

Article

Demineralization Pretreatments for Reducing Biomass Variability in Pyrolysis

Carmen Branca* and Colomba Di Blasi

Cite This: https://doi.org/10.1021/acsomega.3c09321

ACCESS

III Metrics & More

Article Recommendations

Read Online

ABSTRACT: Thermogravimetric and calorimetric analyses are applied to study how washing modifies the pyrolysis rates and heats of five samples of potato plant stems. Hot (553 K) water or dilute (hydrochloric) acid washing of powdered samples causes a reduction in the alkali content by about 62–78 or 97–99%. The feedstock variability is highly reduced, especially for dilute acid treatment. The char yields drastically decrease up to 42–50%, with increases in the peak rates and corresponding temperatures of up to 20–60% and 50–60 K, respectively. Overall, these characteristic parameters closely approach the beech wood values used for



comparison. The shape of the rate curves also testifies the dissolution of nonstructural organic components (pectin, starch, and protein) essentially to the advantage of holocellulose. The ratios between activation energy and order of the global devolatilization reaction increase from about 62–98 kJ/mol (no treatment) to 78–104 kJ/mol (hot water) and 113–124 kJ/mol (dilute acid) (versus 141 kJ/mol for wood). Following washing, the strong exothermic character of the crop residues (global reaction heats from -560 to -180 J/g) is lost. The pyrolysis becomes nearly thermally neutral after hot water washing (heats from -106 to -25 J/g). Furthermore, dilute acid washing makes the process shift from exothermic to endothermic with heats around 70–270 J/g (versus 238 J/g of wood).

INTRODUCTION

The variability of biomass is a serious barrier for the scale-up and commercialization of thermochemical conversion technologies,¹ in particular pyrolysis. A rough biomass classification is sometimes introduced² as terrestrial or aquatic, but variations are huge even in the same class. For instance, the sector of origin of terrestrial biomass, such as energy crops or agricultural, forest, and industry waste, gives rise to largely different properties.³ The main structural components, hemicellulose, cellulose, and lignin, present contents and chemical characteristics varying from one biomass to another.⁴⁻⁶ The composition is further modified by nonstructural organic and inorganic phases. The organic part is often indicated under the item "extractives." These consist of⁴ "various saccharides and carbohydrates, proteins, hydrocarbons, oils, aromatics, lipids, fats, starches, phenols, waxes, chlorophyll, resins, terpenes, terpenoids, acetyls, uronic acids, organic acids, sterols, glycosides, alkaloids, gums, mucilages, dyes, saponins, tannins, and flavonoids." The content and nature are affected by the specific material and can reach rather high values, especially for agro-industrial residues, forestry wastes, and energy crops.⁷ High variability is also observed for the inorganic phase, depending on genetic and environmental issues and physiological and morphological differences among feedstock. Components primarily consist of potassium, calcium, sodium, magnesium, silicon, phosphorus, sulfur, chlorine, and generally, in minor amounts, others such as noble and heavy metals.

Although under practical conditions, physical processes control the pyrolytic conversion,^{10–15} the specific properties of each feedstock always play key roles. For instance, the packedbed pyrolysis of a significant number of lignocellulosic biomasses^{16,17} clearly shows that, while the conversion times are essentially determined by the bulk density, the yields and quality of the products essentially depend on the chemical composition. This is understandable given the large diversity in the macrocomponent pyrolysis products¹⁸ and the catalysis of the inherent metals, in particular the alkali and alkaline earth metals (AAEMs), acting on the selectivity of primary and secondary reactions.^{19–23}

Significant variations are also observed for the same type of biomass, $^{24-30}$ depending on the genotype/cultivar, geographical origin, plant part, harvest year, and so on. Potato crop residue has been recently found $^{31-34}$ to show very large variability in pyrolysis. Indeed, largely different composition, degradation characteristics, and exothermicity magnitude are observed at both the micro- and macroscale. For instance, for

Received:	November 22, 2023
Revised:	January 23, 2024
Accepted:	January 25, 2024

In the second se

five samples originated from approximately the same geographical area, the factors of variations in the volatile products generated by the most important pseudocomponent (cellulose-starch), the global (exothermic) reaction heat, and, for packed beds, the maximum temperature overshoot are approximately in the range of $1.5-2.5^{32,33}$ (the corresponding range for the residue parts³⁴ is about 1.3-1.4, excluding foliage). It is understandable that the large impact of variability on the pyrolysis characteristics makes difficult the optimization of the conversion systems for this kind of residue largely available worldwide.³⁵ Hence, it is useful to look for possible feedstock pretreatments to make the chemicophysical properties more homogeneous and possibly more similar to those of wood. In fact, in addition to very high AAEM contents, for this waste, considerable amounts of pectin, starch, and protein are reported.³² Washing pretreatments, using water or dilute acid, have been demonstrated^{36,37} to remove, at different extent, the soluble parts of both organic and inorganic phases. However, their impacts on biomass variability have not yet been studied. The aspects related to changes in the reaction thermicity also deserve careful consideration as the endo- or exothermic character of lignocellulosic pyrolysis is an extremely important topic in pyrolysis but still largely unknown.^{38,39}

In this study, the role of washing (hot water and dilute acid) is investigated on the pyrolysis and related energetic aspects for the potato crop residues, aimed at ascertaining whether feedstock variability can be reduced and the pyrolytic behavior can approach that typical of woody materials. The analysis is conducted at the microscale using thermogravimetric and calorimetric analyses. The main thermogravimetric parameters, the ratio between the activation energy and the order of the global devolatilization reaction, and the global pyrolysis heats are evaluated and compared with those of beech wood, which is used as a reference material.

MATERIALS AND METHODS

The materials investigated consist of potato plant (Solanum tuberosum L.) stems collected in the small geographical area of Irpinia (Sud Italy) at the end of their life cycle. Five samples (N.1-5) of different cultivars and harvest years are examined, with the same characteristics already presented in previous works^{32,33} (the sample N.5 is from a different batch, but the properties are roughly the same as the previous one). The variability, caused by cultivar and harvest year, was successfully depicted in terms of variable plant aging at the harvest time.^{32–34} So, to facilitate the comparison, the data of Table 1 and those presented and discussed in the following are listed from the least to the most aged item, that is, for the samples N.3, 4, 5, 1, and 2.

As in previous studies, $^{25,28,40-45}$ washing of powdered material (sizes in the range of 50–100 μ m) is carried out using either hot water (hw) at a temperature of 353 K or dilute acid (also indicated in the following as acidic water, aw) of 0.1 mol/L HC1 at ambient conditions. The feedstock to solvent ratio is taken equal to 0.01 g/mL (initial sample mass around 0.5–1 g), with treatment times of 2 h (hw) or 4 h (aw). The suspensions are filtered by means of a 40 μ m metallic wire net, and for aw washing, the collected material is washed with distilled water until neutrality, followed by drying at 353 K.

