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Development and validation of new analytical methods using sea urchin embryo bioassay to evaluate dredged marine sediments --Manuscript Draft--

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Order of Authors:	Rosa Bonaventura Francesca Zito Lorenzo Morroni David Pellegrini Francesco Regoli Annalisa PINSINO, PhD.
Abstract:	<p><p style="margin-right: 0cm; margin-left: 0cm; font-size: medium; font-family: &quot;Times New Roman&quot;, serif; caret-color: rgb(0, 0, 0); color: rgb(0, 0, 0); text-align: justify; line-height: 24px;">Management of dredged materials disposal is regulated by several environmental normative requirements, and it is often supported by the integration of chemical data with ecotoxicological characterization. The reliability of a bioassay to assess the potential toxicity of dredged sediments requires the selection of quality criteria that should be based on simple analytical methods and easily understandable hazard for politicians and environmental managers. The sea urchin embryo-toxicity bioassay is considered an essential component for evaluating the quality of sediments in harbour areas but its use, when based exclusively on the observation of normal vs. abnormal embryos, may alter the interpretation of the results, overestimating the risk assessment. To improve the reliability of this assay in&nbsp;&nbsp;a&nbsp;causative relationship between&nbsp;quality of sediments&nbsp;and sea urchin embryonic development, here we developed and validated three&nbsp;Integrative Toxicity Indexes (ITI 2.0, ITI 3.0, ITI 4.0),&nbsp;modifying the already-known ITI&nbsp;(here ITI 1.0). &nbsp;Based on this aim,&nbsp;we used Taranto harbour as a model pilot-study to&nbsp;compare results to those obtained from standard&nbsp;criteria. Among the tested indexes, the ITI 4.0, discriminating strictly developmental delay and morphological defects from fertilized egg to gastrula stage,&nbsp;resulted in the most promising.&nbsp;</p></p></p>

Development and validation of new analytical methods using sea urchin embryo bioassay to evaluate dredged marine sediments

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Abstract

Management of dredged materials disposal is regulated by several environmental normative requirements, and it is often supported by the integration of chemical data with ecotoxicological characterization. The reliability of a bioassay to assess the potential toxicity of dredged sediments requires the selection of quality criteria that should be based on simple analytical methods and easily understandable hazard for politicians and environmental managers. The sea urchin embryo-toxicity bioassay is considered an essential component for evaluating the quality of sediments in harbour areas but its use, when based exclusively on the observation of normal vs. abnormal embryos, may alter the interpretation of the results, overestimating the risk assessment. To improve the reliability of this assay in establishing a causative relationship between quality of sediments and sea urchin embryonic development, here we developed and validated three Integrative Toxicity Indexes (ITI 2.0, ITI 3.0, ITI 4.0), modifying the already-known ITI (here ITI 1.0). Based on this aim, we used Taranto harbour as a model pilot-study to compare results to those obtained from standard criteria. Among the tested indexes, the ITI 4.0, discriminating strictly developmental delay and morphological defects from fertilized egg to gastrula stage, resulted in the most promising.

Keyword list: *Paracentrotus lividus*; teratogenicity; delay; elutriates; contaminants

Dear Dr. Paolo Roccaro,

Please find here the recent resubmission of our manuscript entitled "Development and validation of new analytical methods using sea urchin embryo bioassay to evaluate dredged marine sediments" Ms. Ref.No. JEMA-D-20-08004. The manuscript has been revised according to the reviewer's comments. We appreciated their work on this manuscript as both made a constructive review and appeared familiar with the topic. We did follow all of their recommendations. We hope that you will find this revised manuscript suitable for publication in the Journal of Environmental Management.

Below, point by point responses:

Reviewers' comments:

Reviewer #1:

The manuscript JEMA-D-20-08004, supplies an important tool to evaluate marine matrices in the light of environmental safety and management. The authors present an improvement of their previous work by use of sea urchin early developmental stages, by supplying three Integrative Toxicity Indexes (ITI) which can put together chemical data with ecotoxicological characterization. This work reports a modification of the already known ITI, developed previously by the same authors. The case of Taranto harbor is examined as a pilot study, aimed at demonstrating the efficiency and possible predictivity of the methods. The developmental aspects of the biological system are described and shown, so that the method is easy to understand and reproducible in other laboratories.

The work is carefully explained, the results are convincing, the references are up-to date and appropriate, the English language is fluent and easy to understand, the figures clearly show the results. Thus, in my opinion the paper deserves publication in the Journal of Environmental management.

I have just one observation:

Q1. Page 14, line 265: from gastrula to fertilized eggs: I would write "from fertilized egg to gastrula stage". At these stages, the possibility of recovery is rather higher than expected, cause of the possibility demonstrated by Giudice (1971) to reaggregate dispersed cell as soon as the environmental conditions are improved. Moreover, the plasticity of sea urchin larvae (Fenaux et al) may cause a morphogenetic damage milder than expected, also at later stages.

R1. The point raised by the reviewer is very important. As for his suggestion, the words "from gastrula to fertilized eggs" have been replaced by "from fertilized egg to gastrula stage" throughout the text (see text).

Reviewer #2: This is a very interesting work describing new analytical methods using the sea urchin embryo bioassay to evaluate dredged marine sediments. Authors evaluated three new Integrative Toxicity Indexes (ITI) in addition to a previous one reported in a study published by the same group in 2016. The method selected among the three (ITI 4.0) represents a powerful tool to assess embryotoxicity of dredged marine sediments and for other ecotoxicological assessments. This method improves the reliability of the sea urchin embryo bioassay, since it weights developmental delay and morphological defects in a balanced way.

Although the paper is well conceived and clearly written, the Authors should address some minor point before the manuscript could be published.

Q1. Par. 2.2 Information about control conditions chosen for this experiment should be given. Please add it.

R1. As suggested, we added some information about control conditions as follows (see highlighted text):

Three replicates were performed for each elutriate sample as well as for controls. **Control embryos were exposed to ASW only, for a total of nine control cultures (each in triplicates).** At 48 hours after fertilization (h), live embryos were observed and photographed using an optical microscope equipped with a digital camera (OLYMPUS CKX3). **The results were considered valid and acceptable only if each set of control presented at least 80% of normal embryos.** Formaldehyde (10% in ASW) was added to each well at the final concentration of 0.015% just prior to count and categorize embryos.

Q2. Line 166-168. It is not clear to me why in the summation notation $i=10$. Written this way it should mean that 10 is the lower bound of summation, and n is the upper bound of summation. This is the standard mathematical notation for the summation. If the meaning is here different, please specify.

R2. The point raised by the reviewer is correct. We acknowledge that the lower bound of summation notation $i=10$ was incorrect and we thank this reviewer to highlighted the mistake. Following your question, the text has been changed as follows:

"The ITIs applied in this study are calculated as follows:

$$ITI = \sum_{i=1}^n (S_i * F_i) / 100$$

Where S_i is the score associated with each abnormal embryonic morphology and F_i is the frequency observed for that abnormality expressed as a percentage (n=13 for ITI 1.0 and n=6 for ITI 2.0, ITI 3.0, and ITI 4.0)."

Q3. Line 182-183. Please indicate a reference for the threshold of 20% to be considered as significant.

R3. The reference has been added as suggested (see text)

Morroni, L., Sartori, D., Costantini, M., Genovesi, L., Magliocco, T., Ruocco, N., Buttino, I. 2019. First molecular evidence of the toxicogenetic effects of copper on sea urchin *Paracentrotus lividus* embryo development. *Water Res.* 160: 415-423. doi: 10.1016/j.watres.2019.05.062

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Highlights

- *Paracentrotus lividus* embryo bioassay was used to evaluate dredged marine sediments
- The harbour of Taranto was chosen as a model pilot-study for sediment evaluation
- New Integrative Toxicity Indices was developed and validated to improve the reliability of the sea urchin embryo assay
- ITI 4.0 weighted developmental delay and malformations in the most balanced way
- ITI 4.0 resulted in a promising tool in the quality assessment of dredged sediments

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1. Introduction

To evaluate the impact of chemical pollutants in the environment, nowadays it is widely recognized the importance to assess the biological effects of contaminants, using an integrated approach with the chemical data. The chemical approach by itself does not provide information on real bioavailability and biological risk of measured pollutants. Ecotoxicological batteries of bioassays have progressively been applied to quantify the potential biological hazard caused by bioavailable multi-factorial contamination, thus providing a more relevant response not restricted by a predetermined list of contaminants (Morrone et al., 2020). For these reasons, these tools are often included in legislative requirements. The last Italian Decree on the management of dredged sediments (DM 173/2016) foresees a list of key species to be used in a battery of bioassays to assess the sediment toxicity. This ecological risk assessment is based on a multidisciplinary weight of evidence (WOE) approach, considering chemical analyses and ecotoxicological bioassays as different lines of evidence (LOEs) through a quantitative integration. As a result of weighted elaboration, a quality classification of marine sediments and different management options were then suggested according to dumping legislation. In recent years WOE approach was validated in several case studies for environmental risk assessment associated with polluted sediments, harbour areas, or complex natural and anthropic impacts on the marine environment (Piva et al., 2011; Benedetti et al., 2014; Pittura et al., 2018; Regoli et al., 2019). Based on this aim, the ecotoxicological bioassays have a crucial role to evaluate the overall environmental quality status and to suggest appropriate management decisions. Data are typically obtained from different species, strains, exposure times, and different end-points including survival, reproduction, and growth (Picone et al., 2016). Among target species, sea urchins are worldwide considered an ideal choice for marine eco-toxicological tests as their embryos are enough sensitive to detect adverse effects related to a huge range of pollutants and natural matrices, including metals and metals mixtures (Morrone et al., 2016, 2018; Bonaventura et al., 2018), micro- and nano-plastics (Pinsino et al., 2017; Oliviero et al., 2019), UV radiation (Lister et al., 2010; Russo et al., 2014), ocean acidification (Passarelli et al 2017; Dorey et al., 2018), sediments (Khosrovyan et al., 2013; Pagano et al., 2017). Sea urchin embryos can be easily obtained in laboratory conditions, and the development to pluteus stage is completed in 24-48 hrs, depending on the species. In the DM 173/2016, the embryo-toxicity test on the mediterranean species, *Paracentrotus lividus*, is measured after 48 hours of development (ASTM, 1995; USEPA, 1995; Environment Canada, 2011), and embryos are conventionally classified in “normal” or “abnormal”, reporting the percentage of abnormally developed embryos (standard toxicity criteria). The general limit of such standard toxicity criteria is that developmental analysis does not distinguish among different malformations, block, and delay of embryogenesis. To overcome this limitation, some recent studies developed new

69 analytical indexes to weigh the teratogenic effects in the sea urchin embryos, by integrating the
70 frequency of abnormal embryos with the severity of such abnormalities (Morrone et al., 2016), or by
71 using a selective criterion such as detailed skeleton malformation (Carballeira et al., 2012). Although
72 these analytic methods result highly performant, they are less rapid and simple than a traditional
73 toxicological testing strategy based on the observation of normal vs. abnormal embryos.

74 To further improve the use of promising analytical methods to establish causative relationships
75 between contaminants and sea urchin embryonic teratogenicity or delay, here we developed and
76 tested three additional new Integrative Toxicity Indexes (ITIs) modifying the pioneer ITI published
77 by Morrone et al (2016), and comparing results to those obtained from standard toxicity approach.
78 Based on this aim, we used sea urchin embryo-toxicity data (48 hours of development as end-point)
79 generated by 43 elutriates obtained from representative sediments samples of Taranto harbour, which
80 was chosen as a model case-study. Notably, Taranto harbour was of interest because environmental
81 and epidemiological investigations in the area have provided evidence of environmental
82 contamination (e.g., particulate matter, heavy metals, polycyclic aromatic hydrocarbons, and organ-
83 halogenated compounds) (Pirastu et al 2013). The new ITIs were based on the frequency of delayed
84 and/or abnormal embryonic morphologies calculated using a simplified scale from 0 (absence of
85 toxicity) to 5 (maximum toxicity). To achieve the intended goals in terms of reliable harbour-sediment
86 hazard assessment and related risk evaluation, studies on the development of fast and reliable methods
87 become mandatory.

