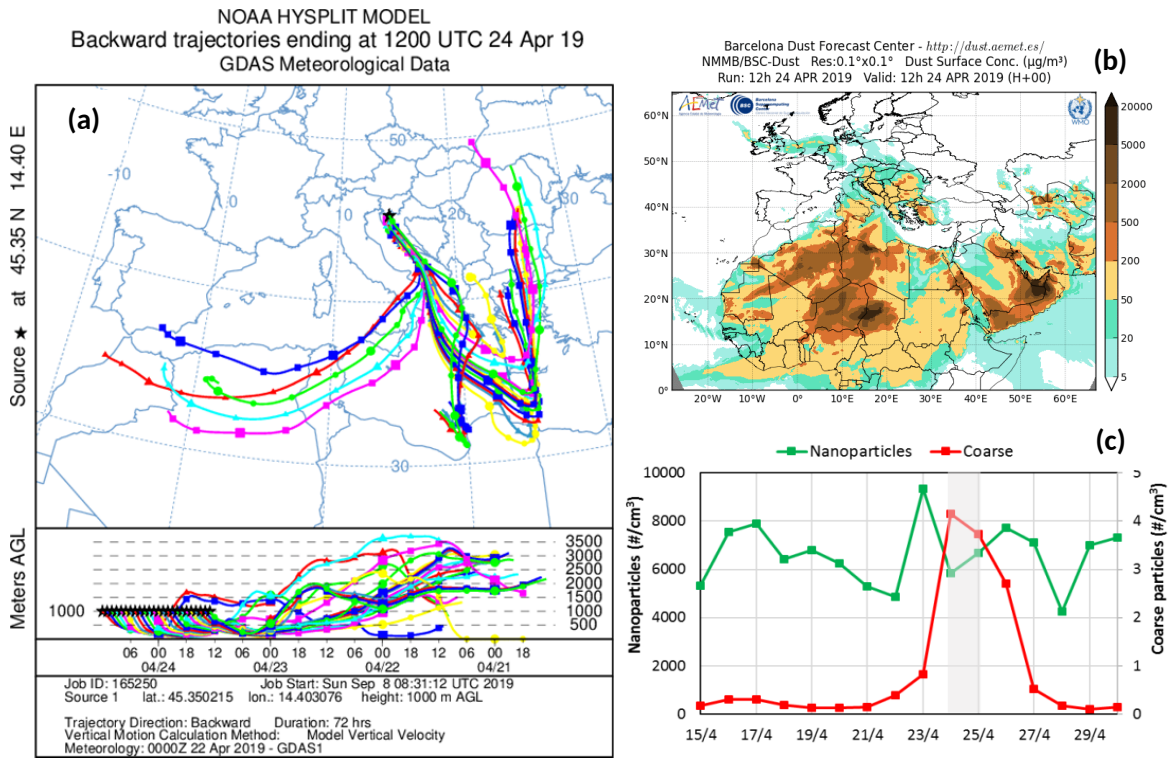
54 Figure S2) Wind roses for the two measurement campaigns (top) and average daily trends of the
55 temperature and relative humidity (bottom) in Rijeka and Venice (with standard errors indicated by
56 the error bars).

57

58

Figure S3

59



60

61 Figure S3) (a) Back-trajectories (<http://arl.noaa.gov/ready/>), (b) NMMB/BSC images
62 (<https://ess.bsc.es/bsc-dust-daily-forecast>) and (c) daily number concentration of coarse particles and
63 nanoparticles for the Saharan dust event recorded in Rijeka starting on 24/04/2019.

64

65 **Table S1**

66

67 Table S1. Average and standard deviation of particle concentration in mass and number at both sites
68 (Rijeka – RJ, Venice – VE) and for the different size fractions. The last line is the average ratio
69 between measurements in Rijeka and in Venice.

	Site	PM ₁ (µg/m ³)	PM _{2.5} (µg/m ³)	PM ₁₀ (µg/m ³)	Nanoparticles (#/cm ³)	Fine (#/cm ³)	Coarse (#/cm ³)
Average	RJ	11.4	13.4	14.8	6552	134.1	0.45
STD	RJ	6.9	8.8	10.4	3738	90.4	0.89
Average	VE	21.8	24.8	28.9	10166	282.0	0.58
STD	VE	13.6	14.1	15.2	6326	188.2	0.42
Ratio (%)	-	52.3	54.0	51.2	64.4	47.5	77.6

70

71

72

73 **Table S2**

74

75 Table S2. Average absolute contributions of shipping to particle concentrations in mass and number
76 at both sites for the different size ranges. In parenthesis the uncertainties.

	PM ₁ (µg/m ³)	PM _{2.5} (µg/m ³)	PM ₁₀ (µg/m ³)	Nanoparticles (#/cm ³)	Fine (#/cm ³)	Coarse (10 ⁻³ #/cm ³)
Rijeka	< 0.03	< 0.03	< 0.03	130 (±30)	< 0.3	1.2 (0.5)
Venice	0.41 (0.18)	0.43 (0.17)	0.53 (0.35)	1063 (±50)	4.8 (1.8)	10 (7)

77

78

79