The untreated and washed residues are characterized in terms of proximate analysis^{28,32-34} and inductively coupled plasma mass spectrometry (ICP-MS) (Agilent 7500ce) analysis of the alkali metals. The decomposition characteristics

Table 1. (A, B) Proximate Analysis and AAEM Contents for the Untreated and hw- and aw-Washed Samples $N.1-S^a$

		-	-		
(A)) sample	VM [wt	%] FC	C [wt %]	ASH [wt %]
N.3		77.0		11.3	11.7
	hw	86.3		10.1	3.6
	aw	86.4		12.1	1.5
N.5		79.0		11.4	9.6
	hw	83.0		14.0	3.0
	aw	86.8		13.0	0.2
N.4		77.0		13.3	9.6
	hw	85.0		10.0	5.0
	aw	87.0		11.3	1.7
N.1		77.1		11.2	11.7
	hw	86.3		10.8	2.9
	aw	88.0		11.0	1.0
N.2		76.9		12.6	10.5
	hw	86.4		9.3	4.3
	aw	86.5		11.9	1.6
beech	wood	86.5		13.1	0.4
(B)	sample	K [ppm]	Ca [ppm]	Mg [ppm]	Na [ppm]
N.3		75 030	12 130	1613	175
hw		12 717	12 120	181	73
aw		591	2460	47	12
N.5		73 800	5768	1860	236
hw		1354	5541	1091	157
aw		191	70	53	67
N.4		50 320	12 200	324	242
hw		10 939	12 577	278	98
aw		109	880	12	53
N.1		82 930	10 770	3864	323
hw		16 134	8685	2208	98
aw		383	1210	8	25
N.2		58 420	15 590	3999	339
hw		5617	12 573	1143	113
aw		121	1288	73	59
beec	h wood	796.4	584	221.6	76.8
^a Beech wood data is included for comparison.					

are evaluated using integral (TG, mass fraction, Y) and differential (DTG, time derivative of the mass fraction, -dY/dt) thermogravimetric analyses (Mettler TGA 1) with a pulverized (sizes below 100 μ m) sample mass of 5 mg, heated at 5 K/min up to 773 K, under a nitrogen flow of 50 mL/min, including a predrying stage at 383 K for 30 min. The characteristic temperatures, rates, and mass fractions^{25,28,46} are evaluated and compared for the various cases.

Accurate kinetic characterization requires the introduction of multistep reaction schemes, to describe macrocomponent dynamics, and parameter fitting methods, as done³² for the untreated feedstocks under exam. However, for engineering evaluations, analytical approaches are available, assuming a single n-order reaction based on mathematical relationships between thermogravimetric and kinetic parameters. More specifically, the mathematical treatment proposed in refs 47,48 to evaluate the ratio between activation energy and reaction order, E/n, is used here for both untreated and washed feedstocks. To apply this analysis, the integral and differential weight loss curves are expressed in terms of conversion, defined as $\alpha = (1 - Y)/(1 - Y_{773})$, with the corresponding $d\alpha/dT$, where Y_{773} is the char yield at complete conversion, assumed to coincide with the mass fraction measured at a temperature of 773 K.^{25,28} The mathematical

expression of the E/n ratio is the same as in refs 47,48 (see eq 4 of ref 48) and uses the thermogravimetric parameters related to the peak rate.

Differential scanning calorimetry (DSC) curves (calorimeter Mettler DSC 1/700) are measured together with the corresponding weight loss data, again using pulverized samples as above, for the untreated and washed feedstocks. A sample mass of 5 mg is heated at 20 K/min up to 773 K, under a nitrogen flow of 160 mL/min, including a predrying stage at 383 K for 20 min, in aluminum open crucibles. The heat reaction curves are computed following the same method previously used, $^{49-54}$ including a radiation correction for open crucibles. To proceed with the computation, the equations are required for the specific heats of wood⁴⁹ and corresponding char⁵⁵ and of crop residues and corresponding chars.⁵² Then, the global reaction heat, *H*, is computed by means of numerical integration of the heat flow curves. The thermogravimetric and calorimetric curves are measured in triplicate, showing good repeatability, as already observed for the untreated samples³² and other feedstocks.^{28,56,57}

Beech wood, a sort of standard woody biomass, is used for comparison of both the thermogravimetric and calorimetric behavior of the untreated and washed food crop residues. Wood pretreatments are not made. In fact, as already stated above, the scope of the investigation is to ascertain whether washing acts to reduce the variability of the residues and to bring their behavior near that of wood.

RESULTS AND DISCUSSION

The effects of the washing pretreatments are first examined following the application of both hot and acidic water. Then, the changes caused on the thermogravimetric parameters, the ratio E/n (activation energy/reaction order) of the global devolatilization reaction, and the global reaction heats are examined.

Proximate Analysis and Alkali Content. The volatile matter (VM), fixed carbon (FC), and inorganic (ASH) contents and the alkali metal (K, Ca, Mg, and Na) contents are summarized in Table 1A,B (data for beech wood is included for comparison^{33,44}). The proximate analysis results show values around 77-79 (VM), 11-13 (FC), and 10-12 (ASH) wt % for the untreated samples N.1-5. The corresponding contents of alkali metals are in the range of 6.3-9.8 wt %. Both washing pretreatments act to reduce or eliminate the soluble part of the inorganic content^{36,37} as well as the possible soil contamination of the samples. Consequently, in the first place, the alkali catalysis on the decomposition reactions, generally favoring charring with respect to devolatilization, is reduced or eliminated, and second, the amount of active material is increased. Together with demineralization, nonstructural soluble organic compounds are also removed to a certain extent.^{36,37} Hence, the active material, constituting the samples, changes in relation to both contents and chemical properties.

The results listed in Table 1B show that the content of alkali metals is highly reduced or practically eliminated after the pretreatments (reductions in the total AAEMs around 62-78 and 97-99 wt % for the hw or the aw washing, respectively). Reductions in the total ash (Table 1A) range from 48 to 75 or 82 to 98 wt % again for the two treatments in the same order. For the hw treatment, the reduction factors, for the most abundant metal, K, vary between about 5 and 10 (factors below 3.5 for the much less abundant Mg and Na), whereas

the Ca content is left practically unvaried. It is plausible that, in some cases (samples N.2 and N.4), the ash consists of a considerable part of components that are not soluble in water; for instance, silicon that anyway does not affect the pyrolysis reactions.³⁷

Following washing, the VM content always increases with values around 83-86 or 86-88 wt % for the hw and the aw washing, respectively. To understand the trend shown by the FC contents, it should be kept in mind that the conditions of proximate analysis^{28,32,33} do not require a kinetic control and that alkali metals play a complex role in conversion. In fact, they catalyze charring reactions so that char formation is favored. However, they also enhance the pyrolysis exothermicity,^{20,21} leading to higher actual reaction temperatures, which are detrimental to char formation. Therefore, the practically complete alkali removal for the aw treatment, by lowering the amount of heat released, generally results in higher FC values. For the hw treatment, the FC contents generally are lower, owing to still non-negligible alkali contents, associated with approximately invariant temperature enhancement. In fact, the exothermicity magnitude during pyrolysis of potassium-loaded wood remains at its maximum for additive contents above 2 wt %.^{20,21} Washing makes closer to the proximate analysis and alkali contents of the agricultural residues and wood, especially for the aw treatment.