88 **2. Material and Methods**

89 **2.1 Sediment sampling and elutriate preparation**

90 Sediments were collected during a large characterization and monitoring project in the Taranto
91 harbour (from September 2016 to February 2017). Elutriates from 43 representative sediment samples
92 collected at different depth levels (from 0 to 150 cm) were prepared according to the guidelines
93 (USEPA 1991; APAT-ICRAM, 2007) and literature studies (Morrone et al., 2016).

94 **2.2 Sea urchin harvesting, embryonic cultures, and exposure**

95 Specimens of the sea urchin *Paracentrotus lividus* were collected along the unpolluted coast of Sicily
96 (Italy), and were brought back to the laboratory. Toxicity tests were performed following the method
97 reported by Morrone et al. (2016) with slight modifications, as described. At least three males and
98 three females were induced to spawn by injecting 0.5 M KCl into the sea urchin body cavity through
99 the peristomal membrane around the teeth. Eggs were collected by placing spawning females on 100
100 ml beakers with 0.45 µm filtered artificial seawater (ASW), while sperms were collected dry (directly
101 from the surface of the sea urchin) using a micropipette with the end of the tip cut off, maintained in

102 a sealed container at room temperature, and used in 30 minutes. Egg quality and sperm motility were
103 inspected by observing the gametes under an optical microscope (OLYMPUS CKX31). Sperms were
104 diluted in 10 ml of ASW and added to the egg suspension (10,000 eggs mL⁻¹ dilution). After
105 fertilization, embryos were maintained in a 24-well plate at the final concentration of 500
106 embryos/ml, at a temperature of 18°C. Embryos were then exposed to elutriates (1:4 ratio of sediment
107 to water), from fertilization (0 h post-fertilization) to the pluteus stage (48 h post-fertilization).
108 Three replicates were performed for each elutriate sample as well as for controls. Control embryos
109 were exposed to ASW only, for a total of nine control cultures (each in triplicates). At 48 hours after
110 fertilization (h), live embryos were observed and photographed using an optical microscope equipped
111 with a digital camera (OLYMPUS CKX3). The results were considered valid and acceptable only if
112 each set of control presented at least 80% of normal embryos. Formaldehyde (10% in ASW) was
113 added to each well at the final concentration of 0.015% just prior to count and categorize embryos.

114

115 **2.3 Toxicity criteria**

116

117 The toxicity of elutriate samples from Taranto harbour was estimated by calculating the percentage
118 of abnormal embryos at pluteus stage (48 h of development), according to the standard criteria, the
119 Integrative Toxicity Index (ITI) published by Morroni et al (2016) (here called ITI 1.0), and three
120 additional new ITIs developed from ITI 1.0, respectively called ITI 2.0, ITI 3.0 and ITI 4.0. The
121 standard criteria calculate the percentage of normal and abnormal embryos without considering
122 different typologies of malformations or the phase in which they appear. Conversely, all the ITI
123 methodologies count the frequency of delayed and/or abnormal embryonic morphologies and
124 quantitatively rank the severity of effects.

125 Embryos were classified as normal if they satisfied the morphological criteria as follows: i) suitable
126 schedule in reaching the developmental endpoint (pluteus at 48 h); ii) correct skeleton development
127 and patterning; iii) right ectoderm, mesoderm, and endoderm germ layer differentiation; iv) conform
128 left/right or dorso/ventral axis symmetry. On the other hand, embryos displaying impairment of the
129 axis symmetry, as well as germ layer defects were marked as abnormal (see **Table 1** showing
130 representative images of normal and abnormal *P. lividus* embryos at different embryonic
131 developmental stages).

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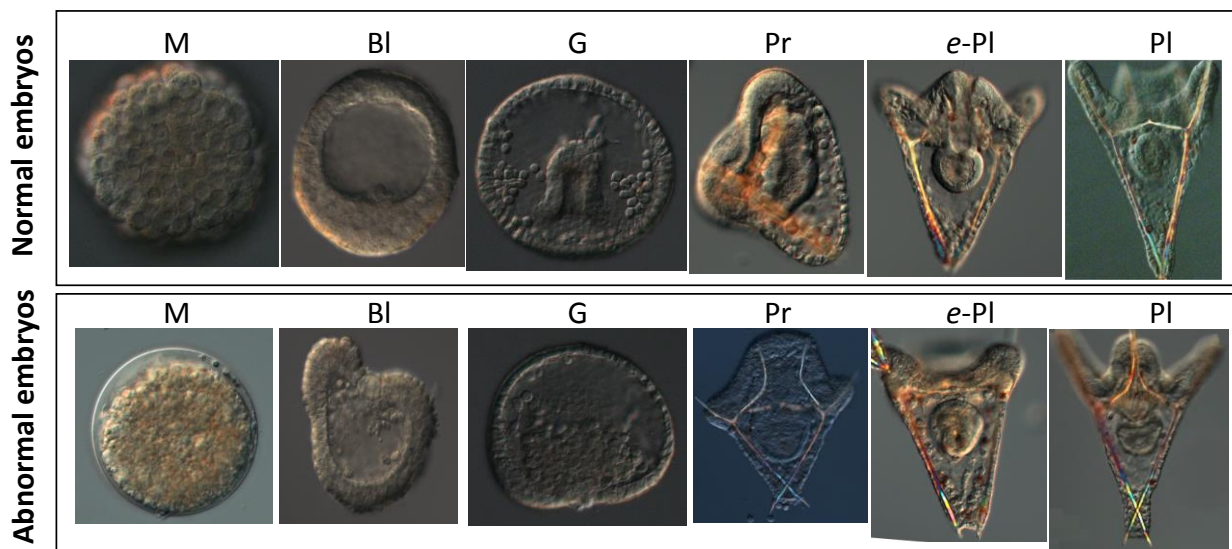
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Table 1 –Developmental stage and abnormalities of *P. lividus* observed in this study



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145 M, Morula; Bl, Blastula; G, Gastrula; Pr, Prism; ePl, early Pluteus; Pl, Pluteus

146 The ITI 1.0 from the previous study used a toxicity scale from 0 (absence of toxicity) to 10 (maximum
147 toxicity), which was here implemented with a second generation of ITIs, using more simplified
148 criteria, grouping embryos on a toxicity scale from 0 (absence of toxicity) to 5 (maximum toxicity)
149 as shown in **Table 2**. The lowest level (0) assigned to each ITIs was associated with a “no effect” in
150 development, including only normal embryos reaching the 48-h endpoint (Pluteus). The only
151 exception was the ITI 4.0 where the zero-effect level was extended also to those embryos displaying
152 a slight delay, considered as a negligible effect (pluteus and early pluteus). A score ranging from 1 to
153 3 was assigned to the ITI 2.0 and ITI 3.0 as follows: 1 for delayed embryos at the pluteus stage (*ePl*);
154 2 for malformed embryos at the pluteus stage (*mPl*); and 3 for delayed embryos at the pluteus stage
155 displaying malformations (*emPl*). For ITI 2.0, level 4 was associated with the delayed embryos from
156 fertilized egg to prism (F-Pr) and level 5 with the malformed F-Pr (F-*mPr*); in ITI 3.0 the levels 4
157 and 5 were assigned to include delayed and/or malformed Pr (Pr and/or *mPr*), and delayed and/or
158 malformed embryos from fertilized egg to gastrula stage (F-G and/or F-*mG*), respectively. On the
159 other hand, the score 1 assigned to the ITI 4.0 was associated with the Pl and *e-Pl* displaying
160 malformations (*mPl* and *emPl*), the score 2 was associated with the stage of the prism (Pr), and the

161 score 3 with the Pr displaying malformations (*mPr*). The highest levels of toxicity (4 and 5) assigned
 162 to the ITI 4.0 were associated with the delayed and/or malformed embryos at the gastrula stage (G
 163 and/or *mG*), and with the delayed and/or malformed embryos from fertilized egg to blastula stage (F-
 164 Bl and/or F-*mBl*). Therefore, the lower degree of toxicity was assigned to Pl with the absence of
 165 abnormalities, while the higher degree was attributed to embryos displaying severe delay, and/or
 166 delay *plus* abnormalities simultaneously.

167 The ITIs applied in this study are calculated as follows:

168
$$ITI = \sum_{i=1}^n (S_i * F_i) / 100$$

169 Where S_i is the score associated with each abnormal embryonic morphology and F_i is the frequency
 170 observed for that abnormality expressed as a percentage (n=13 for ITI 1.0 and n=6 for ITI 2.0, ITI
 171 3.0, and ITI 4.0).

172

173 **Table 2** – Integrative Toxicity Indexes (ITIs) tested in this study

Toxicity categories													
ITI 1.0	Normal		Delayed					Malformed					
	Pl	<i>ePl</i>	Pr	G	<i>mBl</i>	Bl	M	Pl	Pr	G	<i>mBl</i>	Bl	M
Score	0	2	3	4	4.5	5	5.5	6	7	7.5	8	9	10
ITI 2.0	Normal		Delayed			Delayed and Malformed							
	Pl	<i>ePl</i>	F-Pr			Pl	<i>ePl</i>	F-Pr					
Score	0	1	4			2	3	5					
ITI 3.0	Normal		Delayed		Malformed		Delayed and/or Malformed						
	Pl	<i>ePl</i>	Pl	<i>ePl</i>	Pl	<i>ePl</i>	Pr		F-G				
Score	0	1	2	3	4	5	4		5				
ITI 4.0	Normal		Delayed		Malformed		Delayed and/or Malformed						
	Pl/ <i>ePl</i>	Pr	Pl/ <i>ePl</i>	Pr	G		F-Bl						
Score	0	2	1	3	4		5						

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175 F: fertilized egg; M, Morula; *mBl*, mesenchyme Blastula; Bl, Blastula; G, Gastrula; Pr, Prism; *ePl*, early Pluteus; Pl,
 176 Pluteus. F-Bl, from fertilized egg to Blastula; F-G: from fertilized egg to Gastrula; F-Pr: from fertilized egg to Prism.

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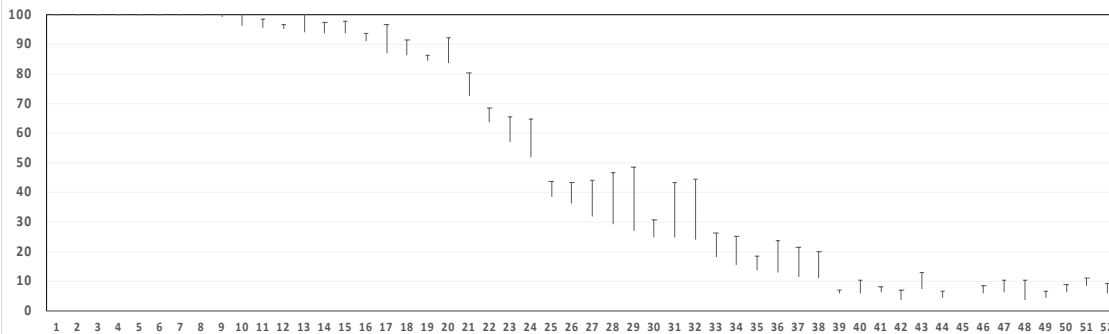
178 **3. Results and Discussion**

179 The evaluation based on the standard criteria and the thoughtful ITIs (from ITI 1.0 to ITI 4.0) of the
 180 impact on the sea urchin embryonic development is shown in Figure 1. Based on the standard criteria,
 181 the majority of tested elutriates (46.5%) presented extremely severe or severe toxicity with a
 182 percentage of abnormal embryos higher than 75% (20 of 43 samples; Figure 1A, red bars); a moderate
 183 number (28%) displayed from severe to moderate toxicity with a percentage of abnormal embryos
 184 ranging from 75% to 20% (12 of 43; Figure 1A, blue bars); the remaining 25.5% did not show any