TG-DTG Curves of the Demineralized Samples. The integral (TG) and differential (DTG) thermogravimetric curves are compared in Figure 1 for the untreated and the



Figure 1. Thermogravimetric curves of the untreated, the hw-washed, and the aw-washed samples N.1 versus the heating temperature (heating rate 5 K/min). Beech wood curves are included for comparison.

hw- and aw-washed samples N.1, where the beech wood curves are also reported. Figures 2-3 provide a general overview of the effects of the hw and aw treatments on the thermogravimetric behavior of samples N.1–5. The TG-DTG curves of lignocellulosic biomass decomposition, in particular woody materials, are characterized by three main zones: a shoulder, a peak rate, and a tail, mainly associated with hemicellulose,



Figure 2. Thermogravimetric curves of the untreated and the hwwashed samples N.1-5 versus the heating temperature (heating rate 5 K/min). Beech wood curves are included for comparison.



Figure 3. Thermogravimetric curves of the untreated and the awwashed samples N.1-5 versus the heating temperature (heating rate 5 K/min). Beech wood curves are included for comparison.

cellulose, and lignin dynamics. In general, there is an overlap among the macrocomponent rates so that the introduction of pseudocomponents is required for kinetic modeling.^{11,12} The weight loss curves of potato crop residues show more complex dynamics, owing to the presence of large quantities of extractives, pectin, starch, and protein, and the significant AAEM amounts that enhance the overlap.³² The different zones of the weight loss curves can be associated with the dominant component, considering the temperature range where the corresponding model compounds decompose. Excluding the pseudocomponent "light extractives," owing to the low amounts of volatiles they release, in the absence of pretreatment, three main pseudocomponents or zones of the weight loss curves are introduced³² and used in the following. They are pectin-hemicellulose (first or low-temperature zone), cellulose-starch (second or intermediate-temperature zone), and lignin-protein (third or high-temperature zone).

Washing effects are qualitatively the same for all the samples N.1-5 and similar to those observed for other materials.^{25,28,40-45,56} The most evident change consists of an increase in the peak rates and the corresponding temperatures with an improved separation between the reaction zones. The aw washing effects are quantitatively higher. Furthermore, there are important changes in the shape of the rate curves. To partly justify the new trends, it is useful to remind that the leachates from washing treatments decompose over the same temperature range as the origin material, as observed for several biomass materials.^{25,58-60} The low-temperature rates become slower, and the peak generally disappears. This is due to the loss of pectin and other thermally labile components. As already observed, for the second zone, the higher absolute peak rate moves at higher temperatures, testifying to an increase in the cellulose content and its crystallinity degree. The increased crystallinity can be attributed to the reduction or removal of alkalis⁶¹ and starch, which is an amorphous substance. The high-temperature peak rate (third zone) is not clearly visible any longer, not only for the partial dissolution of phenolic compounds and proteins. In fact, for the hw washing, it is most likely hidden by the evolution of the second-zone components occurring at higher temperatures. For the aw washing, it is barely visible under the swollen part of the decay zone after the absolute peak rate. From the physical point of view, this feature can be explained considering that the catalysis of alkali metals on lignin decomposition is smaller than that exerted on the cellulose component.²³ Thus, only if they are almost completely removed, such as for the aw treatment, the lignin decomposition significantly moves toward higher temperatures. The peculiar shape of the rate curve might also be due to the specific properties of the lignin whose chemical (and degradation) characteristics are remarkably dependent upon the origin materials (e.g., woods and straws^{62,63}). Overall, following the washing pretreatments, the rate curves tend to qualitatively approach those of woody biomass (a shoulder, a peak, and a tailing zone).

The influences of the pretreatment on the thermogravimetric characteristics can be quantified from the parameters^{25,28,46} listed in Table 2 (definitions in Figures SM1 and SM2 of the Supplementary Material (SM)). They include the peak rate, $-dY_{peak}/dt$, and the peak/shoulder rate of the first or third zone, $-dY_{ps1}/dt$ or dY_{ps2}/dt , with the corresponding temperatures T_{peak} and T_{ps1} or T_{ps2} and mass fractions, Y_{peak} , and Y_{ps1} or Y_{ps2} . It should be specified that the $-dY_{\text{ps2}}/dt$ rate can be defined only for the untreated and the aw-treated samples, where a third reaction zone (or at least its final part) can be identified. In the former case, a local peak rate is observed, whereas in the latter, the characteristic point corresponds to the first minimum or the shoulder of the $d^{2}Y/dt^{2}$ (the shoulder is defined by a nearly zero value of the third time derivative of the mass fraction). For this treatment, there is also an additional characteristic point, at higher temperatures, again identified by the same features as above, corresponding to the conclusion of the third reaction zone (this is reported in parentheses in Table 2).

The temperature range fwhm (the full width of the rate curve at the half-maximum) and the final charred residue

sample	$-dY_{\rm peak}/dt \times 10^3 \ [s^{-1}]$	$-dY_{ps1}/dt \times 10^3 [s^{-1}]$	$-dY_{\rm ps2}/dt \times 10^3$ [s	$[-1]$ Y_{peak}	$Y_{\rm ps1}$	$Y_{\rm ps2}$	Y ₇₇₃
N.3	0.54	0.32	0.21	0.66	0.86	0.55	0.39
hw	0.73	0.30		0.59	0.85		0.28
aw	0.88	0.33	0.70	0.46	0.91	0.44	0.20
N.5	0.62	0.26	0.27	0.68	0.91	0.53	0.39
hw	0.87	0.22		0.56	0.92		0.27
aw	0.95	0.26	0.74	0.48	0.83	0.34	0.18
N.4	0.65	0.38	0.22	0.63	0.90	0.48	0.36
hw	0.78	0.26		0.58	0.89		0.27
aw	0.94	0.27	0.80	0.51	0.84	0.42	0.19
N.1	0.71	0.28	0.22	0.64	0.91	0.51	0.38
hw	0.88	0.29		0.56	0.88		0.27
aw	1.00	0.32	0.85	0.47	0.83	0.37	0.17
N.2	0.71	0.27	0.15	0.63	0.87	0.50	0.38
hw	0.79	0.26		0.53	0.88		0.25
aw	0.82	0.30	0.64	0.49	0.85	0.40	0.22
beech wood	0.92	0.42		0.41	0.80		0.18
sample	T_{peak} [K]	$T_{\rm ps1}$ [K]	$T_{\rm ps2}$ [K]	fwhm [K]	$\Delta T_1 [K]$		ΔT_2 [K]
N.3	555	511	582	70	50		29
hw	577	532		51	50		34
aw	604	519	611 (623)	40	34		31
N.5	558	506	588	48	54		27
hw	584	523		50	53		23
aw	605	534	617 (625)	43	35		30
N.4	557	508	586	53	52		26
hw	579	525		51	50		31
aw	604	538	613 (631)	42	31		39
N.1	570	513	597	41	46		19
hw	580	533		42	45		27
aw	603	544	612 (622)	40	34		31
N.2	576	523	603	34	33		19
hw	593	534		51	48		30
aw	606	536	617 (629)	44	38		34
beech wood	616	558		48	47		22

^{*a*}Peak rate, $-dY_{\text{peak}}/dt$, and peak/shoulder rate of the low- and high-temperature zones, $-dY_{\text{ps1}}/dt$ and $-dY_{\text{ps2}}/dt$, with the corresponding temperatures, T_{peak} and T_{ps1} , and T_{ps2} , and mass fractions, Y_{peak} and Y_{sp1} and Y_{sp2} , temperature range *fwhm*, the final charred residue, Y_{773} , and the temperature intervals ΔT_1 and ΔT_2 (in bracket, for the variable T_{ps2} , the second characteristic temperature of this zone is reported). ^{*b*}Beech wood data is included for comparison.

(mass fraction) at a temperature of 773 K, Y_{773} , are also considered. Finally, as already done for the analysis of lignocellulosic char oxidation curves,⁶⁴ the characteristic temperature ranges determined by the position of the peak rate and the intercepts (on the temperature axis) obtained by extrapolating the tangents at the points of the fwhm, ΔT_1 and ΔT_2 , are also evaluated. This data gives information about the skewness of the rate curves, which influences the kinetic parameters of the global devolatilization reaction.

The main parameters are also shown in Figure 4A,B (T_{psl} , T_{ps2} , T_{peak} , $-dY_{peak}/dt$, and Y_{773}) and the histograms in Figure 5 (ΔT_1 and ΔT_2). Some characteristic differences ($T_{peak} - T_{ps1}$, $T_{ps2} - T_{peak}$) are also represented by the histograms in Figure 6A,B.

As already noticed, quantitative differences are significant between the effects of the two pretreatments, with the acidic treatment being the most important. Also, the pretreatments exert a stronger effect on the less aged samples N.3, 4, and 5, owing to the larger amounts of soluble organic components.³² Following washing, the peak rates increase significantly (up to factors of 20-60%). For this parameter, the differences among the five samples are reduced to about 20% with respect to 30% in the absence of treatment. The mass fraction, in correspondence with the peak rate, becomes smaller as the washed active material preferentially leads to volatile product formation. Indeed, the yields of char decrease to about 42–55%. The temperature $T_{\rm peak}$ also increases, from values in the range of 555–576 K (no pretreatment) to 577–593 and 603–606 K, for the hw or aw treatments. Again, washing reduces the differences among samples, with variations for this parameter decreasing from 21 K (absence of pretreatments) to 16 and 3 K for the treatments in the same order as above.

For the first reaction zone, the temperatures of the peak or shoulder rates, $T_{\rm ps1}$, again increase (from 512–529 K (no treatment) to 525–533 and 545–550 K for the hw and aw washing in the order), progressively reducing the differences among the samples (from 17 to 8 and 5 K). However, the displacement toward higher values of this first zone is less important than that of the second zone discussed above. Indeed, the differences $T_{\rm peak} - T_{\rm ps1}$ vary from about 39–51 K (absence of treatment) to 44–60 and 53–59 K for the hot and acidic water washing. That is, the alkali effects on cellulose decomposition are stronger than those on the hemicellulose. Hence, washing causes a reduction in the overlap between the



Figure 4. (A, B) Effects of the hw and the aw washing on the thermogravimetric parameters T_{ps1} , T_{peak} , T_{ps2} (A) and $-dY_{peak}/dt$, Y_{773} (B) for the samples N.1–5 (the data is plotted from the least to the most aged sample). Beech wood data is included for comparison.



Figure 5. Histograms of the temperature differences ΔT_1 and ΔT_2 for the untreated, the hw-washed, and the aw-washed samples N.1–5 (the data is plotted from the least to the most aged sample). Beech wood data is included for comparison.

two adjacent reaction zones. The analysis of the third reaction zone is more complicated because, as was already observed, washing causes the disappearance of the local peak rate. Moreover, a change in the slope of the rate curve can be identified only for the aw washing. As expected, $T_{\rm ps2}$ becomes

higher and the differences among samples are reduced (values of 611–617 K versus 555–576 K of the untreated sample). Moreover, the differences $T_{\rm ps2} - T_{\rm peak}$ are reduced (7–12 vs 26–30 K), confirming that the effects of alkali removal are more important for cellulose than lignin. Finally, the second characteristic temperature of the third zone varies between 623 and 631 K.

Washing affects the fwhm parameter through changes in the composition of the active material and the influence of alkali metals on the position of the peak rates. As for the latter, the mutual overlap between the decomposition zones of hemicellulose and cellulose is always reduced. That between the cellulose and lignin zones slightly decreases or increases for the hw or the aw washing, respectively. On the other hand, the removal of significant amounts of organic matter, modifying the shape of the rate curves, also modifies the conditions leading to the definition of fwhm. Overall, a tendency of wider fwhm is observed. The untreated sample N.3 owns the widest value in the absence of treatment due to the large contents of pectin making possible the attainment of high rates already at low temperatures (70 K versus 40-50 K for the treated samples). However, it is useful to observe that, following washing, the differences among the samples are again reduced with fwhm variations among samples from 36 K to 15 or 6 K. Despite the swelling of the right side, the skewness of the rate curves is generally from the left side, as indicated by ΔT_1 > ΔT_2 . Washing generally makes ΔT_2 increase (the sample N.2 deviates from this trend). ΔT_1 remains roughly the same or also increases (hw or aw washing; deviations again for the sample N.2).

As already noticed for the proximate analysis, washing makes the residues more similar to wood, especially for the aw washing. For the untreated agricultural residues, the most aged ones (N.1-2) are those more like beech wood, though quantitative differences are anyway large (overall, T_{peak} lower by about 40-60 K). As shown by the histograms in Figure 7, following hot water washing, sample N.2 is still the most similar to wood, while the others also move closer (T_{peak}) differences of 24 K for sample N.2 and 33-40 K for the others). However, the rate curves are more swollen for the decay part. This feature is even more evident for the aw-treated samples, whose peak rate position approaches closer to that of beech wood (T_{peak} differences around 5–10 K). Moreover, the peak rates and the char yields also become comparable. These findings support the speculation that for the aw-treated residues the cellulose contents and properties become approximately the same as for wood and that differences may be attributed to the different nature of lignin and most likely to differences in the hemicellulose as well.