185 significant impact compared to the controls (11 of 43; figure 1A, compare green bars with those grey)
186 being under the threshold of 20% (Morrone et al., 2019).
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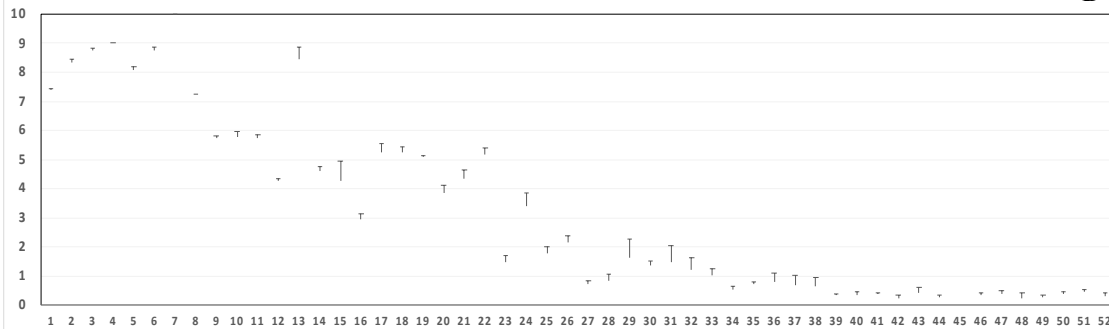
ABNORMAL EMBRYOS (%)

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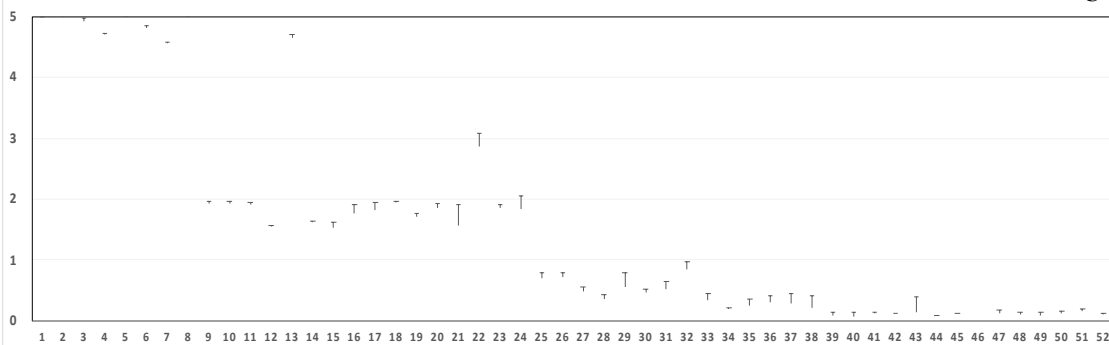
ITI 1.0

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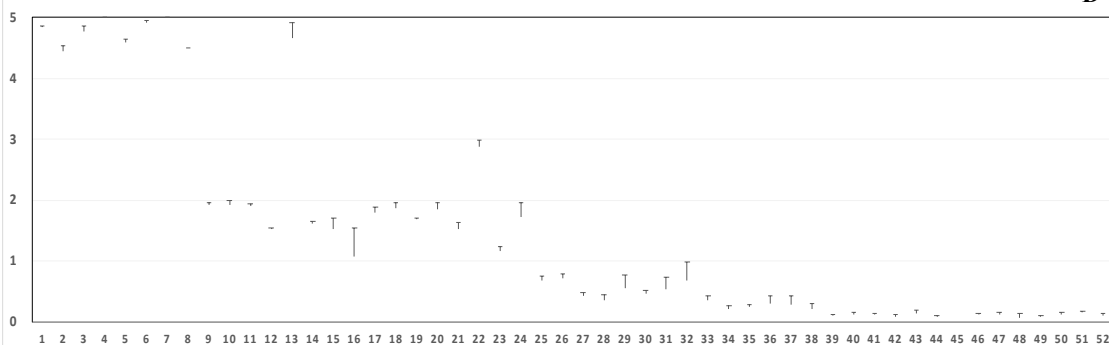
ITI 2.0

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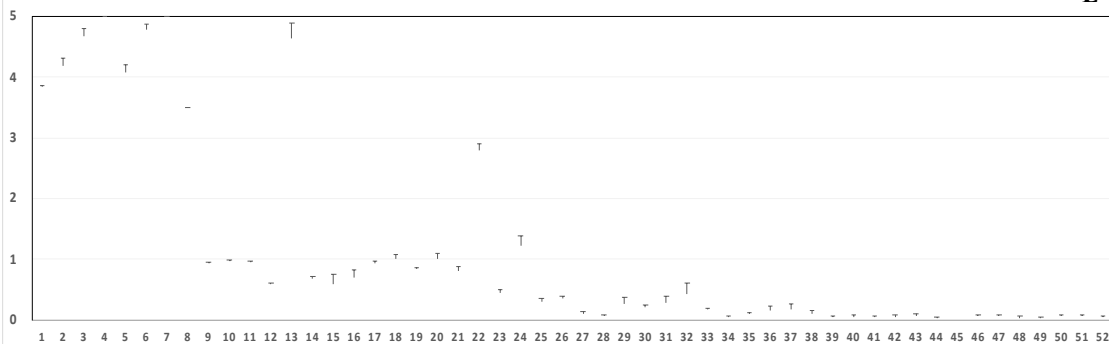
ITI 3.0

D



ITI 4.0

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Figure 1. Sea urchin embryonic development evaluated according to standard criteria and thoughtful ITIs. Histograms represent the results expressed as a mean percentage (%) of abnormal embryos \pm SD (A), and values of the ITI 1.0 (B), ITI 2.0 (C), ITI 3.0 (D), ITI 4.0 (E) \pm SD. Data are referred to each elutriate sample of Taranto harbour, reported on X-axis. Controls (CTR) are reported in the right part of the panels for a total of 9. Red bars: percentage of abnormal embryos higher than 75%; Blue bars: percentage of abnormal embryos ranging from 75% to 20%; Green bars: percentage of abnormal embryos lower than 20%; Grey bars: controls.

The evaluation of embryo-toxicity based on the ITI 1.0 method, which discriminates the frequency of delayed and/or abnormal embryonic morphologies and quantitatively ranks the severity of effects on a pondered scale from 0 (absence of toxicity) to 10 (maximum toxicity), leads to an attenuate scenario of morbidity (Figure 1B). The difference appeared more accentuated for those samples classified as extremely toxic by standard criteria (Figure 1B, red bars): only one sample (number 7) confirmed the maximum level of toxicity (10), while eight samples presented a level ranging from 7 to 9 (number 1-6, 8, and 13), six samples ranged from 5 to 6 (moderate toxicity, number 9-11, 17-19), three samples from 4 to 5 (low toxicity, number 12, 14-15), and two samples were below the value 4 (very low toxicity, number 16, and 20). Therefore, the extremely toxic effects assessed by the standard method for 20 sediment samples were confirmed by ITI 1.0 only for 9 of them (from level 7 to level 10), corresponding to about 50% of the cases. In agreement with our recent report, this result provides evidence that the ITI 1.0 allows to better separate the samples according to the frequency and severity of delayed and/or abnormal morphologies (Morrone et al., 2016). ITI 1.0 is more sensitive than traditional toxicological testing strategy but presents the disadvantage that the rigorous morphological analysis may be applied by trained personnel on the sea urchin embryonic development: in this respect, for unspecialized operators, it may result less rapid, simple, and direct than conventional methods based on observation of normal vs. abnormal embryos only. To simplify the promising approach of ITI for determining a more realistic toxicological evaluation of dredged sediments, we tested the second generation of indices still based on the frequency of delayed and/or abnormal embryonic morphologies but calculated using a simplified scale, in which embryos are grouped on a toxicity scale from 0 (absence of toxicity) to 5 (maximum toxicity) (see Table 2, Material and Methods). Based on this aim, scores assigned by ITI 2.0 range from 1 to 5 as follows: 1 for delayed plutei (*ePl*); 2 for malformed plutei (*mPl*); 3 for delayed and malformed plutei (*emPl*); 4 for delayed embryos from fertilized egg to prism (F-Pr); 5 for malformed embryos from fertilized egg to prism (F-*mPr*). The evaluation of embryo-toxicity of Taranto elutriates based on the ITI 2.0 is reported in Figure 1C. In agreement with the ITI 1.0 results, extremely toxic effects evaluated by standard criteria for the 20 samples were confirmed only for 9 of them (from level 4 to 5); the remaining 11 samples displayed low toxicity with levels below 2. Even if the ITI 2.0 appear less

227 discriminating than ITI 1.0, this index displayed a similar level of performance in samples classified
228 as extremely toxic (compare Figure 1B and 1C).

229 The ITI 3.0 showed a trend comparable to ITI 2.0, with the difference that levels below 2 were much
230 less flattened (Figure 1D). The ITI 3.0 assign the score from 1 to 3 as for ITI 2.0 (1 for *ePl*; 2 for *mPl*;
231 3 for *emPl*), whereas different criteria have been used for attributing values of 4 and 5: 4 for delayed
232 and/or malformed prisms (*emPr*), and 5 for delayed and/or malformed embryos from fertilized egg to
233 gastrula (F-*emG*). These differences increased the ability of ITI 3.0 to discriminate among groups
234 compared to ITI 2.0.

235 Considering the results of the first six samples, the toxic levels were, on average, 5% lower than those
236 obtained from ITI 2.0. This slight increase in sensitivity was mostly observed in sample number 2
237 and 5, with 13% and 5% of the embryos at the prism stage, respectively (not shown). This stage was
238 not well discriminated in ITI 2.0 as score 4 is assigned to embryos at the stages from fertilized egg
239 (F) to prism (Pr).

240 Finally, when we used the ITI 4.0, discriminating strictly developmental delay and morphological
241 defects from fertilized egg to gastrula (1 for *emPl*; 2 for *ePr*; 3 for *mPr*; 4 for *emG*; 5 for F-*emBl*), we
242 still increased the ability to discriminate among groups (see Figure 1E). Notably, several samples
243 showed lower and more distributed values of toxicity, such as samples number 1, 2, 5, and 8 (ITI 4.0
244 values ranging from 3.5 to 4.2). Other samples, as the number 7, maintained the maximum level of
245 toxicity (5), confirming the good discriminatory ability of this index, which consider the different
246 degree of severity assigned at early stages, discriminating between gastrula and pre-gastrula stages
247 (from fertilized egg to blastula) (see Table 2).

248 The sea urchin embryo is a simple model to monitor the developmental stages from fertilization to
249 pluteus stage; *Paracentrotus lividus*, under controlled conditions of temperature (18°C) reaches the
250 pluteus stage after 48 h. The embryonic development requires a prompt and synchronised
251 combination of cell proliferation, fate specification, and movement, controlled by gene regulatory
252 networks (Erkenbrack et al., 2018). Cell fate is specified at the appropriate space and time (blastula-
253 early gastrula stage of development) when cells become able to express a set of differentiated germ
254 layer-exclusive genes (Davidson et al., 1998). Elevated metabolic rates decrease capability for
255 growth, and promote developmental delay of the sea urchin embryos; for example, this happens under
256 acidified seawater conditions (Stumpp et al., 2011).

257 The most documented explanation on sea urchin embryonic delay as an effect of toxicity is the
258 reduction in the ability to uptake calcium and, in turn, to maintain intracellular homeostasis related to
259 a low extracellular pH; calcium-contaminant trafficking competition also affects the normal gene
260 regulatory network controlling development (Stumpp et al., 2011; Pinsino et al., 2011). Several

261 developmental steps such as fertilization, cleavage, neuronal development, skeletogenesis, cell death,
262 and body modelling are known to be dependent on calcium ion trafficking (Webb and Miller 2003).
263 On the other hand, regulatory studies reveal that the embryos present an early sequence of encoded
264 “fail-safe” regulatory devices (Smith and Davidson 2009). Based on this evidence, we speculate that
265 in the early embryonic stages (from fertilized egg to gastrula), when the cell fate has not been
266 specified yet, the probability for embryos to recover from the delay and continue the development is
267 scant, thereby justifying the assignment of the higher degree of toxicity to embryos displaying severe
268 delay, abnormalities, or delay plus abnormalities occurring simultaneously. On the contrary, at the
269 late embryonic phases (from prism to early pluteus), when the cell fate is already specified, the
270 embryos have a high probability to continue the development, thus explaining the assignment of the
271 lower degree of toxicity. The increasing grading of mild, moderate, and severe effects assigned to the
272 severity of delay and teratogenicity, was progressively emphasized from the ITI 2.0 to the ITI 4.0.
273 All these indexes can be considered valid tools to better evaluate the embryotoxicity effects on sea
274 urchin based on objective and solid scientific criteria, with clearly important applicative consequences
275 when assessing the quality of dredged marine sediments: among the various indexes, ITI 4.0 which
276 stress mainly the severity of delay, offers the higher sensitivity and discriminatory efficiency.

277 **4. Conclusions**

278 The use of the WOE integration, which combines and weight different typologies of data and
279 analyses, allows to better discriminate the presence of contaminants and their short or long-term
280 consequences, especially when apparently contrasting results are provided by various LOEs. The
281 possibility to convert complex scientific information into simple hazard indexes, easily
282 understandable by policymakers and environmental managers, can facilitate and orientate the more
283 appropriate and site-specific decisions on environmental sediment management (Morrone et al.,
284 2020). In this context, the sea urchin embryo-toxicity bioassay is considered an essential component
285 for evaluating the quality of sediments in harbour areas, with important environmental and economic
286 implications. Classifications based on the worst result are still in use and significant consequences
287 may arise depending on the choice of the ecotoxicological assays within a battery. In particular, this
288 study demonstrated that care should be taken in the evaluation of embryo-toxicity results suggesting
289 weight developmental delay and morphological defects in a balanced way.

290 The development of such a sensitive method is of great utility to properly achieve a reliable harbour-
291 sediment hazard assessment and related risk evaluation. The results obtained in the present study
292 indicate that ITI 4.0 is a promising approach for dredged sediments, better discriminating samples
293 with intermediate toxicity from those highly toxic.

294

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299

300 **CRedit authorship contribution statement**

301 **Rosa Bonaventura:** Investigation, Data Curation; Writing-Original Draft, Project administration;
302 **Francesca Zito:** Investigation; **Lorenzo Morroni:** Software, Formal analysis, Writing-Review &
303 Editing; **David Pellegrini:** Writing-Review & Editing; **Francesco Regoli:** Writing-Review &
304 Editing, Validation, Software; **Annalisa Pinsino:** Conceptualization, Investigation, Data Curation,
305 Writing-Original Draft, Writing-Review & Editing; Project administration, Funding acquisition.

306 **Declaration of Competing Interest**

307 The authors declare no competing financial interest.

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310 **References**

- 311 1) APAT-ICRAM, 2007. Manuale per la movimentazione di sedimenti marini, pp. 1–67. Available
312 at <www.apat.gov.it> and <www.icram.org >
313
- 314 2) ASTM E1563, 1995. Standard guide for conducting static acute toxicity tests with echinoid
315 embryos. E 1563-95 In Annual Book of ASTM Standards, Philadelphia, PA, vol. 11(5), pp. 999–
316 1017.
317
- 318 3) Benedetti, M., Gorbi, S., Fattorini, D., D’Errico, G., Piva, F., Pacitti, D., Regoli, F., 2014.
319 Environmental hazards from natural hydrocarbons seepage: Integrated classification of risk from
320 sediment chemistry, bioavailability and biomarkers responses in sentinel species. Environ. Pollut.
321 185, 116–126. <https://doi.org/10.1016/j.envpol.2013.10.023>
322
- 323 4) Bonaventura R, Zito F, Chiaramonte M, Costa C, Russo R (2018). Nickel toxicity in *P. lividus*
324 embryos: Dose dependent effects and gene expression analysis. Mar Environ Res. 2018, 139:113-
325 121. <https://doi.org/10.1016/j.marenvres.2018.05.002>
326
- 327 5) Carballeira, C., Ramos-Gómez, J., Martín-Díaz, L., DelValls, T.A., 2012. Identification of
328 specific malformations of sea urchin larvae for toxicity assessment: application to marine
329 pisciculture effluents. Mar Environ Res. 77, 12-22.
330 <https://doi.org/10.1016/j.marenvres.2012.01.001>
331
- 332 6) Davidson, E. H., Cameron, R. A., and Ransick, A. (1998). Specification of cell fate in the sea
333 urchin embryo: summary and some proposed mechanisms. Development. 125, 3269–3290.
334
- 335 7) DM 173/2016. Ministero dell’Ambiente e della Tutela del Territorio e del Mare, Supplemento
336 ordinario alla Gazzetta Ufficiale, n. 208 del 6 settembre 2016-Serie generale. Regolamento
337 recante modalità e criteri tecnici per l’autorizzazione all’immersione in mare dei materiali di
338 escavo di fondali marini.

- 339
340 8) Dorey, N., Martin, S., Oberhänsli, F., Teyssié, J.L., Jeffree, R., Lacoue-Labarthe, T., 2018. Ocean
341 acidification modulates the incorporation of radio-labeled heavy metals in the larvae of the
342 Mediterranean sea urchin *Paracentrotus lividus*. *J Environ Radioact.* 190-191, 20-30.
343 <https://doi.org/10.1016/j.jenvrad.2018.04.017>
344
- 345 9) Environment Canada, 2011. Biological Test Methods: Fertilization Assay using Echinoids (Sea
346 Urchins and Sand Dollars). EPS 1/RM/27.
347
- 348 10) Erkenbrack, E.M., Davidson, E.H., Peter, I.S., 2018. Conserved regulatory state expression
349 controlled by divergent developmental gene regulatory networks in echinoids. *Development*,
350 145(24). doi: 10.1242/dev.167288
351
- 352 11) Khosrovyan, A., Rodríguez-Romero, A., Salamanca, M.J., DelValls, T.A., Riba, I., Serrano, F.,
353 2013. Comparative performances of eggs and embryos of sea urchin (*Paracentrotus lividus*) in
354 toxicity bioassays used for assessment of marine sediment quality. *Mar. Pollut. Bull.* 70(1-2),
355 204-209. <https://doi.org/10.1016/j.marpolbul.2013.03.006>
356
- 357 12) Lister, K.N., Lamare, M.D., Burritt, D.J., 2010. Oxidative damage in response to natural levels of
358 UV-B radiation in larvae of the tropical sea urchin *Tripneustes gratilla*. *Photochem. Photobiol.*
359 86, 1091-1098. <https://doi.org/10.1111/j.1751-1097.2010.00779.x>
360
- 361 13) Morroni, L., Sartori, D., Costantini, M., Genovesi, L., Magliocco, T., Ruocco, N., Buttino, I. 2019.
362 First molecular evidence of the toxicogenetic effects of copper on sea urchin *Paracentrotus lividus*
363 embryo development. *Water Res.* 160: 415-423. doi: 10.1016/j.watres.2019.05.062
364
- 365 14) Morroni, L., d'Errico, G., Sacchi, M., Molisso, F., Armiento, G., Chiavarini, S., Rimauro, J.,
366 Guida, M., Siciliano, A., Ceparano, M., Carraturo, F., Tosti, E, Gallo, A., Libralato, G., Patti,
367 F.P., Gorbi, S., Fattorini, D., Nardi, A., Di Carlo, M., Mezzelani, M., Benedetti, M., Pellegrini,
368 D., Musco, L., Danovaro, R., Dell'Anno, A., Regoli, F., 2020. Integrated characterization and risk
369 management of marine sediments: the case study of the industrialized Bagnoli area (Naples, Italy).
370 *Mar. Environ. Res.*, 160, 104984, <https://doi.org/10.1016/j.marenvres.2020.104984>
371
- 372 15) Morroni, L., Pinsino, A., Pellegrini, D., Regoli, F., Matranga, V., 2016. Development of a new
373 integrative toxicity index based on an improvement of the sea urchin embryo toxicity test.
374 *Ecotoxicol. Environ. Saf.* 123, 2–7. <https://doi.org/10.1016/j.ecoenv.2015.09.026>
375
- 376 16) Morroni, L., Pinsino, A., Pellegrini, D., Regoli, F., 2018. Reversibility of metal induced
377 malformations in sea urchin embryos. *Ecotox. Environ. Saf.* 148, 923-929.
378 <https://doi.org/10.1016/j.ecoenv.2017.11.013>
379
- 380
- 381 17) Oliviero, M., Tato, T., Schiavo, S., Fernández, V., Manzo, S., Beiras, R., 2019. Leachates of
382 micronized plastic toys provoke embryotoxic effects upon sea urchin *Paracentrotus lividus*.
383 *Environ. Pollut.* 247, 706–715. <https://doi.org/10.1016/j.envpol.2019.01.098>
384
- 385 18) Pagano, G., Thomas, P.J., Guida, M., Palumbo, A., Romano, G., Trifuoggi, M. Oral, R.,
386 Trifuoggi, M., 2017. Sea Urchin Bioassays in Toxicity Testing: II. Sediment Evaluation. *Expert*
387 *Opin. Environ. Biol.* 6, 1. doi: 10.4172/2325-9655.1000141
388

- 389 19) Pinsino, A., Roccheri, M.C., Costa, C., Matranga, V., 2011. Manganese interferes with calcium,
390 perturbs ERK signaling, and produces embryos with no skeleton. *Toxicol Sci.*, 123, 217-30.
391 <https://doi.org/10.1093/toxsci/kfr152>
392
- 393 20) Pinsino, A., Bergami, E., Della Torre, C., Vannuccini, M.L., Addis, P., Secci, M., Dawson, K.A.,
394 Matranga, V., Corsi, I., 2017. Amino-modified polystyrene nanoparticles affect signaling
395 pathways of the sea urchin (*Paracentrotus lividus*) embryos. *Nanotoxicology*, 11, 201-209. DOI:
396 10.1080/17435390.2017.1279360
397
- 398 21) Passarelli, M.C., Cesar, A., Riba, I., DelValls, T.A. 2017. Comparative evaluation of sea-urchin
399 larval stage sensitivity to ocean acidification. *Chemosphere* 184, 224–234. DOI:
400 10.1016/j.chemosphere.2017.06.001
401
- 402 22) Picone, M., Bergamin, M., Losso, C., Delaney, E., Arizzi Novelli, A., Ghirardini, A.V., 2016.
403 Assessment of sediment toxicity in the Lagoon of Venice (Italy) using a multi-species set of
404 bioassays. *Ecotoxicol. Environ. Saf.* 123, 32-44, <https://doi.org/10.1016/j.ecoenv.2015.09.002>
405
- 406 23) Pirastu, R., Comba, P., Iavarone, I., Zona, A., Conti, S., Minelli, G., Manno, V., Mincuzzi, A.,
407 Minerba, S., Forastiere, F., Mataloni, F., Biggeri A., 2013. Environment and Health in
408 Contaminated Sites: The Case of Taranto, Italy. *Int J Environ Res Public Health* 3, 753719.
409 <https://doi.org/10.1155/2013/753719>
410
- 411 24) Pittura, L., Avio, C.G., Giuliani, M.E., D'Errico, G., Keiter, S.H., Cormier, B., Gorbi, S., Regol,
412 F., 2018. Microplastics as vehicles of environmental PAHs to marine organisms: Combined
413 chemical and physical hazards to the mediterranean mussels, *Mytilus galloprovincialis*. *Front.*
414 *Mar. Sci.* 5(103), <https://doi.org/10.3389/fmars.2018.00103>
415
- 416 25) Piva, F., Ciapriani, F., Onorati, F., Benedetti, M., Fattorini, D., Ausili, A., Regoli, F., 2011.
417 Assessing sediment hazard through a weight of evidence approach with bioindicator organisms:
418 A practical model to elaborate data from sediment chemistry, bioavailability, biomarkers and
419 ecotoxicological bioassays. *Chemosphere* 83, 475–485,
420 <https://doi.org/10.1016/j.chemosphere.2010.12.064>
421
- 422 26) Regoli, F., D'Errico, G., Nardi, A., Mezzelani, M., Fattorini, D., Benedetti, M., Di Carlo, M.,
423 Pellegrini, D., Gorbi, S., 2019. Application of a Weight of Evidence Approach for Monitoring
424 Complex Environmental Scenarios: the Case-Study of Off-Shore Platforms. *Front. Mar. Sci.* 6,
425 1–15. <https://doi.org/10.3389/fmars.2019.00377>
426
- 427 27) Russo, R., Bonaventura, R., Matranga, V., 2014. Time-and dose-dependent gene expression in
428 sea urchin embryos exposed to UVB. *Mar. Environ. Res.* 93, 85–92.
429 <https://doi.org/10.1016/j.marenvres.2013.08.006>
430
- 431 28) Smith, J., Davidson, E.H., 2009. Regulative recovery in the sea urchin embryo and the stabilizing
432 role of fail-safe gene network wiring. *Proc Natl Acad Sci U S A.*, 106, 18291-18296.
433 <https://doi.org/10.1073/pnas.0910007106>
434
- 435 29) Stumpp, M., Wren, J., Melzner, F., Thorndyke, M.C., Dupont, S.T., 2011. CO2 induced seawater
436 acidification impacts sea urchin larval development I: elevated metabolic rates decrease scope for
437 growth and induce developmental delay. *Comp Biochem Physiol A Mol Integr Physiol.*, 160,
438 331-340. <https://doi.org/10.1016/j.cbpa.2011.06.022>
439