In summary, the thermogravimetric data shows that the feedstock variability is reduced by the washing pretreatment, with changes that are stronger for the less aged samples N.3-5. The behavior of these approaches the same as that of the most aged ones N.1-2. Overall, the washed residues tend to exhibit thermogravimetric curves that become both qualitatively and quantitatively close to that of (untreated) beech wood, a consequence of more similar chemical compositions. The ash removal causes the reaction process to proceed at higher temperatures. The changes in the shape of the washed residues, becoming more like wood, indicate that a large part or the entire amount of nonstructural organic matter is also eliminated. The increase in the peak rate and strong reduction



Figure 6. (A, B) Histograms of the temperature differences $(T_{peak} - T_{ps1})$ (A) and $(T_{ps2} - T_{peak})$ (B) for the untreated, the hw-washed, and the awwashed samples N.1–5 (the data is plotted from the least to the most aged sample). Beech wood data is included for comparison.



Figure 7. Histograms of the differences between the $T_{\rm peak}$ temperature of beech wood and those of the untreated, the hw-washed, and the aw-washed samples N.1–5 (the data is plotted from the least to the most aged sample).

in the char yields support the speculation that the holocellulose content increases.

Analytical Evaluation of the Thermogravimetric **Curves.** The results of the analytical evaluation of the ratio E/n for the global devolatilization reaction for the untreated and washed samples are reported in Table 3 and Figure 8. It is

Table 3. Estimated Ratio Between the Activation Energy, E, and the Reaction Order, n, of the Global Devolatilization Reaction for the Untread, the hw-Washed, and the aw-Washed Samples N.1-5^{*a*}

	E/n [kJ/mol]		
sample	untreated	hw	aw
N.3	61.6	78.2	123.9
N.5	67.0	103.5	120.3
N.4	74.7	85.2	115.1
N.1	88.7	102.1	121.2
N.2	98.2	99.2	112.9
beech wood	141.2		
an 1 11.			

"Beech wood data is included for comparison.

understandable that this evaluation essentially concerns the central zone of the thermogravimetric curves that in the absence of pretreatments represents the dynamics of cellulose–starch decomposition. In this case, the ratio E/n varies from 62 kJ/mol (sample N.3) to 98 kJ/mol (sample N.2); that is, it increases with sample aging. This finding is



Article

Figure 8. Histograms of the ratios of the activation energy to reaction order, E/n, of the global devolatilization reaction (data in Table 3) for the untreated, the hw-washed, and the aw-washed samples N.1 and 5 (the data is plotted from the least to the most aged sample). Beech wood data is included for comparison.

understandable as the cellulose content and crystallinity also increase, as testified by the successively higher peak rate and corresponding temperature.⁵⁷ On the other hand, both features are described by higher activation energies.⁶¹ Values are much smaller than those obtained for beech wood (141 kJ/mol) and for microcrystalline or cotton linter celluloses⁵⁷ (values of 209 and 198 kJ/mol, respectively). Agricultural residue variability gives rise to a variation in the E/n ratio of around 58%.

A significant increase in the E/n ratio is computed for the pretreated samples with value ranges of 78-104 kJ/mol (hw) and 113-124 kJ/mol (aw). The variability among the five samples is reduced to 32 and 10%. The aw treatment exerts a stronger effect on the E/n values, especially for the less aged samples (N.3 and N.5) with an increase between 15 and 101% (versus 1-54% for the hw treatment). Overall, following washing, the variability among the samples is reduced, and the contribution and crystallinity of cellulose are increased. Moreover, the behavior of the washed residues again approaches that of wood.

It is worth noticing that the kinetic parameters E and n are correlated; that is, an increase in E is also associated with an increase in n.⁶⁵ In other words, washing causes an increase in the E parameter significantly stronger than that represented by the ratio E/n. Information about the n values can be gained by looking at the characteristic temperature range ΔT_{2} , considering the empirical correlation between the two parameters previously reported for char oxidation.⁶⁴ It shows that n increases with ΔT_{2} , and that for values of this below 50

G

K, it is comprised between about 0.5 and 1. Considering that, in the absence of treatment, ΔT_2 roughly varies between 19 and 29 K, possible *n* values are around 0.6–0.7. After washing, ΔT_2 is in the range of 23–39 K so that *n* is expected to vary in the range of 0.7–0.9.

Thermograms and Global Pyrolysis Heats. Thermograms and weight loss curves are reported in Figure 9 for



Figure 9. Calorimetric (A) and differential thermogravimetric (B) curves of the untreated, hw-washed, and aw-washed samples N.1 versus the heating temperature (heating rate 20 K/min). Beech wood curves are included for comparison.

sample N.1 in the absence of treatment and for the hw and aw washing (beech wood data is also included). It is evident that both pretreatments act to reduce the global exothermicity of pyrolysis making the residues approach the wood behavior. In this regard, the acidic water treatment is more effective, as expected from the results of the thermogravimetric analysis. In the absence of pretreatment, the thermal decomposition of the sample N.1 is nearly thermally neutral for the first two reaction zones (decomposition of pectin-hemicellulose and cellulose-starch), while high rates of heat release are observed for the third zone (lignin-protein).

Following hw washing, the decomposition for the first zone remains almost neutral, but a moderate endothermicity appears for the second zone. The decomposition for the third zone is still exothermic, but the magnitude of the thermal event is highly reduced. When aw washing is carried out, the thermal behavior of sample N.1 becomes qualitatively and quantitatively similar to that of beech wood. Holocellulose decomposition (first two reaction zones) occurs endothermically, whereas a very moderate exothermicity appears only for the third zone (lignin). In fact, the globally endothermic decomposition of $wood^{49,50,66}$ also shows moderate exothermicity only in correspondence to the tailing zone. Lignin (and proteins) are the main components responsible for char formation,^{11,12,32} which is an exothermic process. The extractives also produce high yields of charred solid residues,^{25,58-60} contributing to heat release. Moreover, primary and secondary charring reactions are also highly favored by alkali metals, thus further enhancing the exothermicity magnitude.^{20,21} Given the small sample mass and the use of open crucibles, it can be speculated that the observed thermal behavior mainly refers to primary reactions. These findings agree with previous DSC data,⁵⁰ where extraction and water washing of an energy crop are reported to cause a progressive shift of the overall thermicity of the decomposition toward endothermicity.

As shown in Figures 10-11, all the samples N.1-5 show the same qualitative trends, without and with washing, although



Figure 10. Calorimetric (A) and differential thermogravimetric (B) curves of the untreated and hw-washed samples N.1-5 versus the heating temperature (heating rate 20 K/min). Beech wood curves are included for comparison.



Figure 11. Calorimetric (A) and differential thermogravimetric (B) curves of the untreated and aw-washed samples N.1 and 5 versus the heating temperature (heating rate 20 K/min). Beech wood curves are included for comparison.

the peak rates of the calorimetric curves are slightly different. A quantitative comparison can be made by considering the global reaction heat (Table 4 and Figure 12). The computation is made over the temperature range of 450-773 K. The (endothermic) heat for beech wood is 238 J/g, a value comparable with those of spruce⁴⁹ and poplar⁵² woods. The exothermic behavior of the untreated samples is described by pyrolysis reaction heats varying from -181 J/g (sample N.3) to -564 J/g (sample N.5). The hot water treatment makes the conversion nearly thermally neutral with global pyrolysis heats varying from about -106 J/g (sample N.2) to -24 J/g (sample

Table 4. Computed Global Reaction Heats for the Untreated, hw-Washed, and aw-Washed Samples $N.1-5^{a}$

	Q [J/g]		
sample	untreated	hw	aw
N.3	-181	-91	+69
N.5	-564	-26	+270
N.4	-248	-24	+141
N.1	-357	-73	+226
N.2	-324	-106	+175
beech wood	+238		

^aBeech wood data is included for comparison.



Figure 12. Histograms of the global reaction heat, Q, for the untreated, hw-washed, and aw-washed samples N.1–5 (the data is plotted from the least to the most aged sample). Beech wood data is included for comparison.

N.4). Acidic water washing shifts from the exothermic to endothermic process with global reaction heats from 69 J/g (sample N.3) to 270 J/g (sample N.5).

These results clearly indicate that the washing pretreatment is apt to modify not only the decomposition characteristics but also the energetic aspects of the process. The removal of both alkali metals and nonstructural components (extractives, pectin, starch, and proteins) reduces the char yields and thus reduces or eliminates the conversion exothermicity. Therefore, similarly to the results already discussed for the decomposition rates, the thermal behavior of the washed samples closely approaches that of wood, especially for the acidic treatment.

CONCLUSIONS

Microscale analysis (thermogravimetry and calorimetric analysis) is carried out to ascertain the role of hot water and acidic washing on the widely different properties of five samples of potato stem waste, using beech wood for comparison. It is found that washing always acts to reduce the variability of the wastes, making their behavior approach that of wood. This finding is due to the removal of nonstructural inorganics (AAEMs) and organics (pectin, starch, and protein) present in large amounts in the untreated residues.

Following the pretreatments, crop residue decomposition tends to occur at higher temperatures and with higher rates and reduced char yields, especially for acidic water pretreatment. The exothermic character of the decomposition reaction also becomes weaker (hot water washing) and then turns into an endothermic character (acidic water). In this way, thermogravimetric curves (devolatilization rates) and calorimetric curves (global reaction heat) become almost coincident with those obtained for beech wood. In conclusion, from a quantitative point of view, acidic water washing is an effective method for making the waste properties more uniform and comparable with those of woody materials.

In addition to quantitative information about the effects induced by washing on the devolatilization characteristics and global reaction heats, an analytical evaluation of the ratios between the activation energy and reaction order has been made. It is shown that they increase, again approaching the wood value, especially for the acidic treatment (E/n around 113–124 kJ/mol versus 141 kJ/mol for wood). Hence, the cellulose contents, and most likely the crystallinity index, increase with washing treatment.

The better effectiveness of acidic water washing, for reducing the variability of wastes and making their decomposition behavior approach that of wood, avoids the solvent preheating required by hot water. There is, however, the disadvantage of generating an acidic leachate byproduct, which should be disposed of or treated for possible reuse. This is an aspect that requires further investigation in relation to the use of alternative organic acidic substances.

ASSOCIATED CONTENT

Supporting Information

The Supporting Information is available free of charge at https://pubs.acs.org/doi/10.1021/acsomega.3c09321.

Definitions of the TGA parameters (PDF)

AUTHOR INFORMATION

Corresponding Author

Carmen Branca – Istituto di Scienze e Tecnologie per l'Energia e la Mobilità Sostenibili (STEMS), C.N.R., 80125 Napoli, Italy; orcid.org/0000-0003-0216-2428; Phone: 39-081-7682232; Email: carmen.branca@stems.cnr.it

Author

Colomba Di Blasi – Dipartimento di Ingegneria Chimica, dei Materiali e della Produzione Industriale, Università degli Studi di Napoli "Federico II", 80125 Napoli, Italy; orcid.org/0000-0001-5499-6251

Complete contact information is available at: https://pubs.acs.org/10.1021/acsomega.3c09321

Notes

The authors declare no competing financial interest.

ACKNOWLEDGMENTS

The authors thank Fernando Stanzione (Istituto di Scienze e Tecnologie per l'Energia e la Mobilità Sostenibili, CNR, Napoli, Italy) for the ICP-MS analysis of the ashes.

REFERENCES

(1) Yan, J.; Oyedeji, O.; H Leal, J. H.; Donohoe, B. S.; Semelsberger, T. A.; Li, C.; Hoover, A. N.; Webb, E.; Bose, E. A.; Zeng, Y.; Williams, C. L.; Schaller, K. D.; Sun, N.; Ray, A. E.; Tanjore, D. Characterizing variability in lignocellulosic biomass: a review. *ACS Sustainable Chem. Eng.* **2020**, *8*, 8059–8085.

(2) Sharifzadeh, M.; Sadeqzadeh, M.; Guo, M.; Borhani, T. N.; Konda, N. V. S. N. M.; Cortada Garcia, M.; Wang, L.; Hallett, J.; Shah, N. The multi-scale challenges of biomass fast pyrolysis and biooil upgrading: Review of the state of art and future research directions. *Prog. Energy Combust. Sci.* **2019**, *71*, 1–80. (3) González Martínez, M.; Floquet, P.; Dupont, C.; da Silva Perez, D.; Meyer, X. Assessing the impact of woody and agricultural biomass variability on its behaviour in torrefaction through Principal Component Analysis. *Biomass Bioenerg*. **2020**, *134*, No. 105474.

(4) Mohan, D.; Pittman, C. U.; Steele, P. H. Pyrolysis of wood/ biomass for bio-oil: a critical review. *Energy Fuels* **2006**, *20*, 848–889.

(5) González Martínez, M.; Dupont, C.; da Silva Perez, D.; Mortha, G.; Thiery, S.; Meyer, X.; Gourdon, C. Understanding the torrefaction of woody and agricultural biomasses through their extracted macromolecular components. Part 1: Experimental thermogravimetric solid mass loss. *Energy* **2020**, *205*, No. 118067.

(6) González Martínez, M.; Dupont, C.; Anca-Couce, A.; da Silva Perez, D.; Boissonnet, G.; Thiery, S.; Meyer, X.; Gourdon, C. Understanding the torrefaction of woody and agricultural biomasses through their extracted macromolecular components. Part 2: torrefaction model. *Energy* **2020**, *210*, No. 118451.