- 440 30) USEPA, 1991. Evaluation of dredged material proposed for ocean disposal testing manual. EPA
441 503/8-91/001.
442
- 443 31) USEPA, 1995. Short-term methods for estimating the chronic toxicity of effluents and receiving
444 waters to west coast marine and estuarine organisms. EPA/600/ R-95/136.
445
- 446 32) Webb, S.E., Miller, A.L., 2003. Calcium signalling during embryonic development. *Nat Rev Mol*
447 *Cell Biol.*, 4, 539-551. DOI: 10.1038/nrm1149
448
449

1 **Development and validation of new analytical methods using sea urchin embryo bioassay to**
2 **evaluate dredged marine sediments**

3

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15

16 **Abstract (200 words)**

17 Management of dredged materials disposal is regulated ~~by in accordance to~~ several environmental
18 normative requirements, and it is often supported by ~~the~~ integration of chemical data with
19 ecotoxicological characterization. The reliability of a bioassay to assess the potential toxicity of
20 dredged sediments requires the selection of quality criteria that should be based on simple analytical
21 methods and easily understandable hazard for politicians and environmental managers. The sea urchin
22 embryo-toxicity bioassay is considered an essential component for evaluating the quality of sediments
23 in harbour areas but its use, when based exclusively on the observation of normal vs. abnormal
24 embryos, may alter the interpretation of the results, overestimating the risk assessment. To improve
25 the reliability of this assay in establishing ~~a~~ causative relationship between quality of sediments and
26 sea urchin embryonic development, here we developed and validated three Integrative Toxicity
27 Indexes (ITI 2.0, ITI 3.0, ITI 4.0), modifying the already-known ITI (here ITI 1.0). Based on this aim,
28 we used Taranto harbour as a model pilot-study to compare results to those obtained from standard
29 criteria. Among the tested indexes, the ITI 4.0, discriminating strictly developmental delay and
30 morphological defects from ~~fertilized egg to gastrula stage~~~~gastrula to fertilized eggs~~, resulted ~~in~~ the
31 most promising.

32

33

34 **Keyword list:** *Paracentrotus lividus*; teratogenicity; delay; elutriates; contaminants

35

36 **1. Introduction**

37 To evaluate the impact of chemical pollutants in the environment, nowadays it is widely recognized
38 the importance to assess the biological effects of contaminants, using an integrated approach with the
39 chemical data. ~~In fact,~~ ~~†~~The chemical approach by itself does not provide information on real
40 bioavailability and biological risk of measured pollutants. Ecotoxicological batteries of bioassays
41 have progressively been applied to quantify the potential biological hazard caused by bioavailable
42 multi-factorial contamination, thus providing a more relevant response not restricted by a
43 predetermined list of contaminants (Morrone et al., 2020). For these reasons, these tools are often
44 included in legislative requirements. The last Italian Decree on the management of dredged sediments
45 (DM 173/2016) foresees a list of key species to be used in a battery of bioassays to assess the sediment
46 toxicity. This ecological risk assessment is based on a multidisciplinary weight of evidence (WOE)
47 approach, considering chemical analyses and ecotoxicological bioassays as different lines of evidence
48 (LOEs) through a quantitative integration. As a result of weighted elaboration, a quality classification
49 of marine sediments and different management options were then suggested according to dumping
50 legislation. In recent years WOE approach was validated in several case studies for environmental
51 risk assessment associated with polluted sediments, harbour areas, or complex natural and anthropic
52 impacts on the marine environment (Piva et al., 2011; Benedetti et al., 2014; Pittura et al., 2018;
53 Regoli et al., 2019). Based on this aim, the ecotoxicological bioassays have a crucial role to evaluate
54 the overall environmental quality status, and to suggest appropriate management decisions. Data are
55 typically obtained from different species, strains, exposure times, and different end-points including
56 survival, reproduction, and growth (Picone et al., 2016). Among target species, sea urchins are
57 worldwide considered an ideal choice for marine eco-toxicological tests as their embryos are enough
58 sensitive to detect adverse effects related to a huge range of pollutants and natural matrices, including
59 metals and metals mixtures (Morrone et al., 2016, 2018; Bonaventura et al., 2018), micro- and nano-
60 plastics (Pinsino et al., 2017; Oliviero et al., 2019), UV radiation (Lister et al., 2010; Russo et al.,
61 2014), ocean acidification (Passarelli et al 2017; Dorey et al., 2018), sediments (Khosrovyan et al.,
62 2013; Pagano et al., 2017). Sea urchin embryos can be easily obtained in laboratory conditions, and
63 the development to pluteus stage is completed in 24-48 hrs, depending on the species. In the DM
64 173/2016, the embryo-toxicity test on the mediterranean species, *Paracentrotus lividus*, is measured
65 after 48 hours of development (ASTM, 1995; USEPA, 1995; Environment Canada, 2011), and
66 embryos are conventionally classified in “normal” or “abnormal”, reporting the percentage of
67 abnormally developed embryos (standard toxicity criteria). The general limit of such standard toxicity

68 criteria is that developmental analysis does not distinguish among different malformations, block,
69 and delay of embryogenesis. ~~To~~~~in order to~~ overcome this limitation, some recent studies developed
70 new analytical indexes to weigh the teratogenic effects in the sea urchin embryos, by integrating the
71 frequency of abnormal embryos with the severity of such abnormalities (Morrone et al., 2016), or by
72 using a selective criterion such as detailed skeleton malformation (Carballeira et al., 2012). Although
73 these analytic methods result highly performant, they are less rapid and simple than a traditional
74 toxicological testing strategy based on the observation of normal *vs.* abnormal embryos.

75 To further improve the use of promising analytical methods to establish causative relationships
76 between contaminants and sea urchin embryonic teratogenicity or delay, here we developed and
77 tested three additional new Integrative Toxicity Indexes (ITIs) modifying the pioneer ITI published
78 by Morrone et al (2016), and comparing results to those obtained from standard toxicity approach.
79 Based on this aim, we used sea urchin embryo-toxicity data (48 hours of development as end-point)
80 generated by 43 elutriates obtained from representative sediments samples of Taranto harbour, which
81 was chosen as a model case-study. Notably, Taranto harbour was of interest because environmental
82 and epidemiological investigations in the area have provided evidence of environmental
83 contamination (e.g., particulate matter, heavy metals, polycyclic aromatic hydrocarbons, and organ-
84 halogenated compounds) (Pirastu et al 2013). The new ITIs were based on the frequency of delayed
85 and/or abnormal embryonic morphologies calculated using a simplified scale from 0 (absence of
86 toxicity) to 5 (maximum toxicity). To achieve the intended goals in terms of reliable harbour-sediment
87 hazard assessment and related risk evaluation, studies on the development of fast and reliable methods
88 become mandatory.

89 **2. Material and Methods**

90 **2.1 Sediment sampling and elutriate preparation**

91 Sediments were collected during a large characterization and monitoring project in the Taranto
92 harbour (from September 2016 to February 2017). Elutriates from 43 representative sediment samples
93 collected at different depth levels (from 0 to 150 cm) were prepared according to the guidelines
94 (USEPA 1991; APAT-ICRAM, 2007) and literature studies (Morrone et al., 2016).

95 **2.2 Sea urchin harvesting, embryonic cultures, and exposure**

96 Specimens of the sea urchin *Paracentrotus lividus* were collected along the unpolluted coast of Sicily
97 (Italy), and were brought back to the laboratory. Toxicity tests were performed following the method
98 reported by Morrone et al. (2016) with slight modifications, as described. At least three males and
99 three females were induced to spawn by injecting 0.5 M KCl into the sea urchin body cavity through
100 the peristomal membrane around the teeth. Eggs were collected by placing spawning females on 100

101 ml beakers with 0.45 μm filtered artificial seawater (ASW), while sperms were collected dry (directly
102 from the surface of the sea urchin) using a micropipette with the end of the tip cut off, maintained in
103 a sealed container at room temperature, and used in 30 minutes. Egg quality and sperm motility were
104 inspected by observing the gametes under an optical microscope (OLYMPUS CKX31). Sperms were
105 diluted in 10 ml of ASW and added to the egg suspension (10,000 eggs mL^{-1} dilution). After
106 fertilization, embryos were maintained in a 24-well plate at the final concentration of 500
107 embryos/ml, at a temperature of 18°C. Embryos were then exposed to elutriates (1:4 ratio of sediment
108 to water), from fertilization (0 h post-fertilization) to the pluteus stage (48 h post-fertilization).
109 Three replicates were performed for each elutriate sample as well as for controls. Control embryos
110 were exposed to ASW only, for a total of nine control cultures (each in triplicates). At 48 hours after
111 fertilization (h), live embryos were observed and photographed using an optical microscope equipped
112 with a digital camera (OLYMPUS CKX3). The results were considered valid and acceptable only if
113 each set of control presented at least 80% of normal embryos. Formaldehyde (10% in ASW) was
114 added to each well at the final concentration of 0.015% just prior to count and categorize embryos.

115 **2.3 Toxicity criteria**

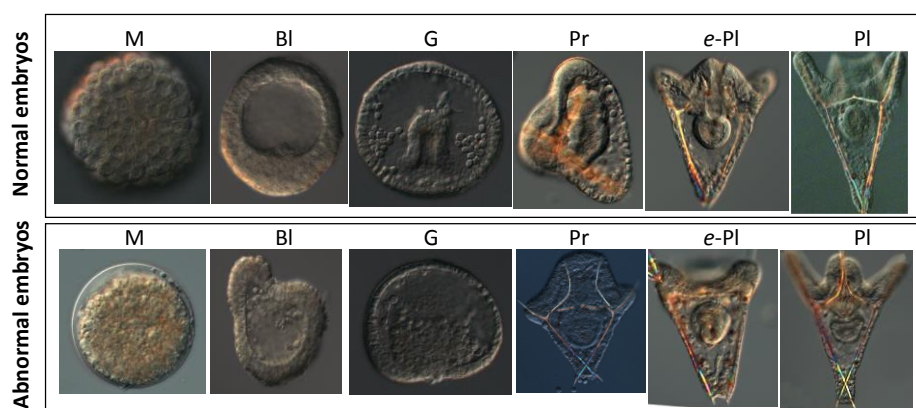
116 The toxicity of elutriate samples from Taranto harbour was estimated by calculating the percentage
117 of abnormal embryos at pluteus stage (48 h of development), according to the standard criteria, the
118 Integrative Toxicity Index (ITI) published by Morroni et al (2016) (here called ITI 1.0), and three
119 additional new ITIs developed from ITI 1.0, respectively called ITI 2.0, ITI 3.0 and ITI 4.0. The
120 standard criteria calculate the percentage of normal and abnormal embryos without considering
121 different typologies of malformations or the phase in which they appear. Conversely, all the ITI
122 methodologies count the frequency of delayed and/or abnormal embryonic morphologies and
123 quantitatively rank the severity of effects.

124 Embryos were classified as normal if they satisfied the morphological criteria as follows: i) suitable
125 schedule in reaching the developmental endpoint (pluteus at 48 h); ii) correct skeleton development
126 and patterning; iii) right ectoderm, mesoderm, and endoderm germ layer differentiation; iv) conform
127 left/right or dorso/ventral axis symmetry. On the other hand, embryos displaying impairment of the
128 axis symmetry, as well as germ layer defects were marked as abnormal (see **Table 1** showing
129 representative images of normal and abnormal *P. lividus* embryos at different embryonic
130 developmental stages).

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Table 1 –Developmental stage and abnormalities of *P. lividus* observed in this study



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M, Morula; Bl, Blastula; G, Gastrula; Pr, Prism; e-Pl, early Pluteus; Pl, Pluteus

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The ITI 1.0 from the previous study used a toxicity scale from 0 (absence of toxicity) to 10 (maximum toxicity), which was here implemented with a second generation of ITIs, using more simplified criteria, grouping embryos on a toxicity scale from 0 (absence of toxicity) to 5 (maximum toxicity) as shown in **Table 2**. The lowest level (0) assigned to each ITIs was associated with a “no effect” in development, including only normal embryos reaching the 48-h endpoint (Pluteus). The only exception was the ITI 4.0 where the zero-effect level was extended also to those embryos displaying a slight delay, considered as a negligible effect (pluteus and early pluteus). A score ranging from 1 to 3 was assigned to the ITI 2.0 and ITI 3.0 as follows: 1 for delayed embryos at the pluteus stage (*e-Pl*); 2 for malformed embryos at the pluteus stage (*m-Pl*); and 3 for delayed embryos at the pluteus stage displaying malformations (*em-Pl*). For ITI 2.0, ~~the~~ level 4 was associated with the delayed embryos from fertilized egg to prism ~~to fertilized egg~~ (*F-Pr-F*) and ~~the~~ level 5 with the malformed F-Pr-F (*F-m-Pr-F*); in ITI 3.0 the levels 4 and 5 were assigned to include delayed and/or malformed Pr (Pr and/or *m-Pr*), and delayed and/or malformed embryos from fertilized egg to gastrula stage ~~to~~

160 ~~fertilized egg~~ (F-G ~~F~~ and/or F-m-G-~~F~~), respectively. On the other hand, the score 1 assigned to the
 161 ITI 4.0 was associated with the Pl and e-Pl displaying malformations (*m*-Pl and *em*-Pl), the score 2
 162 was associated with the stage of the prism (Pr), and the score 3 with the Pr displaying malformations
 163 (*m*-Pr). The highest levels of toxicity (4 and 5) assigned to the ITI 4.0 were associated with the delayed
 164 and/or malformed embryos at the gastrula stage (G and/or *m*-G), and with the delayed and/or
 165 malformed embryos from fertilized egg to blastula stage ~~to fertilized egg~~ (F-Bl-~~F~~ and/or F-*m*-Bl-~~F~~).
 166 Therefore, the lower degree of toxicity was assigned to Pl with the absence of abnormalities, while
 167 the higher degree was attributed to embryos displaying severe delay, and/or delay *plus* abnormalities
 168 simultaneously.

169 The ITIs applied in this study are calculated as follows:

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$$ITI = \sum_{i=1}^n (S_i * F_i) / 100$$

171 Where S_i is the score associated ~~with~~ each abnormal embryonic morphology and F_i is the frequency
 172 observed for that abnormality expressed as a percentage that morphotypes (~~$n_i=10$~~ $n_i=13$ for ITI 1.0 and
 173 $n=6$ for ITI 2.0, ITI 3.0, and ITI 4.0).

175 **Table 2** – Integrative Toxicity Indexes (ITIs) tested in this study

		Toxicity categories													
ITI 1.0	Normal	Delayed						Malformed							
	Pl	e-Pl	Pr	G	mBl	Bl	M	Pl	Pr	G	mBl	Bl	M		
Score	0	2	3	4	4.5	5	5.5	6	7	7.5	8	9	10		
ITI 2.0	Normal	Delayed			Delayed and Malformed										
	Pl	e-Pl	F-Pr- F			Pl	e-Pl	F-Pr- F							
Score	0	1			4						2	3			5
ITI 3.0	Normal	Delayed		Malformed		Delayed and/or Malformed									
	Pl	e-Pl	Pl	e-Pl	Pr	F-G- F									
Score	0	1		2		3		4				5			
ITI 4.0	Normal	Delayed		Malformed		Delayed and/or Malformed									
	Pl/e-Pl	Pr	Pl/e-Pl	Pr	G	F-Bl- F									
Score	0	2		1		3		4				5			

177 F: fertilized egg; M, Morula; mBl, mesenchyme Blastula; Bl, Blastula; G, Gastrula; Pr, Prism; e-Pl, early Pluteus; Pl,
 178 Pluteus. F-Bl-~~F~~, from fertilized egg to Blastula ~~to fertilized egg~~; F-G-~~F~~: from fertilized egg to Gastrula ~~to fertilized egg~~;
 179 F-Pr-~~F~~: from fertilized egg to Prism ~~to fertilized egg~~.

181 **3. Results and Discussion**

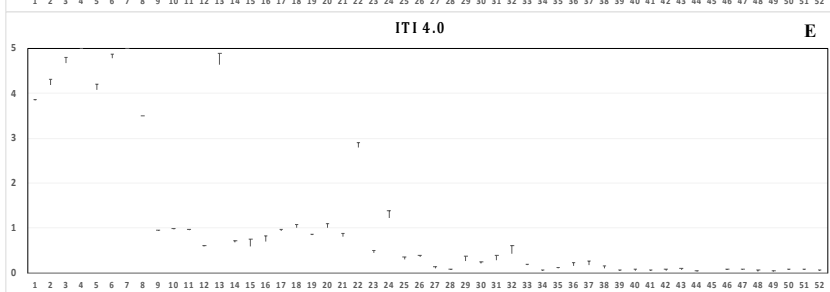
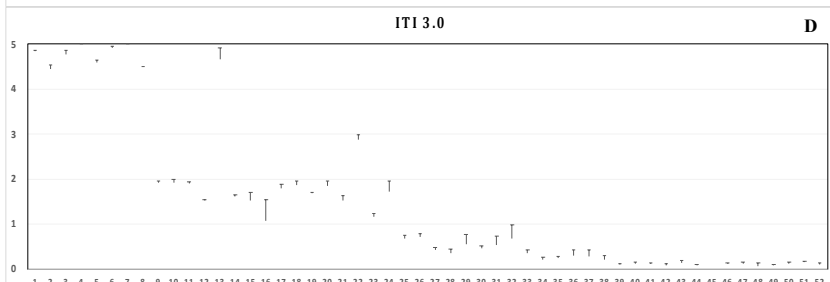
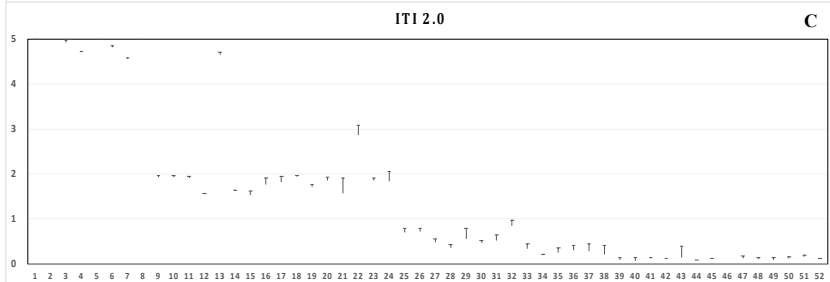
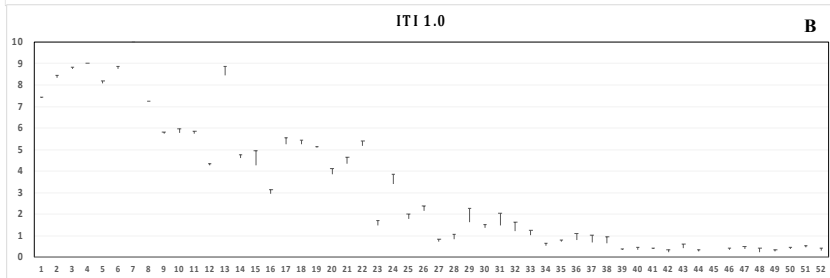
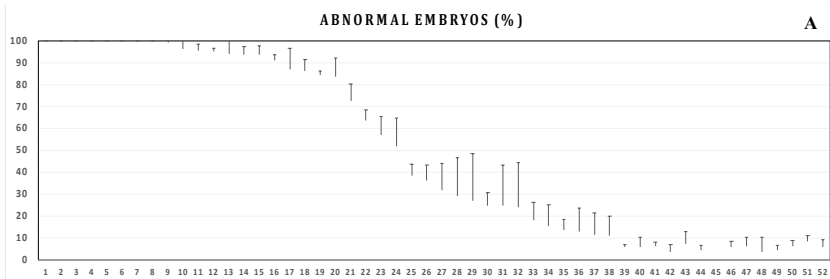
182 The evaluation based on the standard criteria and the thoughtful ITIs (from ITI 1.0 to ITI 4.0) of the
 183 impact on the sea urchin embryonic development is shown in Figure 1. Based on the standard criteria,
 184 the majority of tested elutriates (46.5%) presented ~~an~~ extremely severe or severe toxicity with a
 185 percentage of abnormal embryos higher than 75% (20 of 43 samples; Figure 1A, red bars); a moderate

186 number (28%) displayed from severe to moderate toxicity with a percentage of abnormal embryos
187 ranging from 75% to 20% (12 of 43; Figure 1A, blue bars); the remaining 25.5% did not show any
188 significant impact compared to the controls (11 of 43; figure 1A, compare green bars with those grey)
189 being under the threshold of 20% [\(Morrioni et al., 2019\)](#).

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195 **Figure 1. Sea urchin embryonic development evaluated according to standard criteria and thoughtful**
196 **ITIs.** Histograms represent the results expressed as a mean percentage (%) of abnormal embryos \pm SD (A), and
197 values of the ITI 1.0 (B), ITI 2.0 (C), ITI 3.0 (D), ITI 4.0 (E) \pm SD. Data are referred to each elutriate sample
198 of Taranto harbour, reported on ~~X-axis~~X-axis. Controls (CTR) are reported in the right part of the panels for
199 a total of 9. Red bars: percentage of abnormal embryos higher than 75%; Blue bars: percentage of abnormal
200 embryos ranging from 75% to 20%; Green bars: percentage of abnormal embryos lower than 20%; Grey bars:
201 controls.