(7) Vassilev, S. V.; Baxter, B.; Andersen, L. K.; Vassileva, C. G. An overview of the chemical composition of biomass. *Fuel* **2010**, *89*, 913–933.

(8) Vassilev, S. V.; Baxter, D.; Andersen, L. K.; Vassileva, C. G.; Morgan, T. J. An overview of the organic and inorganic phase composition of biomass. *Fuel* **2012**, *94*, 1–33.

(9) Wang, S.; Dai, G.; Yang, H.; Luo, Z. Lignocellulosic biomass pyrolysis mechanism: A state-of-the-art review. *Prog. Energy Combust. Sci.* **2017**, *62*, 33–86.

(10) Di Blasi, C. The state of the art of transport models for charring solid degradation. *Polym. Int.* **2000**, *49*, 1133–1146.

(11) Di Blasi, C. Modeling chemical and physical processes of wood and biomass pyrolysis. *Prog. Energy Combust. Sci.* **2008**, *34*, 47–90.

(12) Anca-Couce, A. Reaction mechanisms and multi-scale modelling of lignocellulosic biomass pyrolysis. *Prog. Energy Combust. Sci.* **2016**, *53*, 41–79.

(13) Ciesielski, P. N.; Pecha, M. B.; Lattanzi, A. M.; Bharadwaj, V. S.; Crowley, M. F.; Bu, L.; Vermaas, J. V.; Steirer, K. X.; M F Crowley, M. F. Advances in multiscale modeling of lignocellulosic biomass. *ACS Sustainable Chem. Eng.* **2020**, *8*, 3512–3531.

(14) Vikram, S.; Rosha, P.; Kumar, S. Recent modeling approaches to biomass pyrolysis: a review. *Energy Fuels* **2021**, *35*, 7406–7743.

(15) Attanayake, D. D.; Sewerin, F.; Kulkarni, S.; Dernbecher, A.; Dieguez-Alonso, A.; van Wachem, B. Review of modelling of pyrolysis processes with CFD-DEM. *Flow, Turbulence Combust.* **2023**, *111*, 355–408.

(16) Di Blasi, C.; Branca, C.; Galgano, A. Biomass screening for the production of furfural via thermal decomposition. *Ind. Eng. Chem. Res.* **2010**, *49*, 2658–2671.

(17) Di Blasi, C.; Branca, C.; Lombardi, V.; Ciappa, P.; Di Giacomo, C. Effects of particle size and density on the packed-bed pyrolysis of wood. *Energy Fuels* **2013**, *27*, 6781–6791.

(18) Collard, F.-X.; Blin, J. A review on pyrolysis of biomass constituents: Mechanisms and composition of the products obtained from the conversion of cellulose, hemicelluloses and lignin. *Renewable Sustainable Energy Rev.* **2014**, *38*, 594–608.

(19) Trubetskaya, A.; Surup, G.; Shapiro, A.; Bates, R. B. Modeling the influence of potassium content and heating rate on biomass pyrolysis. *Appl. Energy* **2017**, *194*, 199–211.

(20) Di Blasi, C.; Branca, C.; Galgano, A. Influences of potassium hydroxide on rate and thermicity of wood pyrolysis reactions. *Energy Fuels* **2017**, *31*, 6154–6162.

(21) Di Blasi, C.; Branca, C.; Galgano, A. Role of the potassium chemical state in the global exothermicity of wood pyrolysis. *Ind. Eng. Chem. Res.* **2018**, *57*, 11561–11571.

(22) Nzihou, A.; Stanmore, B.; Lyczko, N.; Pham Minh, D. The catalytic effect of inherent and adsorbed metals on the fast/flash pyrolysis of biomass: A review. *Energy* **2019**, *170*, 326–337.

(23) Leng, E.; Guo, Y.; Chen, J.; Liu, S.; E, J.; Xue, Y. A comprehensive review on lignin pyrolysis: Mechanism, modeling and the effects of inherent metals in biomass. *Fuel* **2022**, *309*, No. 122102.

(24) Di Blasi, C.; Branca, C.; Santoro, A.; Gonzalez Hernandez, E. Pyrolytic behavior and products of some wood varieties. *Combust. Flame* **2001**, *124*, 165–177.

(25) Várhegyi, G.; Gronli, M. G.; Di Blasi, C. Effects of sample origin, extraction and hot water washing on the devolatilization kinetics of chestnut wood. *Ind. Eng. Chem. Res.* **2004**, *43*, 2356–2367.

(26) Said, N.; Bishara, T.; García-Maraver, A.; Zamorano, M. Effect of water washing on the thermal behavior of rice straw. *Waste Manage*. **2013**, *33*, 2250–2256.

(27) Intani, K.; Latif, S.; Kabir, A. K. M. R.; Müller, J. Effect of selfpurging pyrolysis on yield of biochar from maize cobs, husks and leaves. *Bioresour. Technol.* **2016**, *218*, 541–551.

(28) Branca, C.; Di Blasi, C.; Galgano, A. Pyrolytic conversion of wastes from cereal, protein and oil-protein crops. *J. Anal. Appl. Pyrolysis* **2017**, *127*, 426–435.

(29) Di Blasi, C.; Galgano, C.; Branca, C. Exothermic events of nut shell and fruit stone pyrolysis. *ACS Sustainable Chem. Eng.* **2019**, *7*, 9035–9049.

(30) Rego, F.; Soares Dias, A. P.; Casquilho, M.; Rosa, F. C.; Rodrigues, A. Pyrolysis kinetics of short rotation coppice poplar biomass. *Energy* **2020**, *207*, 118191.

(31) Di Blasi, C.; Branca, C.; Galgano, A.; Autiero, G. Analysis of the pyrolytic runaway dynamics during agricultural waste conversion. *Energy Fuels* **2018**, *32*, 9530–9540.

(32) Branca, C.; Di Blasi, C. Modeling the effects of cultivar and harvest on the decomposition kinetics of potato crop residues. *Fuel* **2023**, 339, No. 127419.

(33) Branca, C.; Galgano, A.; Di Blasi, C. Dynamics and products of potato crop residue conversion under a pyrolytic runaway regime - Influences of feedstock variability. *Energy* **2023**, *276*, No. 127507.

(34) Branca, C.; Galgano, A.; Di Blasi, C. Multi-scale analysis of the exothermic behavior of agricultural biomass pyrolysis. *J. Anal. Appl. Pyrolysis* **2023**, *173*, No. 106040.

(35) Soltanieh, A.; Jazini, M.; Karimi, K. Biorefinery for efficient xanthan gum, ethanol, and biogas production from potato crop residues. *Biomass Bioenerg.* **2022**, *158*, No. 106354.

(36) Iraola-Arregui, I.; Van Der Gryp, P.; Görgens, J. F. A review on the demineralisation of pre- and post-pyrolysis biomass and tyre wastes. *Waste Manage*. **2018**, *79*, 667–688.

(37) Kumar, R.; Strezov, V.; Weldekidan, H.; He, J.; Singh, S.; Kan, T.; Dastjerdi, B. Lignocellulose biomass pyrolysis for bio-oil production: A review of biomass pre-treatment methods for production of drop-in fuels. *Renewable Sustainable Energy Rev.* 2020, 123, No. 109763.

(38) Di Blasi, C.; Branca, C.; Sarnataro, F. E.; Gallo, A. Thermal runaway in the pyrolysis of some lignocellulosic biomasses. *Energy Fuels* **2014**, *28*, 2684–2696.

(39) Di Blasi, C.; Branca, C.; Galgano, A. On the experimental evidence of exothermicity in wood and biomass pyrolysis. *Energy Technol.* **2017**, *5*, 19–29.

(40) Blasi, C. D.; Branca, C. The effects of water leaching on the isothermal degradation kinetics of straw. *Ind. Eng. Chem. Res.* 2000, 39, 2169–2174.

(41) Blasi, C. D.; Branca, C.; D'Errico, G. Degradation characteristics of straw and washed straw. *Thermochim. Acta* **2000**, *364*, 133– 142.

(42) Deng, L.; Zhang, T.; Che, D. Effect of water washing on fuel properties, pyrolysis and combustion characteristics, and ash fusibility of biomass. *Fuel Process. Technol.* **2013**, *106*, 712–720.

(43) Ma, Q.; Han, L.; Huang, G. Evaluation of different waterwashing treatments effects on wheat straw combustion properties. *Bioresour. Technol.* **2017**, 245, 1075–1083.

(44) Branca, C.; Di Blasi, C.; Galgano, A. Experimental analysis about the exploitation of industrial hemp (Cannabis Sativa) in pyrolysis. *Fuel Process. Technol.* **2017**, *162*, 20–29.

(45) Chen, D.; Wang, Y.; Liu, Y.; Cen, K.; Cao, X.; Ma, Z.; Li, Y. Comparative study on the pyrolysis behaviors of rice straw under different washing pretreatments of water, acid solution, and aqueous (46) Branca, C.; Di Blasi, C. A unified mechanism of the combustion reactions of lignocellulosic fuels. *Thermochim. Acta* **2013**, *565*, 58–64. (47) Kim, S.; Park, J. K. Characterization of thermal reaction by peak temperature and height of DTG curves. *Thermochim. Acta* **1995**, *264*, 137–156.

(48) Kim, S.; Jang, E.-S.; Shin, D.-H.; Lee, K.-H. Using peak properties of a DTG curve to estimate the kinetic parameters of the pyrolysis reaction: application to high density polyethylene. *Polym. Degrad. Stab.* **2004**, *85*, 799–805.

(49) Rath, J.; Wolfinger, M. G.; Steiner, G.; Krammer, G.; Barontini, F.; Cozzani, V. Heat of wood pyrolysis. *Fuel* **2003**, *82*, 81–91.

(50) Gomez, C.; Velo, E.; Barontini, F.; Cozzani, V. Influence of secondary reactions on the heat of pyrolysis of biomass. *Ind. Eng. Chem. Res.* **2009**, *48*, 10222–10233.

(51) Basile, L.; Tugnoli, A.; Stramigioli, C.; Cozzani, V. Influence of pressure on the heat of biomass pyrolysis. *Fuel* **2014**, *137*, 277–284.

(52) Chen, Q.; Yang, R.; Zhao, B.; Li, Y.; Wang, S.; Wu, H.; Zhuo, Y.; Che, C. Investigation of heat of biomass pyrolysis and secondary reactions by simultaneous thermogravimetry and differential scanning calorimetry. *Fuel* **2014**, *134*, 467–476.

(53) Basile, L.; Tugnoli, A.; Cozzani, V. Influence of macrocomponents on the pyrolysis heat demand of lignocellulosic biomass. *Ind. Eng. Chem. Res.* **2017**, *56*, 6432–6440.

(54) Barontini, F.; Biagini, E.; Tognotti, L. Influence of torrefaction on biomass devolatilization. *ACS Omega* **2021**, *6*, 20264–20278.

(55) Grønli, M. G. A Theoretical and experimental Study of the thermal Degradation of biomass. PhD thesis. The Norvegian University of Science and Technology, 1996.

(56) Branca, C.; Di Blasi, C. Thermal degradation behavior and kinetics of industrial hemp stalks and shives. *Thermochim. Acta* **2021**, 697, No. 178878.

(57) Branca, C.; Di Blasi, C. Kinetic assessment of the thermal decomposition of hemp fiber and the impact of pretreatments. *J. Therm. Anal. Calorim.* **2022**, *147*, 14423–14435.

(58) Di Blasi, C.; Branca, C.; Santoro, A.; Perez Bermudez, R. A. Weight loss dynamics of wood chips under fast radiative heating. *J. Anal. Appl. Pyrolysis* **2001**, *57*, 77–90.

(59) Mészáros, E.; Jakab, E.; Varhegyi, G. TG/MS, Py-GC/MS and THM-GC/MS study of the composition and thermal behavior of extractive components of Robinia pseudoacacia. *J. Anal. Appl. Pyrolysis* **2007**, *79*, 61–70.

(60) Branca, C.; Di Blasi, C.; Galgano, A.; Clemente, M. Analysis of the interactions between moisture evaporation and exothermic pyrolysis of hazelnut shells. *Energy Fuels* **2016**, *30*, 7878–7886.

(61) Chen, H.; Liu, Z.; Chen, X.; Chen, Y.; Dong, Z.; Wang, X.; Yang, H. Comparative pyrolysis behaviors of stalk, wood and shell biomass: correlation of cellulose crystallinity and reaction kinetics. *Bioresour. Technol.* **2020**, *310*, No. 123498.

(62) Fougere, D.; Nanda, S.; Clarke, K.; Kozinsk, J. A.; Li, K. Effect of acidic pretreatment on the chemistry and distribution of lignin in aspen wood and wheat straw substrates. *Biomass Bioenerg.* **2016**, *91*, 56–68.

(63) Zhang, L.; Larsson, A.; Moldin, A.; Edlund, U. Comparison of lignin distribution, structure, and morphology in wheat straw and wood. *Ind. Crops Prod.* **2022**, *187*, No. 115432.

(64) Branca, C.; Di Blasi, C. Oxidation reactivity of chars generated from the acid-catalyzed pyrolysis of corncobs. *Fuel Process. Technol.* **2014**, *123*, 47–56.

(65) Branca, C.; Di Blasi, C. Effects of heat/mass transfer limitations and process exothermicity on the kinetic parameters of the devolatilization and oxidation reactions of wood chars. *Thermochim. Acta* **2022**, *716*, No. 179321.

(66) Branca, C.; Di Blasi, C. A summative model for the pyrolysis reaction heats of beech wood. *Thermochim. Acta* **2016**, *638*, 10–16.