202
203 The evaluation of embryo-toxicity based on the ITI 1.0 method, which ~~discriminates~~discriminate the
204 frequency of delayed and/or abnormal embryonic morphologies and quantitatively ~~ranks~~rank the
205 severity of effects on a pondered scale from 0 (absence of toxicity) to 10 (maximum toxicity), leads
206 to an attenuate scenario of morbidity (Figure 1B). The difference appeared more accentuated for those
207 samples classified as extremely toxic by standard criteria (Figure 1B, red bars): only one sample
208 (number 7) confirmed the maximum level of toxicity (10), while eight samples presented a level
209 ranging from 7 to 9 (number 1-6, 8, and 13), six samples ranged from 5 to 6 (moderate toxicity,
210 number 9-11, 17-19), three samples from 4 to 5 (low toxicity, number 12, 14-15), and two samples
211 were below the value 4 (very low toxicity, number 16, and 20). Therefore, the extremely toxic effects
212 assessed by the standard method for 20 sediment samples were confirmed by ITI 1.0 only for 9 of
213 them (from level 7 to level 10), corresponding to about 50% of the cases. In agreement with our recent
214 report, this result provides ~~the~~ evidence that the ITI 1.0 allows to better separate the samples according
215 to the frequency and severity of delayed and/or abnormal morphologies (Morrone et al., 2016). ITI
216 1.0 is more sensitive than traditional toxicological testing strategy but presents the disadvantage that
217 the rigorous morphological analysis may be applied by trained personnel on the sea urchin embryonic
218 development: in this respect, for unspecialized operators, it may result less rapid, simple, and direct
219 than conventional methods based on observation of normal *vs.* abnormal embryos only. To simplify
220 the promising approach of ITI for determining a more realistic toxicological evaluation of dredged
221 sediments, we tested ~~the~~ second generation of indices still based on the frequency of delayed and/or
222 abnormal embryonic morphologies but calculated using a simplified scale, in which embryos are
223 grouped on a toxicity scale from 0 (absence of toxicity) to 5 (maximum toxicity) (see Table 2,
224 Material and Methods). Based on this aim, scores assigned by ITI 2.0 range from 1 to 5 as follows: 1
225 for delayed plutei (*e*-PI); 2 for malformed plutei (*m*-PI); 3 for delayed and malformed plutei (*em*-PI);
226 4 for delayed embryos from fertilized egg to prism ~~to fertilized eggs~~ (F-Pr-F); 5 for malformed
227 embryos from fertilized egg to prism ~~to fertilized eggs~~ (F-m-Pr-F). The evaluation of embryo-toxicity
228 of Taranto elutriates based on the ITI 2.0 is reported in Figure 1C. In agreement with the ITI 1.0
229 results, extremely toxic effects evaluated by standard criteria for the 20 samples were confirmed only
230 for 9 of them (from level 4 to 5); the remaining 11 samples displayed ~~a~~ low toxicity with levels below

231 2. Even if the ITI 2.0 appear less discriminating than ITI 1.0, this index displayed a similar level of
232 performance in samples classified as extremely toxic (compare Figure 1B and 1C).

233 The ITI 3.0 showed a trend comparable to ITI 2.0, with the difference that levels below 2 were much
234 less flattened (Figure 1D). The ITI 3.0 assign the score from 1 to 3 as for ITI 2.0 (1 for *e*-PI; 2 for *m*-
235 PI; 3 for *em*-PI), whereas different criteria have been used for attributing values of 4 and 5: 4 for
236 delayed and/or malformed prisms (*em*-Pr), and 5 for delayed and/or malformed embryos from
237 ~~fertilized egg to gastrula~~ ~~to fertilized eggs~~ (~~F-em-G-F~~). These differences increased the ability of ITI
238 3.0 to discriminate among groups compared to ~~the~~ ITI 2.0.

239 Considering the results of the first six samples, the toxic levels were, on average, 5% lower than those
240 obtained from ITI 2.0. This slight increase in sensitivity was mostly observed in ~~the~~ sample number
241 2 and 5, with 13% and 5% of the embryos at the prism stage, respectively (~~not shown~~). This stage
242 was not well discriminated in ITI 2.0 as ~~the~~ score 4 is assigned to embryos at the stages from ~~fertilized~~
243 ~~egg (F) to prism (Pr)~~ ~~to fertilized egg (F)~~.

244 Finally, when we used the ITI 4.0, discriminating strictly developmental delay and morphological
245 defects from ~~fertilized egg to gastrula~~ ~~to fertilized eggs~~ (1 for *em*-PI; 2 for *e*-Pr; 3 for *m*-Pr; 4 for *em*-
246 G; 5 for ~~F-em-BI-F~~), we still increased the ability to discriminate among groups (see Figure 1E).
247 Notably, several samples showed a lower and more distributed values of toxicity, such as samples
248 number 1, 2, 5, and 8 (ITI 4.0 values ranging from 3.5 to 4.2). Other samples, as the number 7,
249 maintained the maximum level of toxicity (5), confirming the good discriminatory ability of this
250 index, which consider the different degree of severity assigned at early stages, discriminating between
251 gastrula and pre-gastrula stages (from ~~fertilized egg to blastula~~ ~~to fertilized eggs~~) (see Table 2).

252 The sea urchin embryo is a simple model to monitor the developmental stages from fertilization to
253 pluteus stage; *Paracentrotus lividus*, under controlled conditions of temperature (18°C) reaches the
254 pluteus stage after 48 h. The embryonic development requires a prompt and synchronised
255 combination of cell proliferation, fate specification, and movement, controlled by gene regulatory
256 networks (Erkenbrack et al., 2018). Cell fate is specified at the appropriate space and time (blastula-
257 early gastrula stage of development) when cells become able to express a set of differentiated germ
258 layer-exclusive genes (Davidson et al., 1998). Elevated metabolic rates decrease capability for
259 growth, and promote developmental delay of the sea urchin embryos; for example, this happens under
260 acidified seawater conditions (Stumpp et al., 2011).

261 The most documented explanation on sea urchin embryonic delay as an effect of toxicity, is the
262 reduction in the ability to uptake calcium and, in turn, to maintain intracellular homeostasis related to
263 a low extracellular pH; calcium-contaminant trafficking competition also affects the normal gene
264 regulatory network controlling development (Stumpp et al., 2011; Pinsino et al., 2011). ~~Several~~A

265 ~~number of~~ developmental steps such as fertilization, cleavage, neuronal development, skeletogenesis,
266 cell death, and body modelling are known to be dependent on calcium ion trafficking (Webb and
267 Miller 2003). On the other hand, regulatory studies reveal that the embryos present an early sequence
268 of encoded “fail-safe” regulatory devices (Smith and Davidson 2009). Based on this evidence, we
269 speculate that in the early embryonic stages (from ~~fertilized egg to gastrula~~ ~~to fertilized egg~~), when
270 the cell fate has not been specified yet, the probability for embryos to recover from the delay and
271 continue the development is ~~really~~ scant, thereby justifying the assignment of the higher degree of
272 toxicity to embryos displaying severe delay, abnormalities, or delay plus abnormalities occurring
273 simultaneously. On the contrary, at the late embryonic phases (from prism to early pluteus), when the
274 cell fate is already specified, the embryos have a high probability to continue the development, thus
275 explaining the assignment of the lower degree of toxicity. The increasing grading of mild, moderate,
276 and severe effects assigned ~~to~~ the severity of delay and teratogenicity, was progressively
277 emphasized from the ITI 2.0 to the ITI 4.0. All these indexes can be considered valid tools to better
278 evaluate the ~~embryotoxicity~~ ~~embryo toxicity~~ effects on sea urchin based on objective and solid
279 scientific criteria, with clearly important applicative consequences when assessing the quality of
280 dredged marine sediments: among the various indexes, ITI 4.0 which stress mainly the severity of
281 delay, offers the higher sensitivity and discriminatory efficiency.

282 4. Conclusions

283 The use of the WOE integration, which ~~combine~~ ~~combine~~ and weight different typologies of data and
284 analyses, allows to better discriminate the presence of contaminants and their short or long-term
285 consequences, especially when apparently contrasting results are provided by various LOEs. The
286 possibility to convert complex scientific information into simple hazard indexes, easily
287 understandable ~~by~~ ~~from~~ ~~policy makers~~ ~~policy makers~~ and environmental managers, can facilitate and
288 orientate the more appropriate and site-specific decisions on environmental sediment management
289 (Morrone et al., 2020). In this context, the sea urchin embryo-toxicity bioassay is considered an
290 essential component for evaluating the quality of sediments in harbour areas, with important
291 environmental and economic implications. Classifications based on the worst result are still in use
292 and significant consequences may arise depending on the choice of the ecotoxicological assays within
293 a battery. In particular, this study demonstrated that care should be taken in the evaluation of embryo-
294 toxicity results suggesting ~~to~~ weight developmental delay and morphological defects in a balanced
295 way.

296 The development of such a sensitive method is of great utility to properly achieve a reliable harbour-
297 sediment hazard assessment and related risk evaluation. The results obtained in the present study

298 indicate that ITI 4.0 is a promising approach for dredged sediments, better discriminating samples
299 with intermediate toxicity from those highly toxic.

300

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305

306 **CRediT authorship contribution statement**

307 **Rosa Bonaventura: Investigation, Data Curation; Writing-Original Draft, Project administration;**
308 **Francesca Zito: Investigation; Lorenzo Morroni: Software, Formal analysis, Writing-Review &**
309 **Editing; David Pellegrini: Writing-Review & Editing; Francesco Regoli: Writing-Review &**
310 **Editing, Validation, Software; Annalisa Pinsino: Conceptualization, Investigation, Data Curation,**
311 **Writing-Original Draft, Writing-Review & Editing; Project administration, Funding acquisition.**

312 **Declaration of Competing Interest**

313 **The authors declare no competing financial interest.**

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316 **References**

- 317 1) APAT-ICRAM, 2007. Manuale per la movimentazione di sedimenti marini, pp. 1–67. Available
318 at <www.apat.gov.it> and <www.icram.org >
319
- 320 2) ASTM E1563, 1995. Standard guide for conducting static acute toxicity tests with echinoid
321 **embryoembryos**. E 1563-95 In Annual Book of ASTM Standards, Philadelphia, PA, vol. 11(5),
322 pp. 999–1017.
323
- 324 3) Benedetti, M., Gorbi, S., Fattorini, D., D’Errico, G., Piva, F., Pacitti, D., Regoli, F., 2014.
325 Environmental hazards from natural hydrocarbons seepage: Integrated classification of risk from
326 sediment chemistry, bioavailability and biomarkers responses in sentinel species. Environ. Pollut.
327 185, 116–126. <https://doi.org/10.1016/j.envpol.2013.10.023>
328
- 329 4) Bonaventura R, Zito F, Chiamonte M, Costa C, Russo R (2018). Nickel toxicity in *P. lividus*
330 embryos: Dose dependent effects and gene expression analysis. Mar Environ Res. 2018, 139:113-
331 121. <https://doi.org/10.1016/j.marenvres.2018.05.002>
332
- 333 5) Carballeira, C., Ramos-Gómez, J., Martín-Díaz, L., DelValls, T.A., 2012. Identification of
334 specific malformations of sea urchin larvae for toxicity assessment: application to marine
335 pisciculture effluents. Mar Environ Res. 77, 12-22.
336 <https://doi.org/10.1016/j.marenvres.2012.01.001>
337
- 338 6) Davidson, E. H., Cameron, R. A., and Ransick, A. (1998). Specification of cell fate in the sea
339 urchin embryo: summary and some proposed mechanisms. Development. 125, 3269–3290.
340

- 341 7) DM 173/2016. Ministero dell’Ambiente e della Tutela del Territorio e del Mare, Supplemento
342 ordinario alla Gazzetta Ufficiale, n. 208 del 6 settembre 2016-Serie generale. Regolamento
343 recante modalità e criteri tecnici per l’autorizzazione all’immersione in mare dei materiali di
344 escavo di fondali marini.
345
- 346 8) Dorey, N., Martin, S., Oberhänsli, F., Teyssié, J.L., Jeffree, R., Lacoue-Labarthe, T., 2018. Ocean
347 acidification modulates the incorporation of radio-labeled heavy metals in the larvae of the
348 Mediterranean sea urchin *Paracentrotus lividus*. *J Environ Radioact.* 190-191, 20-30.
349 <https://doi.org/10.1016/j.jenvrad.2018.04.017>
350
- 351 9) Environment Canada, 2011. Biological Test Methods: Fertilization Assay using Echinoids (Sea
352 Urchins and Sand Dollars). EPS 1/RM/27.
353
- 354 10) Erkenbrack, E.M., Davidson, E.H., Peter, I.S., 2018. Conserved regulatory state expression
355 controlled by divergent developmental gene regulatory networks in echinoids. *Development*,
356 145(24). doi: 10.1242/dev.167288
357
- 358 11) Khosrovyan, A., Rodríguez-Romero, A., Salamanca, M.J., DelValls, T.A., Riba, I., Serrano, F.,
359 2013. Comparative performances of eggs and embryos of sea urchin (*Paracentrotus lividus*) in
360 toxicity bioassays used for assessment of marine sediment quality. *Mar. Pollut. Bull.* 70(1-2),
361 204-209. <https://doi.org/10.1016/j.marpolbul.2013.03.006>
362
- 363 ~~12)~~ Lister, K.N., Lamare, M.D., Burritt, D.J., 2010. Oxidative damage in response to natural levels of
364 UV-B radiation in larvae of the tropical sea urchin *Tripneustes gratilla*. *Photochem. Photobiol.*
365 86, 1091-1098. <https://doi.org/10.1111/j.1751-1097.2010.00779.x>
366
- 367 ~~13)~~ Morrone, L., Sartori, D., Costantini, M., Genovesi, L., Magliocco, T., Ruocco, N., Buttino, I.
368 2019. First molecular evidence of the toxicogenetic effects of copper on sea urchin *Paracentrotus*
369 *lividus* embryo development. *Water Res.* 160: 415-423. doi: 10.1016/j.watres.2019.05.062.
370
- 371 ~~14)~~ Morrone, L., d’Errico, G., Sacchi, M., Molisso, F., Armiento, G., Chiavarini, S., Rimauro, J.,
372 Guida, M., Siciliano, A., Ceparano, M., Carraturo, F., Tosti, E., Gallo, A., Libralato, G., Patti,
373 F.P., Gorbi, S., Fattorini, D., Nardi, A., Di Carlo, M., Mezzelani, M., Benedetti, M., Pellegrini,
374 D., Musco, L., Danovaro, R., Dell’Anno, A., Regoli, F., 2020. Integrated characterization and risk
375 management of marine sediments: the case study of the industrialized Bagnoli area (Naples, Italy).
376 *Mar. Environ. Res.*, 160, 104984, <https://doi.org/10.1016/j.marenvres.2020.104984>
377
- 378 ~~15)~~ Morrone, L., Pinsino, A., Pellegrini, D., Regoli, F., Matranga, V., 2016. Development of a
379 new integrative toxicity index based on an improvement of the sea urchin embryo toxicity test.
380 *Ecotoxicol. Environ. Saf.* 123, 2–7. <https://doi.org/10.1016/j.ecoenv.2015.09.026>
381
- 382 ~~16)~~ Morrone, L., Pinsino, A., Pellegrini, D., Regoli, F., 2018. Reversibility of metal induced
383 malformations in sea urchin embryos. *Ecotox. Environ. Saf.* 148, 923-929.
384 <https://doi.org/10.1016/j.ecoenv.2017.11.013>
385
- 386 ~~17)~~ Oliviero, M., Tato, T., Schiavo, S., Fernández, V., Manzo, S., Beiras, R., 2019. Leachates of
387 micronized plastic toys provoke embryotoxic effects upon sea urchin *Paracentrotus lividus*.
388 *Environ. Pollut.*, 247, 706–715. <https://doi.org/10.1016/j.envpol.2019.01.098>
389
390

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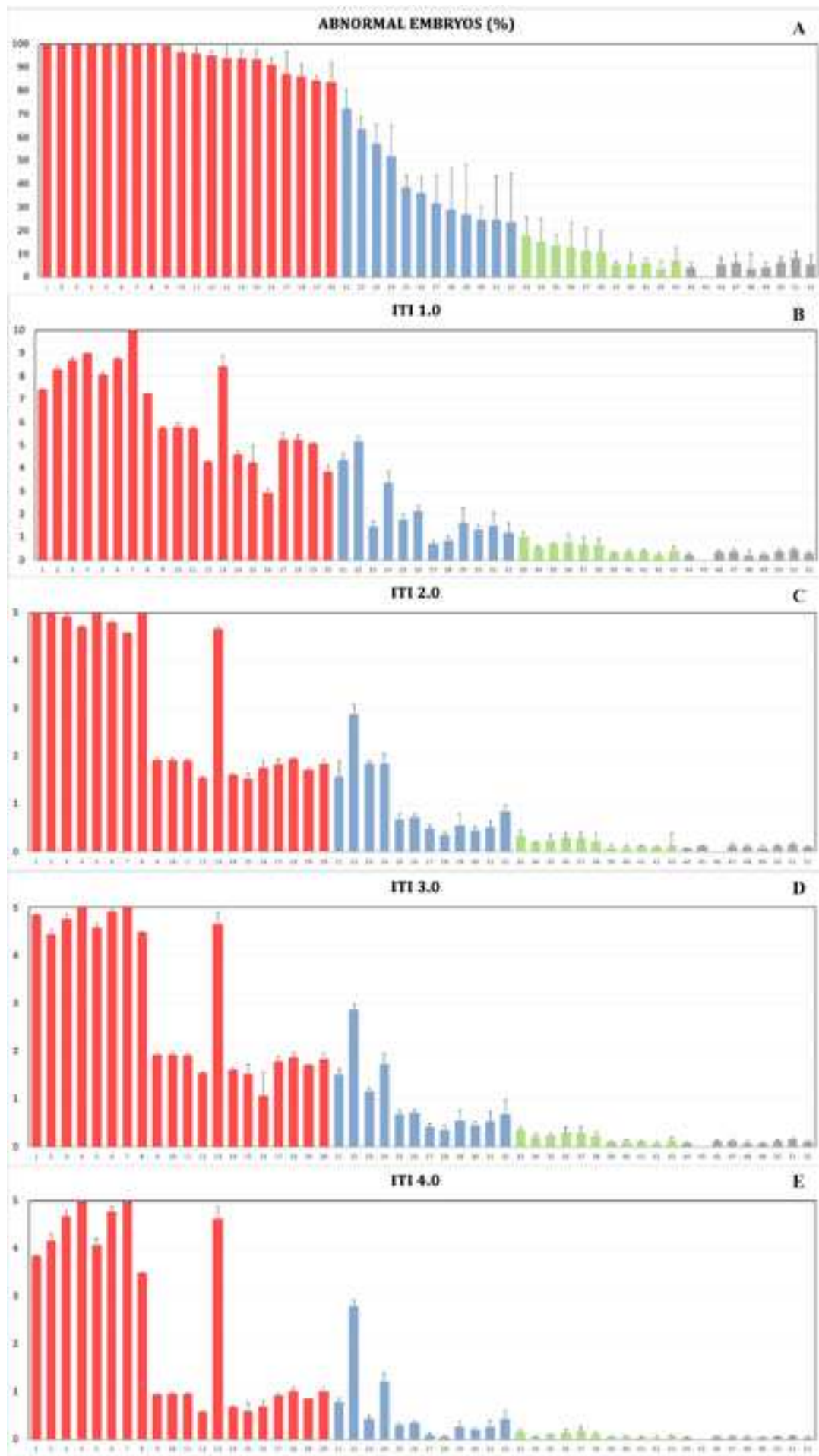
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391 ~~46~~18) Pagano, G., Thomas, P.J., Guida, M., Palumbo, A., Romano, G., Trifuoggi, M. Oral, R.,
392 Trifuoggi, M., 2017. Sea Urchin Bioassays in Toxicity Testing: II. Sediment Evaluation. Expert
393 Opin. Environ. Biol. 6, 1. doi: 10.4172/2325-9655.1000141
394
395 ~~47~~19) Pinsino, A., Roccheri, M.C., Costa, C., Matranga, V., 2011. Manganese interferes with
396 calcium, perturbs ERK signaling, and produces embryos with no skeleton. Toxicol Sci., 123, 217-
397 30. <https://doi.org/10.1093/toxsci/kfr152>
398
399 ~~48~~20) Pinsino, A., Bergami, E., Della Torre, C., Vannuccini, M.L., Addis, P., Secci, M., Dawson,
400 K.A., Matranga, V., Corsi, I., 2017. Amino-modified polystyrene nanoparticles affect signaling
401 pathways of the sea urchin (*Paracentrotus lividus*) embryos. Nanotoxicology, 11, 201-209. DOI:
402 10.1080/17435390.2017.1279360
403
404 ~~49~~21) Passarelli, M.C., Cesar, A., Riba, I., DelValls, T.A. 2017. Comparative evaluation of sea-
405 urchin larval stage sensitivity to ocean acidification. Chemosphere 184, 224–234. DOI:
406 10.1016/j.chemosphere.2017.06.001
407
408 ~~20~~22) Picone, M., Bergamin, M., Losso, C., Delaney, E., Arizzi Novelli, A., Ghirardini, A.V., 2016.
409 Assessment of sediment toxicity in the Lagoon of Venice (Italy) using a multi-species set of
410 bioassays. Ecotoxicol. Environ. Saf. 123, 32-44, <https://doi.org/10.1016/j.ecoenv.2015.09.002>
411
412 ~~24~~23) Pirastu, R., Comba, P., Iavarone, I., Zona, A., Conti, S., Minelli, G., Manno, V., Mincuzzi,
413 A., Minerba, S., Forastiere, F., Mataloni, F., Biggeri A., 2013. Environment and Health in
414 Contaminated Sites: The Case of Taranto, Italy. Int J Environ Res Public Health 3, 753719.
415 <https://doi.org/10.1155/2013/753719>
416
417 ~~22~~24) Pittura, L., Avio, C.G., Giuliani, M.E., D’Errico, G., Keiter, S.H., Cormier, B., Gorbi, S.,
418 Regol, F., 2018. Microplastics as vehicles of environmental PAHs to marine organisms:
419 Combined chemical and physical hazards to the mediterranean mussels, *Mytilus*
420 *galloprovincialis*. Front. Mar. Sci. 5(103), <https://doi.org/10.3389/fmars.2018.00103>
421
422 ~~23~~25) Piva, F., Ciaprini, F., Onorati, F., Benedetti, M., Fattorini, D., Ausili, A., Regoli, F., 2011.
423 Assessing sediment hazard through a weight of evidence approach with bioindicator organisms:
424 A practical model to elaborate data from sediment chemistry, bioavailability, biomarkers and
425 ecotoxicological bioassays. Chemosphere 83, 475–485,
426 <https://doi.org/10.1016/j.chemosphere.2010.12.064>
427
428 ~~24~~26) Regoli, F., D’Errico, G., Nardi, A., Mezzelani, M., Fattorini, D., Benedetti, M., Di Carlo, M.,
429 Pellegrini, D., Gorbi, S., 2019. Application of a Weight of Evidence Approach for Monitoring
430 Complex Environmental Scenarios: the Case-Study of Off-Shore Platforms. Front. Mar. Sci. 6,
431 1–15. <https://doi.org/10.3389/fmars.2019.00377>
432
433 ~~25~~27) Russo, R., Bonaventura, R., Matranga, V., 2014. Time- and dose-dependent gene expression
434 in sea urchin embryos exposed to UVB. Mar. Environ. Res. 93, 85–92.
435 <https://doi.org/10.1016/j.marenvres.2013.08.006>
436
437 ~~26~~28) Smith, J., Davidson, E.H., 2009. Regulative recovery in the sea urchin embryo and the
438 stabilizing role of fail-safe gene network wiring. Proc Natl Acad Sci U S A., 106, 18291-18296.
439 <https://doi.org/10.1073/pnas.0910007106>
440

441 ~~27~~²⁹ Stumpp, M., Wren, J., Melzner, F., Thorndyke, M.C., Dupont, S.T., 2011. CO₂ induced
442 seawater acidification impacts sea urchin larval development I: elevated metabolic rates decrease
443 scope for growth and induce developmental delay. *Comp Biochem Physiol A Mol Integr Physiol.*,
444 160, 331-340. <https://doi.org/10.1016/j.cbpa.2011.06.022>
445
446 ~~28~~³⁰ USEPA, 1991. Evaluation of dredged material proposed for ocean disposal testing manual.
447 EPA 503/8-91/001.
448
449 ~~29~~³¹ USEPA, 1995. Short-term methods for estimating the chronic toxicity of effluents and
450 receiving waters to west coast marine and estuarine organisms. EPA/600/R-95/136.
451
452 ~~30~~³² Webb, S.E., Miller, A.L., 2003. Calcium signalling during embryonic development. *Nat Rev*
453 *Mol Cell Biol.*, 4, 539-551. DOI: 10.1038/nrm1149
454
455



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Declaration of interests

X The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: