

Manuscript Number:

Title: Teamwork makes the dream work: disentangling cross-taxon congruence across soil biota in *Pinus nigra* Arnold plantations

Article Type: Research Paper

Keywords: Black pine; cross-taxon congruence; soil biodiversity; forest management; above/below-ground interactions

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Abstract: Soil biota plays a fundamental role in many ecological processes, throughout a complex network of above/below-ground interactions. Thus, there has been an increasing interest in the use of correlates for biodiversity assessment, demonstrating their reliability with respect to proxies based on environmental data alone. Despite co-variation of species richness and composition in grassland soils has been widely discussed in literature, only few studies were developed for forest artificial stands. In this context, a EU-Life project (SelPiBioLife, LIFE13 BIO/IT/000282) started in 2014 aiming at evaluating the application of an innovative forest management along with its effects on soil biodiversity in *Pinus nigra* plantations (Apennines, Italy). The robustness of cross-taxon congruence before any silvicultural treatment was performed across different taxa (bacteria, vascular plants, above ground mushrooms, ectomycorrhizae on root tips, mycelium in the soil, carabids, microarthropods, nematodes), also exploring how abiotic (soil and spatial-topographic variables) and forest biotic predictors (dendrometric variables) drive the community concordances among taxa, through the use of variation partitioning analysis. Correlations between groups were performed through Mantel and partial Mantel tests. Almost all the distribution patterns showed strong inter-group congruence and also a relationship with abiotic/biotic variables. However, only bacteria/mycelium and mushrooms/mycelium correlations remained significant after removing the environmental effect. Considering variation partitioning, the variance attributed solely to pure effect of biotic or abiotic predictors was significant only in some cases (e.g. bacteria); remarkably, in all dependent taxa, total and partial shared effect of all predictors always explained the highest portion of total -variation. In conclusion, the crucial importance of soil microbiome in affecting the ecosystem functioning of *Pinus nigra* forests was evaluated. Furthermore, results suggested a mutualistic relationship between all predictors, confirming the clutched network of ecological linkages in

soil as well as the impossibility to assess the mutual surrogate efficiency of taxa avoiding habitat influence.

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Siena, 01 August 2018

Subject: Submission to *Science of the Total Environment*

Dear Editor,

Attached for your consideration is an original manuscript entitled “Teamwork makes the dream work: disentangling cross-taxon congruence across soil biota in *Pinus nigra* Arnold plantations”. The manuscript is submitted as a research paper.

This work is part of a wider multidisciplinary project (SelPiBioLife, LIFE13 BIO/IT/000282), started in 2014, aiming at evaluating the application of an innovative forest management technique along with its effects on soil biodiversity in *Pinus nigra* plantations in two mountain areas of the Apennines (Italy). Using data obtained before any silvicultural treatment, the main aim of the present research was to test the robustness of cross-taxon congruence across organisms belonging to four kingdom-levels: Bacteria, Plantae (vascular plants), Fungi (studied as ectomycorrhizae (ECM) on root tips, mushrooms above ground and mycelium in the soil) and Animalia (carabids, microarthropods, nematodes), also exploring how abiotic (soil and spatial-topographic variables) and forest biotic predictors (dendrometric variables) drive the community concordances among taxa.

The “before treatment” research activity guaranteed a high-quality dataset concerning eight different groups of taxa and a complete and objective inventory of twenty-six environmental variables, comprising soil, spatial-topographic and dendrometric factors as proxies of forest structure. This research detains an interesting potential since, to the best of our knowledge, no other researches took into account as many soil taxa as those collected in the SelPiBioLife project, focusing both on biosphere as well as on lithosphere dynamics. For these reasons we are confident that our work could well fit the Aims and Scope of the journal.

Results pointed out the crucial importance of soil microbiome and microbial taxa diversity in affecting the ecosystem functioning of artificial black pine stands. Furthermore, the overall results suggested a mutualistic relationship between all the considered variables, both biotic and abiotic, confirming the complex network of clutched ecological linkages in black pine soils as well as the impossibility to assess the mutual surrogate efficiency of taxa avoiding the environmental and habitat influence.

We are confident that our manuscript will contribute to enhance knowledge about the ecological linkages between above/below-ground biota in affecting the ecosystem functioning in artificial black pine stands, also for application in biodiversity conservation purposes.

Thank you in advance for your attention,

With best regards,

Debora Barbato

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**Teamwork makes the dream work: disentangling cross-taxon congruence across soil biota in *Pinus nigra* Arnold
plantations**

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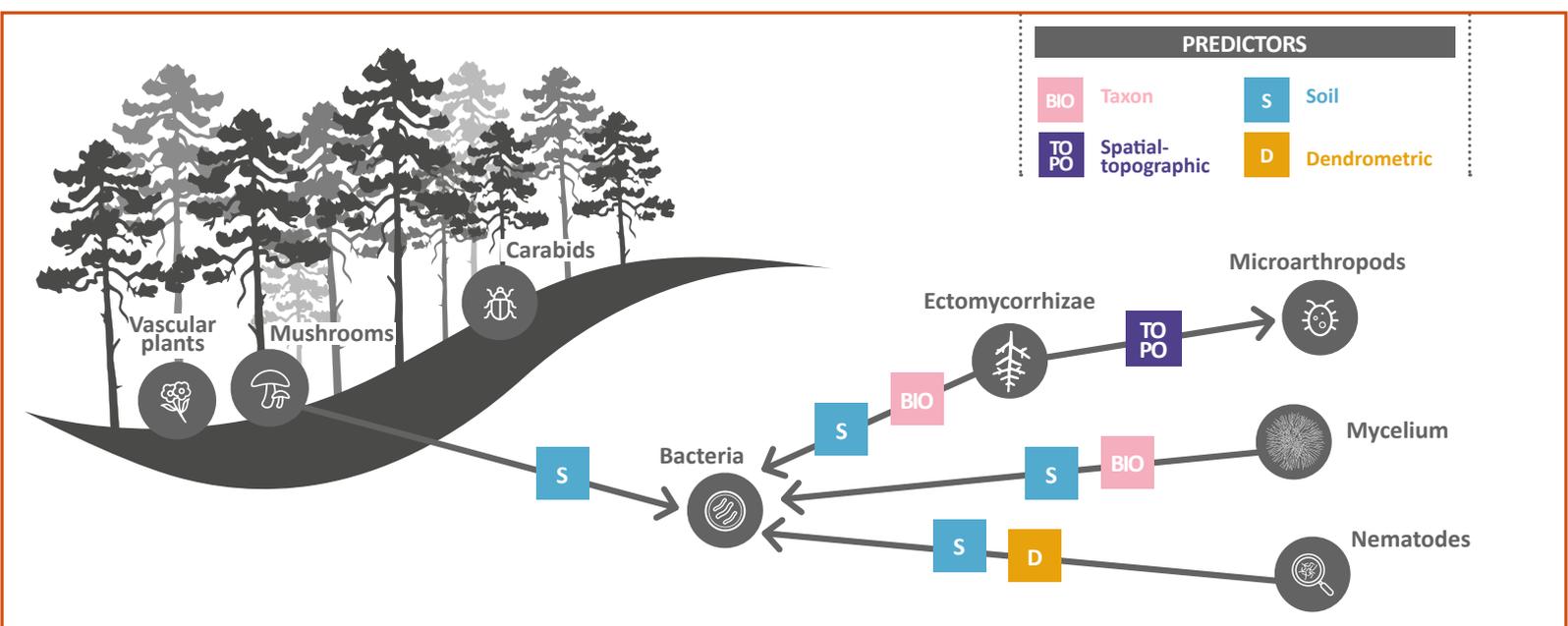
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Highlights

- We evaluated cross-taxon congruence in black pine artificial stands
- We analysed links between soil chemical properties, forest and soil biotic communities
- Fundamental role of soil microbiome
- Complex network of clustered ecological linkages in *Pinus nigra* soil
- Impossibility to assess cross-taxon efficiency avoiding the environmental influence

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27 **Abstract**

28

29 Soil biota plays a fundamental role in many ecological processes, throughout a complex network of above/below-
30 ground interactions. Thus, there has been an increasing interest in the use of correlates for biodiversity assessment,
31 demonstrating their reliability with respect to proxies based on environmental data alone. Despite co-variation of
32 species richness and composition in grassland soils has been widely discussed in literature, only few studies were
33 developed for forest artificial stands. In this context, a EU-Life project (SelPiBioLife, LIFE13 BIO/IT/000282) started
34 in 2014 aiming at evaluating the application of an innovative forest management along with its effects on soil
35 biodiversity in *Pinus nigra* plantations (Apennines, Italy). The robustness of cross-taxon congruence before any
36 silvicultural treatment was performed across different taxa (bacteria, vascular plants, above ground mushrooms,
37 ectomycorrhizae on root tips, mycelium in the soil, carabids, microarthropods, nematodes), also exploring how abiotic
38 (soil and spatial-topographic variables) and forest biotic predictors (dendrometric variables) drive the community
39 concordances among taxa, through the use of variation partitioning analysis. Correlations between groups were
40 performed through Mantel and partial Mantel tests. Almost all the distribution patterns showed strong inter-group
41 congruence and also a relationship with abiotic/biotic variables. However, only bacteria/mycelium and
42 mushrooms/mycelium correlations remained significant after removing the environmental effect. Considering variation
43 partitioning, the variance attributed solely to pure effect of biotic or abiotic predictors was significant only in some
44 cases (e.g. bacteria); remarkably, in all dependent taxa, total and partial shared effect of all predictors always explained
45 the highest portion of total variation. In conclusion, the crucial importance of soil microbiome in affecting the
46 ecosystem functioning of *Pinus nigra* forests was evaluated. Furthermore, results suggested a mutualistic relationship
47 between all predictors, confirming the clutched network of ecological linkages in soil as well as the impossibility to
48 assess the mutual surrogate efficiency of taxa avoiding habitat influence.

49

50

51 **Keywords:** Black pine; cross-taxon congruence; soil biodiversity; forest management; above/below-ground interactions

52

53 **1. Introduction**

54

55 Promoting environmental conservation efforts and mitigating the effects of human-induced changes, such as habitat
56 destruction and degradation and/or climate change, requires a solid understanding of the distribution and main drivers of
57 patterns of biodiversity in space and time (Gaston, 1996, 2000; Howard et al., 1998; Gioria et al., 2011). However, the
58 collection of extensive information on biodiversity is hampered by the complexity of spatio-temporal dimensions of this
59 variable. To address these issues, there has been an increasing interest in the use of correlates for biodiversity (Noss,
60 1990; Prendergast et al., 1993; Margules and Pressey, 2000), which proved that the use of surrogate taxa in
61 conservation planning is substantially more effective than that of surrogates based solely on environmental data
62 (Rodrigues and Brooks, 2007).

63 Cross-taxon congruence analysis can be expressed as the spatio-temporal correlation in patterns of species richness
64 and/or diversity (Pearson and Carroll, 1999). In other words, it occurs when diversity and/or compositional patterns of
65 different biological groups spatially covary following a common biogeographical history, a similar response to
66 environmental gradients or biotic interactions among taxa (Rooney and Azeria, 2015).

67 Despite some criticisms (Lewandowski et al., 2010; Westgate et al., 2014), the use of surrogates to predict community
68 patterns remains an effective tool, in particular when economical resources are scarce or where taxonomy cannot
69 provide a complete species inventory. Actually, the distribution of well-known taxa may provide insights into the
70 processes structuring the spatial distribution of other groups; furthermore, their identification could greatly aid
71 biodiversity monitoring and conservation initiatives (Duan et al., 2016).

72 Albeit most studies have traditionally examined cross-taxon congruency in above-ground terrestrial systems, some
73 attempts are needed to identify potential surrogacy even in soil bioindicators (Keith et al., 2012). In fact, soil hosts
74 approximately a quarter of Earth's biodiversity (FAO, 2015). This immense variety includes micro- (e.g. bacteria, fungi,
75 protozoa and nematodes), meso- (e.g. mites and springtails) and macro- (e.g. earthworms and termites) organisms and it
76 plays a fundamental role in several essential ecological processes, which are considered as life support functions (LSF,
77 Schlöter et al., 2018). In addition, soil interacts with above-ground biodiversity in a complex and intricate network of
78 biological activity (Wardle et al., 2004), so called "ecosystem multifunctionality" (Maestre et al. 2012).

79 A large body of recent literature deals with co-variation of species diversity in grasslands soils, especially species
80 richness, whereas few of them are in forests and even less rely to plantation forests (Irwin et al., 2014). As the area of
81 plantation forest expands worldwide and natural, unmanaged forests decline, the role of this artificial forest for
82 biodiversity conservation, especially those of conifer species, resulted to be not properly understood (Marchi et al.,

83 2018), representing a clear gap in the process of developing reliable bioindicators for soil health across a wide range of
84 different land use management and soil types (Keith et al., 2012).

85 Conifer forest plantations detain a long history, especially in Italy, where forest management has traditionally focused
86 not only on the mere implementation of timber production but also on the biodiversity conservation and soil protection
87 (Cantiani and Chiavetta, 2015).

88 The natural cycle of nutrients in a forest is highly influenced by the “soil system”, mainly through the degradation of
89 dead organic matter. Therefore, the fertility and sustainability of a natural soil depends significantly on the
90 transformation rate of organic materials mediated by the intricate soil biota network. Thus, modern forestry
91 management techniques should be able to meet the compromise between the economic needs of public and private
92 entities and the conservation of biodiversity, including all its soil components (Irwin et al., 2014) as well as the many
93 interactions between above-below ground subsystems and abiotic and biotic drivers (Wall et al., 2010; Wagg et al.,
94 2014, Duan et al., 2016, Fujii et al., 2017).

95 In this context, a multidisciplinary EU-Life project (SelPiBioLife, LIFE13 BIO/IT/000282) was established in 2014
96 aiming at demonstrating the application of an innovative forest management technique along with its effects on a wide
97 range of organisms belonging to four kingdom-levels: Bacteria, Plantae (vascular plants), Fungi (including above
98 ground mushrooms, ectomycorrhizae (ECM) on root tips and mycelium in the soil) and Animalia (carabids,
99 microarthropods, nematodes). The main goals of SelPiBioLife project were first to provide an accurate description of
100 the native biodiversity before any type of forest management (“before treatment” phase) and then to test the effects of a
101 selective silvicultural treatment on the overall level of soil biodiversity (“after treatment” phase) in *Pinus nigra*
102 plantations, using a multi-taxon analysis. This new approach is in line with the EU Biodiversity Strategy to 2020 and
103 the Global and European Atlas of Soil Biodiversity.

104 The “before treatment” research activity guaranteed a high-quality dataset, and a complete and objective inventory of
105 various environmental variables, comprising soil, spatial-topographic and dendrometric factors as proxies of forest
106 structure. This research detains an interesting potential since, to the best of our knowledge, no other researches took into
107 account as many soil taxa as those collected in the SelPiBioLife project to carry out a cross-taxon analysis.

108 Based on data collected in the “before treatment” phase of SelPiBioLife, the main aims of our study were:

- 109
- 110 • to test the significance of cross-taxon congruence in community composition of different biological groups in
111 black pine plantations;

- 112 • to assess the amount of variance explained by biotic (dendrometric variables) and abiotic predictors (soil
113 physico-chemical and spatial-topographic variables) along with their shared and unexplained effects driving the
114 community concordances among taxa;
- 115 • to explore ecological linkages between above/below-ground biota in affecting the ecosystem functioning in
116 artificial black pine stands.

117

118 **2. Materials and methods**

119

120 *2.1 Study area*

121 The study was carried out in two *Pinus nigra* Arn. plantations in Central Italy. The first study site (zone 1), called
122 Monte Amiata, it is located in Southern Tuscany (Castiglion d'Orcia, Siena; 42°56'8''N, 11°38'13''E, mean elevation
123 780 m a.s.l.). Average slope is 15% having mainly a N-E orientation. Geological substrate is characterized by clay,
124 calcareous and marly lithofacies, while soils are deep and very rich in organic matter. Erosion processes are apparently
125 lacking. Average annual temperature is 12.5°C (max: 21.7 °C in July; min: 4.5 °C in January). Mean annual rainfall add
126 up to about 687 mm, with November as the rainiest month. Pine forest is mostly dominated by *Pinus nigra* Arn. (more
127 than 90% of plants which are, on average, 44 years old), *Quercus pubescens* Wild. and *Quercus cerris* L. (Cantiani,
128 2016).

129 Pratomagno (zone 2) is in the Northeastern part of Tuscany (Pratomagno, Arezzo; lat 45°27'8''N, long 9°11'13''E,
130 mean elevation 960 m a.s.l.). It is lithologically characterized by quartz-feldspar sandstones alternated by siltstones and
131 argillites. Soils are generally moderately deep, though locally deeper, due to strong erosion. Average annual
132 temperature is 10.5°C (max: 19°C in July; min: 1.5°C in January). Rainfall follows the Apennines sub-mountain regime
133 (mean annual rainfall: 997 mm), with a maximum peak in autumn and a second one in spring. Pine forest has absolute
134 predominance of *Pinus nigra* Arn. with an average age of 57 years, which is locally associated with *Abies alba* Mill.
135 (especially at higher elevation) and occasionally with broadleaved species such as *Fagus sylvatica* L., *Fraxinus ornus*
136 L., and *Quercus cerris* L. (Cantiani, 2016).

137

138 *2.2 Sampling design*

139 Fifty-four sampling units (plots) were allocated according to a two-stage sampling design (Elzinga et al., 2001): in each
140 of the two zones a central homogeneous area of 20 ha was chosen. Then, nine squares sectors of 1 ha were marked on
141 the ground; afterwards, 3 concentric plots with a variable size spanning from 10 m radius for the biodiversity analysis to
142 15 m radius for the dendrometric and structural assessment were materialized in the field for forest surveys.

143 *2.3 Taxa datasets*

144 Eight different taxonomic groups (bacteria, vascular plants, above ground mushrooms, ectomycorrhizae (ECM) on root
145 tips, mycelium in the soil, carabids, microarthropods, nematodes) were sampled using separate protocols in 2014-2015
146 at plot or sector level, before adopting any silvicultural treatment. Organisms were identified to species or
147 morphospecies, except for bacteria, mycelium, microarthropods and nematodes that were identified to order
148 (micorarthropods), family (nematodes) or genus level (bacteria and mycelium). Prior to analysis, in order to obtain
149 comparable values, data at plot level (vascular plants, above ground mushrooms, ECM, microarthropods) were
150 aggregated at the sector spatial scale using different approaches: median method for vascular plants to reduce sensitivity
151 to outliers and sum method for the other groups. Abundances were then standardized to relative abundances (range
152 between 0-1). In total, eighteen sectors were used for the analyses: 9 relative to Monte Amiata and 9 for Pratomagno.

153 The sampling methodology for each taxonomic group is:

154

155 - for microbial analyses (bacteria and mycelium) an amount of 500 g of soil was randomly sampled in each sector from
156 the upper layer (0-20 cm) and homogenized for laboratory analysis. After sieving at 2mm, total DNA was extracted
157 from 0,5 g of soil by means of the commercial kit FastDNA™ SPIN Kit for Soil (MP Biomedicals). The
158 characterization of microbial community structure was carried out through a high-throughput sequencing approach by
159 means of Miseq Illumina technology (IGA Technology Services s.r.l., Italy), targeting the 16S DNA ribosomal genes
160 using primers 515F and 806R (Caporaso et al., 2012) and ITS using primers FF390 and FR1 (Vainio and Hantula,
161 2000).

162

163 - For vascular plants, the cover abundance was estimated in each plot for all species according to the Braun Blanquet
164 phytosociological method (Braun-Blanquet 1932). For an appropriate numerical treatment, cover- abundance values
165 were transformed according to van der Maarel method (van der Maarel, 2007).

166

167 - For above ground mushrooms, mycocoenological observations were made collecting and counting all epigeous fruit
168 bodies visible to the naked eye, larger than 1 mm according among others to Arnolds (1981), at plot level in each sector.
169 Sampling activity was performed every two weeks during the period of highest fungal production (autumn) while once
170 in spring. Species identification was performed with the usual morphological techniques and employing general analytic
171 keys and monographs.

172

173 - For ECM, mycorrhized root tips present in 1 soil core of 30 cm in length and 6 cm in diameter per plot were described

174 as morphotypes according to Agerer (1991, 1987-2002) and counted. Morphotypes were molecularly identified using a
175 direct PCR approach as described by Iotti and Zambonelli (2006).

176

177 - Carabids were sampled in each sector using three randomly located pitfall traps (Greenslade 1964; Adis 1979; Van
178 den Berghe 1992), characterized by plastic glasses (height: 12 cm; diameter to the mouth: 8.5 cm) buried up to the edge
179 and containing a saturated solution of sodium chloride, wine vinegar and little pure alcohol at 95% for the maintenance
180 of the sample. The collection of material was carried out at intervals of 10-15 days during the season of activity of
181 Coleoptera Carabidae.

182

183 - Concerning microarthropods, three soil samples were randomly collected at plot level in each sector, with a special 10
184 cm cubic corer for mesofauna sampling. Thereafter, extraction from soil samples was made using modified Berlese-
185 Tullgren funnels following the standard methodology (Parisi et al., 2005).

186

187 - For nematodes, three soil samples were randomly collected in each sector with a hand auger at the depth of 15 cm in
188 the top layer of bulk soil, after removing surface residues. Samples were then mixed to form a composite sample.
189 Nematodes were then extracted from 100 ml of soil using the cotton-wood filter method for 48 hours at room
190 temperature (approximately 20°C), counted and determined.

191

192 Further information on the adopted sampling design is available at www.selpibiolife.eu/en/

193

194 *2.4 Explanatory variables*

195 Soil physic-chemical properties, geographical coordinates, topographic factors and dendrometric parameters as proxies
196 of forest structure, were collected in each plot (Table 1).

197 A soil sample was randomly taken from the upper layer (0-20 cm) and then analyzed to determine the percentage of
198 clay (<0.002), coarse sand (2.0-0.2mm), fine sand (0.2-0.05mm), silt (0.050-0.002mm), total limestone, total nitrogen
199 and total organic matter. Soil pH and electric conductivity (ms/cm) were quantified too.

200 Measured dendrometric variables were: number of trees per hectare, basal area per hectare, average diameter at breast
201 height, average height of the stand (i.e total height of the plant of average diameter), dominant height (i.e arithmetic
202 mean of the total height of 100 trees per hectare with the largest diameter at breast height), diametric differentiation
203 (Pommerening, 2002) Clark and Evans Index (Clark and Evans, 1954), Vertical Evenness (Neuman and Starlinger,
204 2001), and Photosynthetic Active Radiation on the ground (PAR). Dendrometric variables are part of a wider and open-

205 access dataset, developed to assess the main structural and mensurational parameters of the studied black pine stands as
206 requested by SelPiBioLife project actions (Cantiani and Marchi, 2017).

207 To account for spatial-topographic factors, nine variables were included: aspect (sessagesimal degrees), elevation
208 (metres), flow direction of water (i.e. the direction of the greatest drop in elevation or the smallest rise if all neighbours
209 are higher), roughness (difference between the maximum and the minimum value of a cell and its 8 surrounding cells),
210 slope (percentage), Topographic Position Index (TPI; difference between the value of a cell and the mean value of its 8
211 surrounding cells), Terrain Ruggedness Index (TRI; the mean of the absolute differences between the value of a cell and
212 the value of its 8 surrounding cells), x and y geographical coordinates expressed as metric units in ETRS89/UTM 32N
213 reference system (EPSG 25832).

214 All environmental parameters obtained at plot level were aggregated at sector level using average method.

215

216 *2.5 Statistical analysis*

217 *2.5.1 Taxon Richness and Complementarity analyses*

218 To compare alpha and beta diversity among each taxon in relation to the sampling effort, rarefaction curves were
219 calculated using exact method. Generally, rarefaction curves enable to compare two or more datasets considering the
220 same sampling effort – in this case the same number of sampled sectors, nine in Amiata and nine in Pratomagno
221 (Gotelli and Colwell 2001; Bacaro et al. 2016).

222 For each taxon, classic Diversity indices (Simpson Index and the Pielou Equitability index) were calculated using the
223 available taxonomic resolution. Box and Whisker plots were used to represent the distribution of resulting values.
224 Moreover, aiming at testing differences in beta diversity among taxa, the R function ‘betadispersion 2’ proposed by
225 Bacaro et al. (2012, 2013) was applied using Bray-Curtis dissimilarity matrices as response variables. This procedure
226 consists of shuffling within taxon dissimilarities among taxa and disregarding between-taxa dissimilarities. In other
227 term, this analysis corresponds to a multivariate analogue of Levene's test for homogeneity of variances. By repeating
228 this operation using permutations (999), a distribution of the test statistics under the null hypothesis of no differences in
229 mean plot-to-plot dissimilarities within taxa was obtained. Where the calculated statistic resulted significant, pairwise
230 comparisons of group mean dispersions were evaluated using the Tukey's Honest Significant Differences between taxa
231 as proposed in Anderson (2006).

232

233 *2.5.2 Cross-taxon congruence between soil taxa and variation partitioning*

234 Mantel tests were used to perform pairwise cross-taxon correlation analyses among taxa, using Bray-Curtis dissimilarity
235 matrices. Similarly, Mantel correlation was calculated between each taxon and the Euclidean environmental distance

236 matrix obtained by all predictors (soil, spatial-topographic, dendrometric variables). Prior to the calculation,
237 environmental factors were standardized using *decostand* function with *normalize* method in R *vegan* package
238 (Oksanen et al., 2017).

239 Partial Mantel tests (Smouse et al., 1986) were then used to evaluate if a significant taxa concordance remained after the
240 conditional effect of the environment variables was removed. The significance of correlation of each Mantel and partial
241 Mantel statistic was obtained using 9999 permutations.

242 Using variation partitioning approach (Borcard et al., 1992; Legendre, 2008), we partitioned the variation in each
243 biological group that could be explained by another taxon as well as by the environment divided into three distinct
244 predictor sets: two abiotic (soil, spatial-topographic) and one biotic (dendrometric as a proxy of forest structure). The
245 outputs obtained allowed us to distinguish the proportion of total variation in each taxon set as dependent variable,
246 explained by the (a) pure effect of another taxon, (b) pure effect of soil, (c) pure effect of spatial-topographic predictors,
247 (d) pure effects of dendrometric variables, (e+f+g+h+i+j+k+l+m+n) partial shared effects of two/three set of factors, (o)
248 total shared effect of all the variables considered along with the variation unexplained.

249 The partitioning was based on the adjusted R^2 statistic as recommended by Peres-Neto et al., (2006); 999 permutations
250 were used to assess the significance of constraints.

251 Before variance partitioning analysis, we minimized multicollinearity by performing a Principal Component Analysis
252 (PCA) separately in each predictor set (taxon, soil, spatial-topographic, dendrometric), extracting the first axis of each
253 PCA (PC1 site scores) to use as explanatory variables in next analysis, for a total of four PC1 (PC1 taxon, PC1 soil,
254 PC1 topo and PC1 dendro).

255 All the statistical analyses were performed using R version 3.4.4 (R Development Core Team 2018).

256

257

258 **3. Results**

259

260 *3.1 Taxon Richness and Complementarity analyses*

261 Almost all the rarefaction curves, except mushrooms, reached the asymptote, meaning that the sampling efforts have
262 been reliable to capture the whole diversity in the study sites (Figure 1). Concerning patterns of diversity indices (both
263 Simpson and Pielou Equitability indices, Figure 2), alpha diversity was greater for bacteria, mycelium and plants
264 whereas nematodes and microarthropods were the less diverse groups in terms of taxonomic composition. Bacteria,
265 ECM and plants displayed the greater equitability (greater values of Pielou's evenness) whilst microarthropods and
266 nematodes showed the lower evenness meaning that is the presence of few entities that dominate the assemblage.

267 Beta diversity average dissimilarities from individual observation units to their group centroid in multivariate space
268 resulted to be significantly different among taxa ($F(1, 61320) = 5031.5, p < 0.001$). Boxplot displaying distance to
269 centroid for each taxon are reported in Figure 3; post-hoc tests are reported in Table 2.

270

271 3.2 Cross-taxon congruence between soil taxa

272 Considering Mantel test results (Table 3), the highest congruence was found between bacteria and mycelium ($r = 0.88$;
273 $p < 0.001$) and the lowest between nematodes and plants ($r = -0.02$; $p > 0.05$). In general, the distribution pattern of
274 almost all the groups analyzed showed highly supported inter-group congruence, while nematodes were not
275 significantly correlated with other taxa. All biological groups detained close relationships with the overall dataset of
276 environmental predictors: the highest one was found in bacteria ($r = 0.97$; $p < 0.001$) followed by mycelium ($r = 0.87$;
277 $p < 0.001$) and plants ($r = 0.84$; $p < 0.001$).

278 A partial Mantel test (Table 4), comparing the taxonomic matrices once the environmental matrix contribution was
279 removed, showed that only bacteria and mushrooms maintained a moderate and significant ($r = 0.24$; $p < 0.01$; $r = 0.19$
280 $p < 0.05$, respectively) correlation with mycelium, independent of environmental factors. No other cross-taxon
281 correlation remained significant.

282

283

284 3.3 Variation partitioning results

285 Considering the “pure_taxon” fraction, only ECM and mycelium were able to explain a significant degree of variance in
286 bacteria (respectively 2.9%, $p = 0.02$; 1.6%, $p = 0.05$; Figure 4C and 4F). The pure taxon effect reflects the cross-taxon
287 congruence between pairs of taxa not associated with environmental-spatial predictors and hence related purely to
288 proxies of potential biotic interactions. The other pure taxon fractions detained an irrelevant effect on all the other
289 groups.

290 In bacteria, the variation attributed solely to pure soil effect was 1.75% ($p = 0.05$), 2.19% ($p = 0.05$), 2.41% ($p = 0.04$),
291 2.14% ($p = 0.06$), when ECM (Figure 4C), mushrooms (Figure 4E), mycelium (Figure 4F) and nematodes (Figure 4G)
292 were set as independent variable, respectively.

293 Pure effect of spatial-topographic predictors is invoked to explain a large and significant portion of variation in
294 microarthropods, when ECM were one of the predictor taxon (11.40%, $p = 0.05$, Figure 4C). The unique significant pure
295 effect of dendrometric parameters as proxies of vegetation structure was found in bacteria when nematodes were set as
296 explanatory taxon (2.07%, $p = 0.05$, Figure 4G).

297 Remarkably, in all the dependent taxa, both totally shared and partially shared effects of all sets of predictors always
298 explained the highest portion of the total variation.

299

300

301 **4. Discussion**

302

303 Apart from some studies concerning the influence of forest management on microbiological soil properties (Lucas-
304 Borja et al., 2016), as well as the evaluation of diversity parameters in few well-known taxa (Spake et al., 2016;
305 Martinez-Jauregui et al., 2016; Mueller et al., 2016; Hanzelka and Reif, 2016), at present no studies have
306 simultaneously assessed diversity of various types of cryptic and less studied soil organisms in *Pinus nigra* plantation
307 forests.

308 The important role of soil microbiome and microbial taxa diversity was retrieved as fundamental in affecting the
309 ecosystem functioning of artificial black pine stands. In fact, bacteria and mycelium provided the highest alpha diversity
310 values, as they represent the smallest and the most abundant organisms within soil matrix, explaining a huge amount of
311 genetic and functional diversity in any environmental sample (e.g. Thompson et al., 2017). Moreover, the high values of
312 Pielou's evenness of bacteria (Figure 2, right) indicated a low environmental selective pressure on bacterial community,
313 thus with no significant dominant species; these results are in line with previous studies which reported that these
314 microorganisms do not respond to large-scale environmental gradients in the same way as other soil meso- or macro-
315 organisms do (Decaëns, 2010).

316 The distribution pattern of almost all the taxa showed highly supported inter-group congruence, displaying close
317 relationships with the overall dataset of environmental-spatial variables (Table 3).

318 However, only bacteria/mycelium and mushrooms/mycelium correlations remained significant after removing the
319 environmental effect (Table 4), demonstrating a close biotic bacterial-fungal interaction (BFI) (Deveau et al., 2018).

320 According to recent research, BFIs are more widespread than expected and their dynamics may be crucial for many
321 ecosystems functions (Bonfante and Anca, 2009; Mocali et al., 2015; Deveau et al., 2018). On the other hand, citing the
322 well-known wood wide web network existing underground between symbionts and the root system, it is not clear why
323 ECM/plants do not have a similar important correlation (Helgason et al., 1998)

324 Overall results suggested Bacteria as the taxon most explained by the set of variables used in this study. Furthermore,
325 they resulted to be involved in the only cases where the variance attributed solely to pure effect of biotic or abiotic
326 predictors was significant (Figure 4C, 4E, 4F, 4G). In particular, among all the biotic predictors, only ECM (Figure 4C)
327 and mycelium (Figure 4F) were able to explain a significant degree of variation ascribed to pure taxon effects in

328 bacterial communities. These results are consistent with the ecological linkages and biotic interactions in forest soil
329 microbiome. For instance, ECM and bacteria are well-known to be strictly associated in rhizospheres of forest trees
330 (Baldrian, 2017b): this is demonstrated also for *Pinus sylvestris* rhizosphere where ectomycorrhizal fungi select for
331 distinct and specific associated bacterial populations (Nurmiaho-Lassila et al., 1997; Marupakula et al., 2016).
332 Mycorrhizae associated with forest trees can also model the composition of local microbial communities, selecting for a
333 broad spectrum of species with distinct enzyme functions and thus providing potential proxies for key biogeochemical
334 processes (Cheeke et al., 2017). Moreover, a large number of fungal and bacterial families engage complex interactions,
335 ranging from mutualism and antagonism (Deveau et al., 2018): while some bacteria are well known to have a positive
336 effect on symbiosis, the so-called "mycorrhiza helper bacteria" (MHB) (Garbaye, 1994; Frey-Klett et al., 2007), others
337 may have negative effect on fungi, causing diseases or using mycophagy as strategy to derive nutrition from living
338 fungi (Leveau and Preston, 2008). The co-occurrence of fungi and bacteria may result in unique contributions to
339 biogeochemical cycles (Tarkka and Deveau, 2016): for example, organic matter transformations as well as nutrient
340 bioavailability in soil strictly depend on the interaction between fungi and bacteria (Baldrian, 2017a).

341 The importance of soil as life support system was confirmed by its significant pure contribution towards bacteria
342 variability when fungi (mushrooms, ECM, mycelium) were set as explanatory variables (Figure 4E, 4C, 4F,
343 respectively). In general, spatial heterogeneity of forest topsoils determines the composition of microbial communities
344 mainly through soil/litter chemistry and vegetation composition. The latter was reported being the second major driver
345 shaping microbial communities after soil features and it detains a greater effect on fungal communities than on bacteria
346 (Baldrian, 2017a). Thus, whereas soil pH is recognized as one of the best predictors for bacterial community
347 composition, fungal community structure is more closely associated to other soil features, nutrient status or tree
348 diversity (Lauber et al., 2008). In this study, it is likely that some environmental parameter such as soil properties (i.e.
349 organic matter, soil texture, C/N ratio, etc.) or the presence of specific predators (i.e. mites and collembola), rather than
350 vegetation diversity, provided direct or indirect selection of fungal community, according with previous analyses (i.e.
351 Dirilgen et al., 2016). Despite most studies reported that soil fungal communities of forest soils are strongly dependent
352 on vegetation type (Shi et al., 2014; Urbanová et al., 2015; Baldrian, 2017b), fungal diversity appeared to be unlikely
353 related to plant diversity, with the obvious exception of ECM, suggesting that plant-soil feedbacks seem not to influence
354 the diversity of soil fungi at the global scale (Tedersoo et al., 2014). However, as distinct microbial communities
355 develop on decomposing leaf litters of different tree species, the comparisons amongst microbial communities in
356 different litters and forest floors at the same stage of decay are needed, in order to definitively deduce the influence of
357 tree species. Thus, disentangling the controversial linkages between tree species and soil microbial communities

358 requires the consideration of several factors, including soil pedoclimatic features, root exudates and the effects of soil
359 fauna (Prescott and Graystone, 2013).

360 Pure soil effect appeared to be significant when nematodes were set as predictor taxon to explain bacterial variance
361 (Figure 4G). In fact, bacterial diversity is strictly related to free nematodes community composition in soil as it includes
362 several bacteria-feeding groups, which can significantly alter the proportion of bacteria by grazing (Sundin et al., 1990;
363 Yeates et al., 1993; Yeates, 2003). Additionally, the relationship between nematodes and bacteria might be highly
364 specific, resulting in mutualistic or even symbiotic interactions (Akhurst, 1980; Forst and Clarke, 2002). Ingham et al.
365 (1985) found up to 70% of the bacterial- and fungal-feeding nematodes in the 4–5% of the total of a thin soil-layer
366 around the rhizosphere. The significance of the relationship between dendrometric parameters and nematode-bacteria
367 interaction (Figure 4G) could likely have its explanation in the variation of radical biomass, which in turn depends on
368 the epigeal biomass of the forest (Price et al., 2010; Pretzsch et al., 2012, Schepaschenko et al., 2017). In fact, the
369 radical horizon represents the natural environment of this trophic liaison (Wang et al., 2000; Eisenhauer et al. 2017).
370 However, little is still known about the relationship between plant structural components and biodiversity in managed
371 forests (Jokela et al., 2018) and further in-depth analysis are needed.

372 Because elevation gradients show a large number of correlated environmental factors, some studies have used
373 topographic factors as a surrogate for habitat heterogeneity (Moura et al., 2016). Habitat heterogeneity on its side means
374 extensive trophic niche differentiation and diverse resources availability. The significant pure effect of spatial-
375 topographic predictors in microarthropods, when ECM were one of the predictor taxon (Figure 4C), could make sense
376 observing the distribution of food resources, since it has been demonstrated the strong feeding preference of Oribatida
377 and Protura on certain ECM (Schneider and Maraun, 2005; Malmström and Persson, 2011).

378 Remarkably, in all the dependent taxa, total shared and partial shared effect of all sets of predictors always explained
379 the highest portion of total variation, testifying the highly intricate and dynamic interplay of environmental factors and
380 the potential biotic interactions in explaining cross-taxon congruence in turnover patterns in *Pinus nigra* plantations
381 (Duan et al., 2016).

382 Although above–below ground biota have been traditionally considered as isolated and independent subsystems
383 (Wardle et al., 2004), our study confirmed the complex network of clutched ecological linkages in controlling
384 ecosystem properties and processes in *Pinus nigra* soil. Considering the composite interaction between organisms
385 belonging to different taxonomic level, it resulted very difficult to assign some biota exclusively to the above or below-
386 ground environment, demonstrating the indissoluble nature of all biological relationships within soil. Furthermore, taxa
387 traditionally used in cross-taxon congruence analysis such as vascular plants and carabids, did not give further
388 information as above ground indicators, shifting the focus on the importance of below-ground system in regulating the

389 whole network of ecological processes. For this reason, a better understanding of the factors that influence soil
390 biodiversity as well as the interaction mechanisms and consequences is pivotal.

391 **5. Conclusions**

392 Our research disentangled cross-taxon congruence across soil biota of four Kingdoms in *Pinus nigra* plantations, for the
393 first time using a high quality multi-taxonomic dataset and an objective inventory of various environmental variables,
394 both biotic and abiotic. This study highlighted the crucial importance of soil microbiome and especially bacteria as
395 general biodiversity indicator potential in *Pinus nigra* plantations. However, although the understanding of bacterial
396 ecology in forest soils has advanced dramatically in recent years, it is still incomplete. The exact extent of the
397 contribution of bacteria to forest ecosystem processes will be recognized only in the future, when the activities of all
398 soil community members will be studied simultaneously (Lladò et al., 2017). Apart from the fundamental role of
399 microbial taxa, this research demonstrated that it is not a unique factor but rather the mutualistic relationship of all
400 variables, both biotic and abiotic, to regulate the above-below ground subsystems in *Pinus nigra* plantations. Purging
401 cross-taxon congruence from the effect of environment, only bacteria/mycelium and mushrooms/mycelium
402 relationships maintained a strong covariation, showing the impossibility to assess the mutual surrogate efficiency of
403 taxa avoiding the environmental and habitat influence. For this reason, in the development of improved indicators of
404 soil quality in artificial black pine stands, it is decisive to analyze not only multiple taxa but also their relationships with
405 biotic/abiotic features, in order to disentangle all the ecological linkages between above-below ground biota.

406

407

408 **Acknowledgements**

409 This work was supported by the SelPiBio LIFE project (Innovative silvicultural treatments to enhance soil biodiversity
410 in artificial black pine stands, i.e., LIFE13 BIO/IT/000282) for demonstration of innovative silvicultural treatments in
411 artificial black pine stands.

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414 **Tables captions**

415

416 **Table 1** List of soil, dendrometric and spatial-topographic variables used as environmental predictor factors

417

418 **Table 2** Output of Tukey' HSD on beta diversity analysis among taxa. Asterisks express statistical significance: *** =

419 < 0.05; *ns* = not significant

420

421 **Table 3** Mantel Test - Pearson's product moment. Asterisks express statistical significance: *** = $p < 0.001$; ** = $p <$

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438 pure or combined effect of another taxon (main title of x axis), soil, spatial-topographic and dendrometric predictor sets
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440 (b) pure effect of soil (pure_soil), (c) pure effect of spatial-topographic predictors (pure_topo), (d) pure effects of
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443 Unexplained variation was provided too. In the graphical representation, negative values of R^2 (i.e cases where the

444 explanatory variables explain less variation than random normal variable) were interpreted as zeros. For this reason,
445 overall explained variance could exceed 100% in some cases.
446

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448

Table 1 List of soil, dendrometric and spatial-topographic variables used as environmental predictor factors

449

Predictor set	Variable name	Variable description
<i>Soil</i>	clay	% clay (<0.002)
	coarse sand	% coarse sand (2.0-0.2mm)
	fine sand	% fine sand (0.2-0.05mm)
	silt	% silt (0.050-0.002mm)
	tot limestone	% total limestone
	nitro	% total nitrogen
	organic matter	% total organic matter
	e-cond	electric conductivity (ms/cm)
	pH	soil pH
<i>Dendrometric</i>	nr ha	number of trees per hectare
	g ha	basal area per hectare
	d aver	average diameter at breast height
	h aver	average height of the stand
	h dom	dominant height
	size siff	diametric differentiation
	CE	Clark and Evans Index
	VE	Vertical Evenness
	PAR	Photosyntetic Active Radiation on the ground
	<i>Spatial-topographic</i>	aspect
elevation		elevation (m),
flowdir		flow direction of water
rough		roughness
tpi		Topographic Position Index
tri		Terrain Ruggedness Index
x		x geographical coordinates (metric units in ETRS89/UTM 32N)
y		y geographical coordinates (metric units in ETRS89/UTM 32N)

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453 **Table 2** Output of Tukey' HSD on beta diversity analysis among taxa. Asterisks express statistical significance: *** =454 < 0.05; *ns* = not significant

455

<i>Taxa</i>	Bacteria	Carabids	ECM	Microarthropods	Mushrooms	Mycelium	Nematodes	Vascular plants
Bacteria	-							
Carabids	***	-						
ECM	***	<i>ns</i>	-					
Microarthropods	***	*	***	-				
Mushrooms	***	<i>ns</i>	***	***	-			
Mycelium	***	***	***	<i>ns</i>	***	-		
Nematodes	***	<i>ns</i>	***	<i>ns</i>	**	**	-	
Vascular plants	***	<i>ns</i>	***	<i>ns</i>	***	***	<i>ns</i>	-

456

457 **Table 3** Mantel Test - Pearson's product moment. Asterisks express statistical significance: *** = $p < 0.001$; ** = $p <$ 458 0.01 ; * = $p < 0.05$; *ns* = not significant

<i>Taxa</i>	Bacteria	Carabids	ECM	Microarthropods	Mushrooms	Mycelium	Nematodes	Vascular plants	Environment
Bacteria	-								
Carabids	0.416**	-							
ECM	0.592***	0.252*	-						
Microarthropods	0.391**	0.065	0.269**	-					
Mushrooms	0.683***	0.343**	0.284**	0.326**	-				
Mycelium	0.876***	0.380**	0.552***	0.316**	0.672***	-			
Nematodes	0.091	0.118	0.129	0.142	0.136	0.050	-		
Vascular plants	0.787***	0.325*	0.5401***	0.344**	0.513***	0.740***	-0.020	-	
Environment	0.966***	0.451**	0.615***	0.394**	0.692***	0.875***	0.110*	0.838***	-

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464 **Table 4** Partial Mantel Test – Environmental effect - Pearson's product moment. Asteriks express statistical

<i>Taxa</i>	Bacteria	Carabids	ECM	Microarthropods	Mushrooms	Mycelium	Nematodes	Vascular plants
Bacteria	-							
Carabids	-0.086	-						
ECM	-0.009	-0.036	-					
Microarthropods	0.047	-0.137	0.038	-				
Mushrooms	0.080	0.048	-0.247	0.081	-			
Mycelium	0.245**	-0.034	0.036	-0.063	0.192*	-		
Nematodes	-0.057	0.077	0.080	0.108	0.085	-0.094	-	
Vasculat plants	-0.158	-0.107	0.057	0.027	-0.170	0.027	-0.207	-

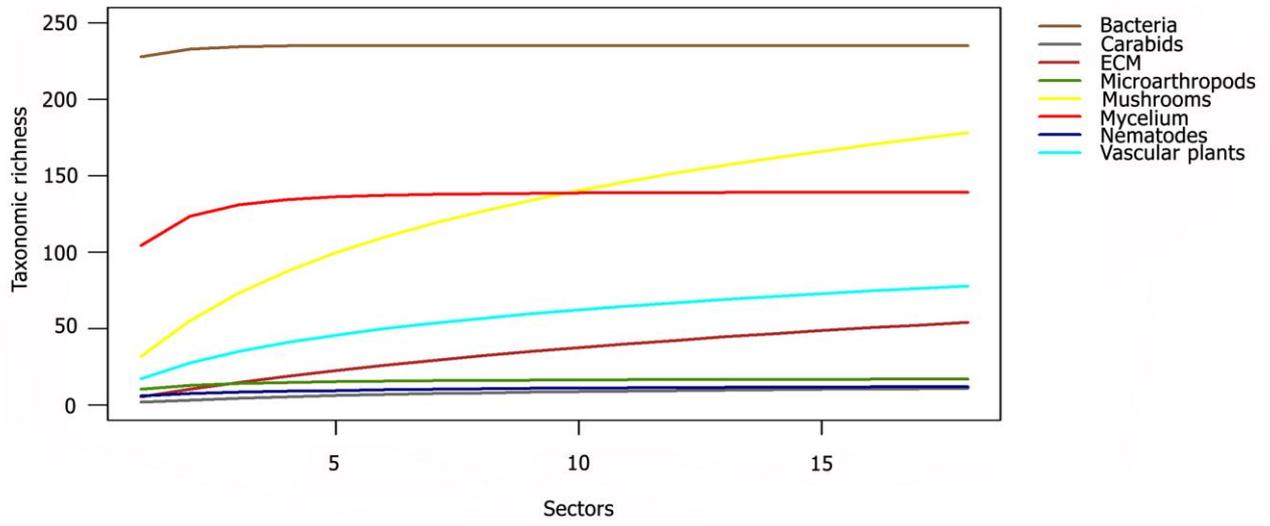
465 significance: *** = $p < 0.001$; ** = $p < 0.01$; * = $p < 0.05$; ns = not significant

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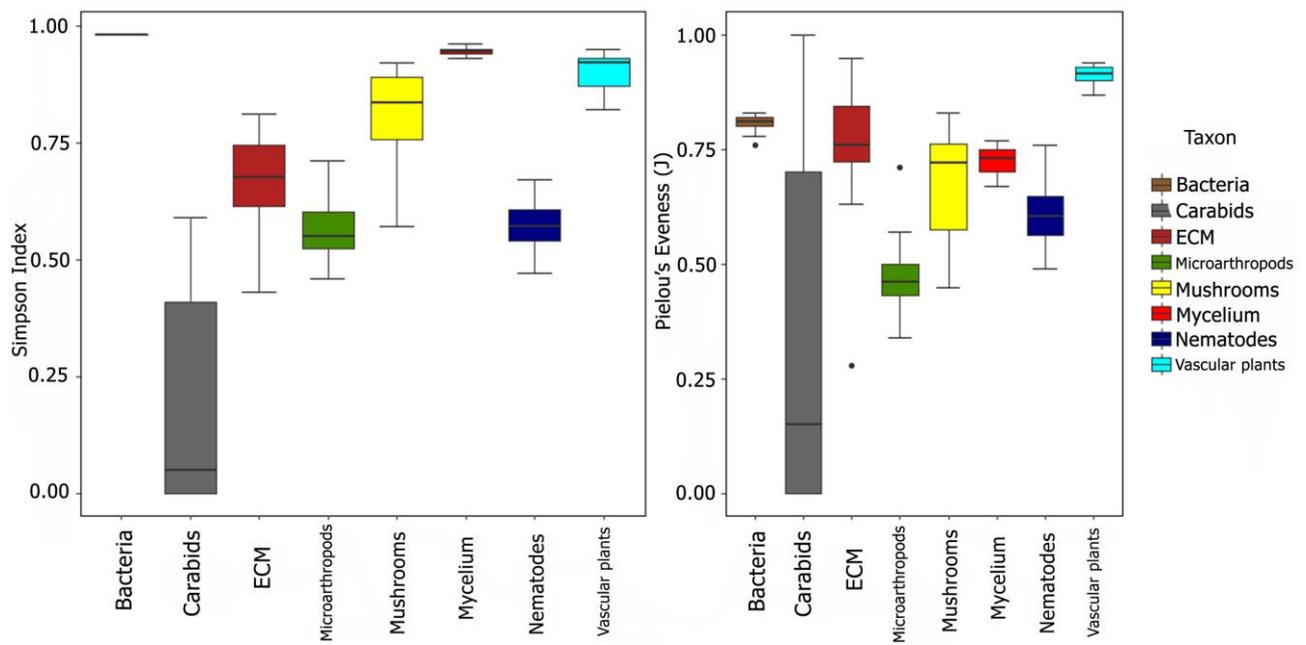
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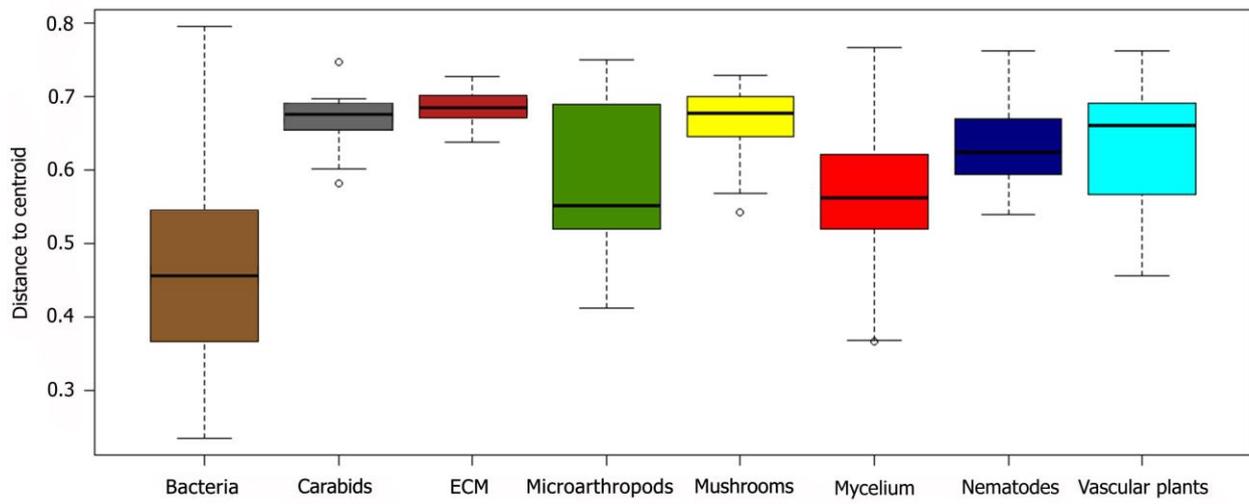


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476 **Figure 2** Summary of Simpson (left) and Pielou Equitability indices (right) for each taxon

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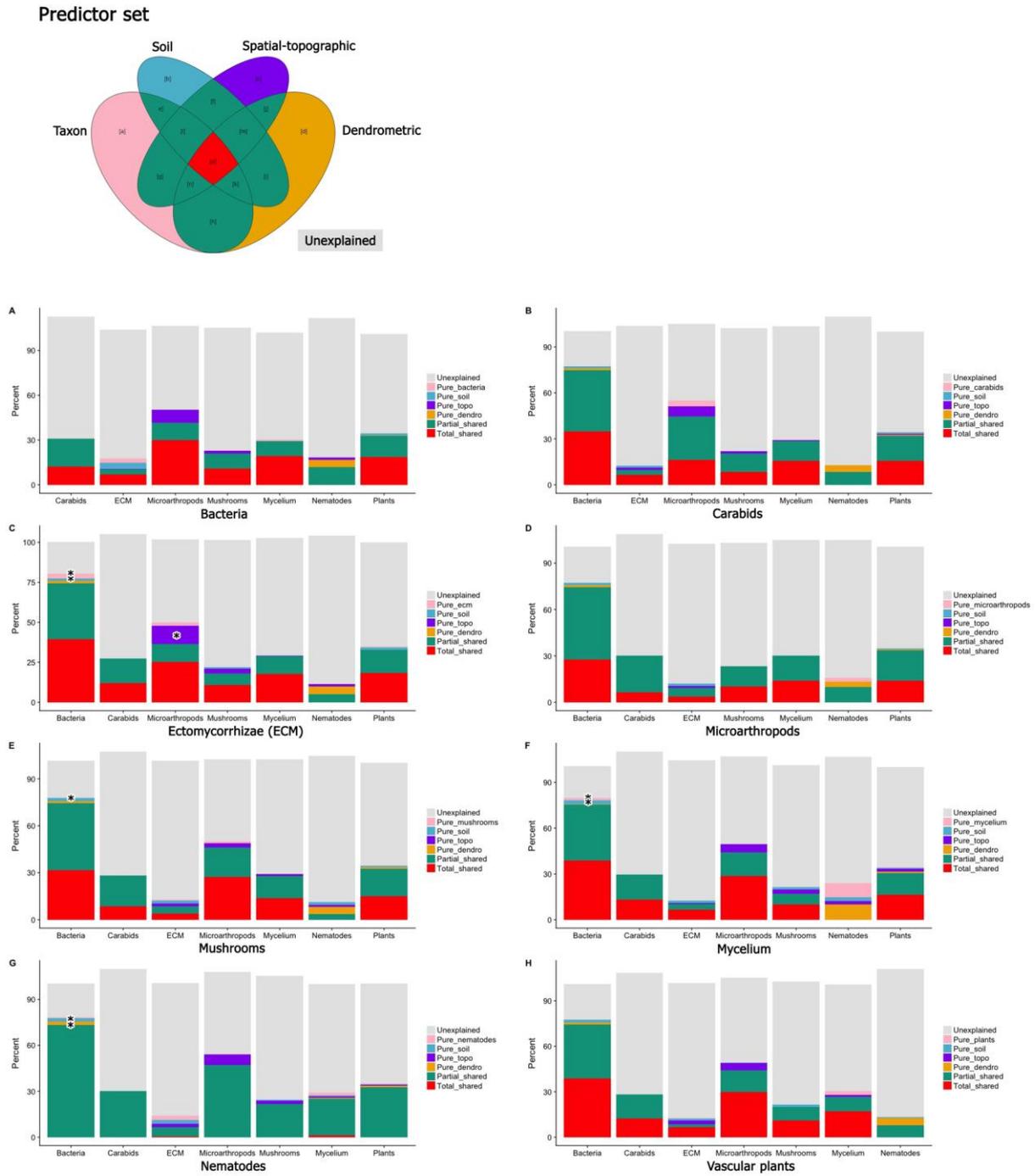


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480 **Figure 3** Boxplot displaying taxonomic β -diversity. Distance to centroid were measured by calculating the Bray Curtis

481 distance of group members to the group centroid for each taxon

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Figure 4 Variation partitioning of total variation of each taxon as dependent variable (barcharts categories) showing the pure or combined effect of another taxon (main title of x axis), soil, spatial-topographic and dendrometric predictor sets as independent variable(s). Venn’s diagram (above) can be portioned in: (a) pure effect of another taxon (pure_taxon), (b) pure effect of soil (pure_soil), (c) pure effect of spatial-topographic predictors (pure_topo), (d) pure effects of dendrometric variables (pure_dendro), (e+f+g+h+i+j+k+l+m+n) partial shared variation explained by the overlapping effects of two/three set of factors, (partial_shared), (o) total shared effect between all sets of predictors (total_shared).

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493 overall explained variance could exceed 100% in some cases.

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496 **References**

497

498 Adis, J., 1979. Problems of interpreting arthropod sampling with pitfall traps. *Zool. Anz.* 202:177-184.

499 Agerer, R., 1991. Characterization of ectomycorrhiza. In: Norris JR, Read DJ, Varma A (Eds.), *Techniques for the*
500 *study of mycorrhiza. Methods in microbiology vol. 23. Academic Press, London, pp. 25-73.*

501 Agerer, R., 1987-2002. *Colour Atlas of Ectomycorrhizas.* Einhorn-Verlag. Schwäbisch Gmünd.

502 Akhurst, R.J., 1980. Morphological and functional dimorphism in *Xenorhabdus* spp., bacteria symbiotically associated
503 with the insect pathogenic nematodes *Neoplectana* and *Heterorhabditis*. *Microbiology* 121 (2):303-309.

504 Anderson, M.J., 2006. Distance-based tests for homogeneity of multivariate dispersions. *Biometrics* 62 (1):245–253.

505 Arnolds, E., 1981. Ecology and coenology of macrofungi in grassland and moist heathland in Drenthe, the Netherlands.
506 *Biblioteca Mycologica*, 407 pp.

507 Bacaro, G., Altobelli, A., Cameletti, M., Ciccarelli, D., Martellos, S., Palmer, M.W., Ricotta, C., Rocchini, D., Scheiner,
508 S.M., Tordoni, E., Chiarucci, A., 2016. Incorporating spatial autocorrelation in rarefaction methods: Implications for
509 ecologists and conservation biologists. *Ecol. Indic.* 69:233-238.

510 Bacaro, G., Gioria, M., Ricotta, C., 2012. Testing for differences in beta diversity from plot-to-plot dissimilarities. *Ecol.*
511 *Res.* 27:285–292.

512 Bacaro, G., Gioria, M., Ricotta, C., 2013. Beta diversity reconsidered. *Ecol. Res.* 28:537–540

513 Baldrian, P., 2017a. Forest microbiome: diversity, complexity and dynamics. *Fems Microbiol. Rev.* 41 (2):109-130.

514 Baldrian, P., 2017b. Microbial activity and the dynamics of ecosystem processes in forest soils. *Curr. Opin.*
515 *Microbiol.* 37:128-134.

516 Bonfante, P., Anca, I.A., 2009. Plants, mycorrhizal fungi, and bacteria: a network of interactions. *Annu. Rev.*
517 *Microbiol.* 63:363-383.

518 Borcard, D., Legendre, P., Drapeau, P., 1992. Partialling out the spatial component of ecological variation. *Ecology*
519 73:1045-1055.

520 Braun-Blanquet, J., 1932. *Plant sociology: the study of plant communities.* McGraw-Hill, New York.

521 Cantiani, P., 2016. Il diradamento selettivo. Accrescere stabilità e biodiversità in boschi artificiali di pino nero. Manuale
522 tecnico SelPiBioLife. Compagnia delle Foreste. Arezzo, Italia, 62 pp.

523 Cantiani, P., Chiavetta, U., 2015. Estimating the mechanical stability of *Pinus nigra* Arn. using an alternative approach
524 across several plantations in central Italy. *iForest* 8 (6):846.

525 Cantiani, P., Marchi, M., 2017. A spatial dataset of forest mensuration collected in black pine plantations in central
526 Italy. *Ann. For. Sci.* 74 (3):50.

527 Caporaso, J.G., Lauber, C.L., Walters, W.A., Berg-Lyons, D., Huntley, J., Fierer, N., Owens, S.M., Betley, J., Fraser, L.,
528 Bauer, M., Gormley, N., Gilbert, J.A., Smith, G., Knight, R., 2012. Ultra-high-throughput microbial community
529 analysis on the Illumina HiSeq and MiSeq platforms. *ISME J.* 6:1621–1624

530 Cheeke, T.E., Phillips, R.P., Brzostek, E.R., Rosling, A., Bever, J.D., Fransson, P., 2017. Dominant mycorrhizal
531 association of trees alters carbon and nutrient cycling by selecting for microbial groups with distinct enzyme
532 function. *New Phytol.* 214 (1):432-442.

533 Clark, P.J., Evans, F.C., 1954. Distance to nearest neighbor as a measure of spatial relationships in populations. *Ecology*
534 445-453.

535 Decaëns, T., 2010. Macroecological patterns in soil communities. *Global Ecol. Biogeogr.* 19 (3):287-302.

536 Deveau, A., Bonito, G., Uehling, J., Paoletti, M., Becker, M., Bindschedler, S., Hacquard, S., Hervé, V., Labbé, J.,
537 Lasovetsky, O.A. Mieszkin, S., Millet, L.J., Vajna, B., Junier, P., Bonfante, P., Krom, B.P., Olsson, S., van Elsas,
538 J.D., Wick, L.Y., 2018. Bacterial-Fungal Interactions: ecology, mechanisms and challenges. *Fems Microbiol. Rev.*
539 42 (3):335-352.

540 Dirilgen, T., Arroyo, J., Dimmers, W.J., Faber, J., Stone, D., da Silva, P.M., Carvalho, F., Schmelz, R., Griffiths, B.S.,
541 Francisco, R., Creamer, R.E., Sousa, J.-P., Bolger, T., 2016. Mite community composition across a European
542 transect and its relationships to variation in other components of soil biodiversity. *Appl. Soil Ecol.* 97:86-97.

543 Duan, M., Liu, Y., Yu, Z., Baudry, J., Li, L., Wang, C., Axmacher, J.C., 2016. Disentangling effects of abiotic factors
544 and biotic interactions on cross-taxon congruence in species turnover patterns of plants, moths and beetles. *Sci. Rep-
545 UK* 6:23511.

546 Eisenhauer, N., Lanoue, A., Strecker, T., Scheu, S., Steinauer, K., Thakur, M. P., Mommer, L., 2017. Root biomass and
547 exudates link plant diversity with soil bacterial and fungal biomass. *Sci. Rep-UK* 7:44641.

548 Elzinga, C.L., Salzer, D.W., Willoughby, J.W., Gibbs, J.P., 2001. *Monitoring Plant and Animal Populations.* Blackwell
549 Science, Malden, Massachussets.

550 FAO, 2015. Soils and biodiversity. Soils host a quarter of our planet's biodiversity. Viewed online at:
551 <http://www.fao.org/3/a-i4551e.pdf>

552 Forst, S., Clarke, D., 2002. Bacteria-nematode symbiosis. *Entomopathogenic nematology*, 57-77.

553 Frey-Klett, P., Garbaye, J. A., Tarkka, M., 2007. The mycorrhiza helper bacteria revisited. *New Phytol.* 176 (1):22-36.

554 Fujii, S., Mori, A.S., Koide, D., Makoto, K., Matsuoka, S., Osono, T., Isbell, F., 2017. Disentangling relationships
555 between plant diversity and decomposition processes under forest restoration. *J. Appl. Ecol.* 54 (1):80-90.

556 Garbaye, J., 1994. Tansley review no. 76 helper bacteria: a new dimension to the mycorrhizal symbiosis. New
557 Phytol. 128 (2):197-210.

558 Gaston, K.J., 1996. Biodiversity. A Biology of Number and Differences. Blackwell, Oxford

559 Gaston, K.J., 2000. Global patterns in biodiversity. Nature 405 (6783):220.

560 Gioria, M., Bacaro, G., Feehan, J., 2011. Evaluating and interpreting cross-taxon congruence: potential pitfalls and
561 solutions. Acta Oecol. 37 (3):187-194.

562 Gotelli, N.J., Colwell, R.K., 2001. Quantifying biodiversity: procedures and pitfalls in the measurement and comparison
563 of species richness. Ecol. Letters 4:379–391.

564 Greenslade, P.J.M., 1964. Pitfall trapping as a method for studying populations of Carabidae (Coleoptera). J. Anim.
565 Ecol. 33:301-310.

566 Hanzelka, J., Reif, J., 2016. Effects of vegetation structure on the diversity of breeding bird communities in forest
567 stands of non-native black pine (*Pinus nigra* A.) and black locust (*Robinia pseudoacacia* L.) in the Czech Republic.
568 Forest Ecol. Manag. 379:102-113.

569 Helgason, T., Daniell, T. J., Husband, R., Fitter, A. H., Young, J. P. W., 1998. Ploughing up the wood-wide web?
570 Nature 394 (6692):431.

571 Howard, P.C., Viskanic, P., Davenport, T.R., Kigenyi, F.W., Baltzer, M., Dickinson, C.J., Lwanga, J.S., Matthews,
572 R.A., Balmford, A., 1998. Complementarity and the use of indicator groups for reserve selection in Uganda. Nature
573 394 (6692):472.

574 Ingham, R.E., Trofymow, J.A., Ingham, E.R., Coleman, D.C., 1985. Interactions of bacteria, fungi, and their nematode
575 grazers: effects on nutrient cycling and plant growth. Ecol. Monogr. 55 (1):119-140.

576 Iotti, M., Zambonelli, A., 2006. A quick and precise technique for identifying ectomycorrhizas by PCR. Mycol. Res.
577 110: 60-65.

578 Irwin, S., Pedley, S.M., Coote, L., Dietzsch, A.C., Wilson, M.W., Oxbrough, A., Sweeney, O., Moore, K.M., Martin,
579 R., Kelly, D.L., Mitchell, F.J.G., Kelly, T.C., O'Halloran, J., 2014. The value of plantation forests for plant,
580 invertebrate and bird diversity and the potential for cross-taxon surrogacy. Biodivers. Conserv. 23 (3):697-714.

581 Jokela, J., Juutilainen, K., Korpela, L., Kouki, J., Kuntsi, S., Koivula, M., Siitonen, J., 2018. Cross-taxon congruence
582 and relationships to stand characteristics of vascular plants, bryophytes, polyporous fungi and beetles in mature
583 managed boreal forests. Ecol. Indic. 85:137-145.

584 Keith, A.M., Boots, B., Hazard, C., Niechoj, R., Arroyo, J., Bending, G.D., Bolger, T., Breen, J., Clipson, N., Doohan,
585 F.M., Griffin, C.T., Schmidt, O., 2012. Cross-taxa congruence, indicators and environmental gradients in soils under
586 agricultural and extensive land management. Eur. J. Soil Biol. 49:55-62.

587 Lauber, C.L., Strickland, M.S., Bradford, M.A., Fierer, N., 2008. The influence of soil properties on the structure of
588 bacterial and fungal communities across land-use types. *Soil Biol. Biochem.* 40 (9):2407-2415.

589 Legendre, P., 2008. Studying beta diversity. *Ecological variation partitioning by multiple regression and canonical*
590 *analysis.* *J. Plant Ecol.* 1:3–8.

591 Leveau, J.H., Preston, G.M., 2008. Bacterial mycophagy: definition and diagnosis of a unique bacterial–fungal
592 interaction. *New Phytol.* 177 (4):859-876.

593 Lewandowski, A.S., Noss, R.F., Parsons, D.R., 2010. The effectiveness of surrogate taxa for the representation of
594 biodiversity. *Conserv. Biol.* 24 (5):1367-1377.

595 Lladó, S., López-Mondéjar, R., Baldrian, P., 2017. Forest soil bacteria: diversity, involvement in ecosystem processes,
596 and response to global change. *Microbiol. Mol. Biol. R.* 81 (2):e00063-16.

597 Maestre, F.T., Quero, J.L., Gotelli, N.J., Escudero, A., Ochoa, V., Delgado-Baquerizo, M., □ García-Gómez M.,
598 Bowker, M.A., Soliveres, S., Escolar, C., García-Palacios, P., Berdugo, M., Valencia, E., Gozalo, B., Gallardo, A.,
599 Aguilera, L., Arredondo, T., Blones, J., Boeken, B., Bran, D., Conceição, A.A., Cabrera, O., Chaieb, M., Derak,
600 M., Eldridge, D.J., Espinosa, C.I., Florentino, A., Gaitán, J., Gabriel Gatica, M., Ghiloufi, W., Gómez-González, S.,
601 Gutiérrez, J.R., Hernández, R.M., Huang, X., Huber-Sannwald, E., Jankju, M., Miriti, M., Moneris, J., Mau, R.L.,
602 Morici, E., Naseri, K., Ospina, A., Polo, V., Prina, A., Pucheta, E., Ramírez-Collantes, D.A., Romão, R., Tighe,
603 M., Torres-Díaz, C., Val, J., Veiga, J.P., Wang, D., Zaady, E., 2012. Plant species richness and ecosystem
604 multifunctionality in global drylands. *Science* 335 (6065):214-218.

605 Malmström, A., Persson, T., 2011. Responses of Collembola and Protura to tree girdling—some support for
606 ectomycorrhizal feeding. *Soil Org.* 83:279-285.

607 Marchi, M., Paletto, A., Cantiani, P., Bianchetto, E., De Meo, I., 2018. Comparing Thinning System Effects on
608 Ecosystem Services Provision in Artificial Black Pine (*Pinus nigra* JF Arnold) Forests. *Forests* 9 (4):188.

609 Margules, C.R., Pressey, R. L., 2000. Systematic conservation planning. *Nature*, 405 (6783):243.

610 Martínez-Jauregui, M., Díaz, M., de Ron, D.S., Soliño, M., 2016. Plantation or natural recovery? Relative contribution
611 of planted and natural pine forests to the maintenance of regional bird diversity along ecological gradients in
612 Southern Europe. *Forest Ecol. Manag.* 376:183-192.

613 Marupakula, S., Mahmood, S., Finlay, R.D., 2016. Analysis of single root tip microbiomes suggests that distinctive
614 bacterial communities are selected by *Pinus sylvestris* roots colonized by different ectomycorrhizal fungi. *Environ.*
615 *Microbiol.* 18 (5):1470-1483.

616 Mocali, S., Landi, S., Curto, G., Dallavalle, E., Infantino, A., Colzi, C., d'Errico, G., Roversi, P.F., D'Avino, L.,
617 Lazzeri, L., 2015. Resilience of soil microbial and nematode communities after biofumigant treatment with defatted
618 seed meals. *Ind. Crop. Prod.* 75:79-90.

619 Moura, M.R., Villalobos, F., Costa, G.C., Garcia, P.C., 2016. Disentangling the role of climate, topography and
620 vegetation in species richness gradients. *PloS one*, 11 (3):e0152468.

621 Mueller, K.E., Eisenhauer, N., Reich, P.B., Hobbie, S.E., Chadwick, O.A., Chorover, J., Dobies, T., Hale, C.M.,
622 Jagodziński, A.M., Kalucka, I., Kasprovicz, M., Kieliszewska-Rokicka, B., Modrzyński, J., Rozen, A., Skorupski,
623 M., Sobczyk, Ł., Stasińska, M., Trocha, L.K., Weiner, J., Wierzbicka, A., Oleksyn, J., 2016. Light, earthworms, and
624 soil resources as predictors of diversity of 10 soil invertebrate groups across monocultures of 14 tree species. *Soil*
625 *Biol. Biochem.* 92:184-198.

626 Neuman, M., Starlinger, F., 2001. The significance of different indices for stand structure and diversity in forests. *Forest*
627 *Ecol. Manag.* 145:91-106.

628 Noss, R.F., 1990. Indicators for monitoring biodiversity: a hierarchical approach. *Conserv. Biol.* 4 (4), 355-364.

629 Nurmiaho-Lassila, E.L., Timonen, S., Haahtela, K., Sen, R., 1997. Bacterial colonization patterns of intact *Pinus*
630 *sylvestris* mycorrhizospheres in dry pine forest soil: an electron microscopy study. *Can. J. Microbiol.* 43 (11):1017-
631 1035.

632 Oksanen, J., Blanchet F.G., Friendly M., Kindt R., Legendre P., McGlinn D., Minchin P.R., O'Hara R.B., Simpson
633 G.L., Solymos P., Stevens M.H.H., Szoecs, E., Wagner, H., 2017. *vegan: Community Ecology Package*. R package
634 version 2.4-4. <https://CRAN.R-project.org/package=vegan>.

635 Parisi, V., Menta, C., Gardi, C., Jacomini, C., Mozzanica, E., 2005. Microarthropod communities as a tool to assess soil
636 quality and biodiversity: a new approach in Italy. *Agr. Ecosys. Environ.* 105:323-333.

637 Pearson, D.L., Carroll, S.S., 1999. The influence of spatial scale on cross-taxon congruence patterns and prediction
638 accuracy of species richness. *J. Biogeogr.* 26:1079–1090.

639 Peres-Neto, P., Legendre, P., Dray, S., Borcard, D., 2006. Variation partitioning of species data matrices: estimation and
640 comparison of fractions. *Ecology* 87:2614-2625

641 Pommerening, A., 2002. Approaches to quantifying forest structures. *Forestry* 75:305-324.

642 Prendergast, J.R., Quinn, R.M., Lawton, J.H., Eversham, B.C., Gibbons, D.W., 1993. Rare species, the coincidence of
643 diversity hotspots and conservation strategies. *Nature* 365 (6444):335.

644 Prescott, C.E., Grayston, S.J., 2013. Tree species influence on microbial communities in litter and soil: current
645 knowledge and research needs. *Forest Ecol. Manag.* 309:19-27.

646 Pretzsch, H., Biber, P., Uhl, E., Hense, P., 2012. Coarse root–shoot allometry of *Pinus radiata* modified by site
647 conditions in the Western Cape province of South Africa. *South. Forests* 74 (4):237-246.

648 Price, C.A., Gilooly, J.F., Allen, A.P., Weitz, J.S., Niklas, K.J., 2010. The metabolic theory of ecology: prospects and
649 challenges for plant biology. *New Phytol.* 188:696–710.

650 R Development Core Team, 2018. R: A language and environment for statistical computing. R Foundation for
651 Statistical Computing. Vienna, Austria. Available at: <http://cran.rproject.org/>.

652 Rodrigues, A.S., Brooks, T.M., 2007. Shortcuts for biodiversity conservation planning: the effectiveness of surrogates.
653 *Annu. Rev. Ecol. Evol. Syst.*, 38:713-737.

654 Rooney, R.C., Azeria, E.T., 2015. The strength of cross- taxon congruence in species composition varies with the size
655 of regional species pools and the intensity of human disturbance. *J. Biogeogr.* 42 (3):439-451.

656 Shi, L.L., Mortimer, P.E., Slik, J.F., Zou, X.M., Xu, J., Feng, W.T., Qiao, L., 2014. Variation in forest soil fungal
657 diversity along a latitudinal gradient. *Fungal Divers.* 64 (1):305-315.

658 Schepaschenko, D., Shvidenko, A., Usoltsev, V., Lakyda, P., Luo, Y., Vasylyshyn, R., Lakyda, I., Myklush, Y., See, L.,
659 McCallum, I., Fritz, S., Kraxner, F., Obersteiner, M., 2017. A dataset of forest biomass structure for
660 Eurasia. *Scientific Data* 4:170070.

661 Schloter, M., Nannipieri, P., Sørensen, S.J., van Elsas, J.D., 2018. Microbial indicators for soil quality. *Biol. Fert. Soils*
662 54 (1):1-10.

663 Schneider, K., Maraun, M., 2005. Feeding preferences among dark pigmented fungal taxa (“Dematiacea”) indicate
664 limited trophic niche differentiation of oribatid mites (Oribatida, Acari). *Pedobiologia* 49 (1):61-67.

665 Smouse, P.E., Long J.C., Sokal R.R., 1986. Regression and Correlation Extensions of the Mantel Test of Matrix
666 Correspondence. *Syst. Zool.* 35 (4):627-632.

667 Spake, R., Barsoum, N., Newton, A.C., Doncaster, C.P., 2016. Drivers of the composition and diversity of carabid
668 functional traits in UK coniferous plantations. *Forest Ecol. Manag.* 359:300-308.

669 Sundin, P., Valeur, A., Olsson, S., Odham, G., 1990. Interactions between bacteria-feeding nematodes and bacteria in
670 the rape rhizosphere: effects on root exudation and distribution of bacteria. *FEMS Microbiol. Lett.* 73 (1):13-22.

671 Tarkka, M., Deveau, A., 2016. An Emerging Interdisciplinary Field: Fungal–Bacterial Interactions. In *Environmental*
672 *and Microbial Relationships* (pp. 161-178). Springer, Cham.

673 Tedersoo, L., Bahram, M., Põlme, S., Kõljalg, U., Yorou, N. S., Wijesundera, R., Villarreal Ruiz, L., Vasco-Palacios,
674 A.M., Quang Thu, P., Suija, A., Smith, M.E., Sharp, C., Saluveer, E., Saitta, A., Rosas, M., Riit, T., Ratkowsky, D.,
675 Pritsch, K., Põldmaa, K., Piepenbring, M., Phosri, C., Peterson, M., Parts, K., Pärtel, K., Otsing, E., Nouhra, E.,
676 Njouonkou, A.L., Nilsson, R.H., Morgado, L.N., Mayor, J., May, T.W., Majuakim, L., Lodge, D.J., Lee, S.S.,

677 Larsson, K.-H., Kohout, P., Hosaka, K., Hiiesalu, I., Henkel, T.W., Harend, H., Guo, L.-d., Greslebin, A., Grelet, G.,
678 Geml, J., Gates, G., Dunstan, W., Dunk, C., Drenkhan, R., Dearnaley, J., De Kesel, A., Dang, T., Chen, X.,
679 Buegger, F., Brearley, F.Q., Bonito, G., Anslan, S., Abell, S., Abarenkov, K., 2014. Global diversity and geography
680 of soil fungi. *Science* 346 (6213):1256688.

681 Thompson, L.R., Sanders, J.G., McDonald, D., Amir, A., Ladau, J., Locey, K.J., Prill, R.J., Tripathi, A., Gibbons, S.M.,
682 Ackermann, G., Navas-Molina, J.A., Janssen, S., Kopylova, E., Vázquez-Baeza, Y., González, A., Morton, J.T.,
683 Mirarab, S., Zech Xu, Z., Jiang, L., Haroon, M.F., Kanbar, J., Zhu, Q., Song, S., Kosciulek, T., Bokulich, N.A.,
684 Lefler, J., Brislawn, C.J., Humphrey, G., Owens, S.M., Hampton-Marcell, J., Berg-Lyons, D., McKenzie, V., Fierer,
685 N., Fuhrman, J.A., Clauset, A., Stevens, R.L., Shade, A., Pollard, K.S., Goodwin, K.D., Jansson, J.K., Gilbert, J.A.,
686 Knight R., the Earth Microbiome Project Consortium., 2017. A communal catalogue reveals Earth's multiscale
687 microbial diversity. *Nature* 551(7681).

688 Urbanová, M., Šnajdr, J., Baldrian, P., 2015. Composition of fungal and bacterial communities in forest litter and soil is
689 largely determined by dominant trees. *Soil Biol. Biochem.* 84:53-64.

690 Vainio, E.J., Hantula, J., 2000. Direct analysis of wood-inhabiting fungi using denaturing gradient gel electrophoresis of
691 amplified ribosomal DNA. *Mycol. Res.* 104:927–936

692 van der Maarel, E., 2007. Transformation of cover- abundance values for appropriate numerical
693 treatment- Alternatives to the proposals by Podani. *J. Veg. Sci.* 18 (5):767-770.

694 Wagg, C., Bender, S.F., Widmer, F., van der Heijden, M.G., 2014. Soil biodiversity and soil community composition
695 determine ecosystem multifunctionality. *P. Natl. Acad. Sci. USA* 111 (14):5266-5270

696 Wall, D.H., Bardgett, R.D., Kelly, E., 2010. Biodiversity in the dark. *Nat. Geosci.* 3 (5):297-298

697 Wang, J.R., Letchford, T., Comeau, P., Kimmins, J.P., 2000. Above-and below-ground biomass and nutrient
698 distribution of a paper birch and subalpine fir mixed-species stand in the Sub-Boreal Spruce zone of British
699 Columbia. *Forest Ecol. Manag.* 130 (1-3):17-26.

700 Wardle, D.A., Bardgett, R.D., Klironomos, J.N., Setälä, H., van der Putten, W.H., Wall, D.H., 2004. Ecological
701 linkages between aboveground and belowground biota. *Science* 304 (5677):1629-33

702 Westgate, M.J., Barton, P.S., Lane, P.W., Lindenmayer, D.B., 2014. Global meta-analysis reveals low consistency of
703 biodiversity congruence relationships. *Nat. Commun.* 5:3899

704 Yeates, G. W., 2003. Nematodes as soil indicators: functional and biodiversity aspects. *Biol. Fert. Soils* 37 (4):199-210.

705 Yeates, G.W., Bongers, T.D., De Goede, R.G.M., Freckman, D.W., Georgieva, S.S., 1993. Feeding habits in soil
706 nematode families and genera—an outline for soil ecologists. *J. Nematol.* 25 (3):315.

Table 1[Click here to download Table: Table 1.doc](#)**Table 1** List of soil, dendrometric and spatial-topographic variables used as environmental predictor factors.

Predictor set	Variable name	Variable description
<i>Soil</i>	clay	% clay (<0.002)
	coarse sand	% coarse sand (2.0-0.2mm)
	fine sand	% fine sand (0.2-0.05mm)
	silt	% silt (0.050-0.002mm)
	tot limestone	% total limestone
	nitro	% total nitrogen
	organic matter	% total organic matter
	e-cond	electric conductivity (ms/cm)
	pH	soil pH
<i>Dendrometric</i>	nr ha	number of trees per hectare
	g ha	basal area per hectare
	d aver	average diameter at breast height
	h aver	average height of the stand
	h dom	dominant height
	size siff	diametric differentiation
	CE	Clark and Evans Index
	VE	Vertical Evenness
	PAR	Photosyntetic Active Radiation on the ground
<i>Spatial-topographic</i>	aspect	aspect (sessagesimal degrees)
	elevation	elevation (m),
	flowdir	flow direction of water
	rough	roughness
	tpi	Topographic Position Index
	tri	Terrain Ruggedness Index
	x	x geographical coordinates (metric units in ETRS89/UTM 32N)
	y	y geographical coordinates (metric units in ETRS89/UTM 32N)

Table 2[Click here to download Table: Table 2.doc](#)

Table 2 Output of Tukey' HSD on beta diversity analysis among taxa. Asterisks express statistical significance: *** = < 0.05; *ns* = not significant

<i>Taxa</i>	Bacteria	Carabids	ECM	Microarthropods	Musbrooms	Mycelium	Nematodes	Vascular plants
Bacteria	-							
Carabids	***	-						
ECM	***	<i>ns</i>	-					
Microarthropods	***	*	***	-				
Mushrooms	***	<i>ns</i>	***	***	-			
Mycelium	***	***	***	<i>ns</i>	***	-		
Nematodes	***	<i>ns</i>	***	<i>ns</i>	**	**	-	
Vascular plants	***	<i>ns</i>	***	<i>ns</i>	***	***	<i>ns</i>	-

Table 3[Click here to download Table: Table 3.doc](#)

Table 3 Mantel Test - Pearson's product moment. Asterisks express statistical significance: *** = < 0.05; *ns* = not significant

<i>Taxa</i>	Bacteria	Carabids	ECM	Microarthropods	Mushrooms	Mycelium	Nematodes	Vascular plants	Environment
Bacteria	-								
Carabids	0.416**	-							
ECM	0.592***	0.252*	-						
Microarthropods	0.391**	0.065	0.269**	-					
Mushrooms	0.683***	0.343**	0.284**	0.326**	-				
Mycelium	0.876***	0.380**	0.552***	0.316**	0.672***	-			
Nematodes	0.091	0.118	0.129	0.142	0.136	0.050	-		
Vascular plants	0.787***	0.325*	0.5401***	0.344**	0.513***	0.740***	-0.020	-	
Environment	0.966***	0.451**	0.615***	0.394**	0.692***	0.875***	0.110*	0.838***	-

Table 4[Click here to download Table: Table 4.doc](#)

Table 4 Partial Mantel Test – Environmental effect - Pearson’s product moment. Asterisks express statistical significance: *** = $p < 0.001$; ** = $p < 0.01$; * = $p < 0.05$; *ns* = not significant

<i>Taxa</i>	Bacteria	Carabids	ECM	Microarthropods	Mushrooms	Mycelium	Nematodes	Vascular plants
Bacteria	-							
Carabids	-0.086	-						
ECM	-0.009	-0.036	-					
Microarthropods	0.047	-0.137	0.038	-				
Mushrooms	0.080	0.048	-0.247	0.081	-			
Mycelium	0.245**	-0.034	0.036	-0.063	0.192*	-		
Nematodes	-0.057	0.077	0.080	0.108	0.085	-0.094	-	
Vasculat plants	-0.158	-0.107	0.057	0.027	-0.170	0.027	-0.207	-

Figure 1

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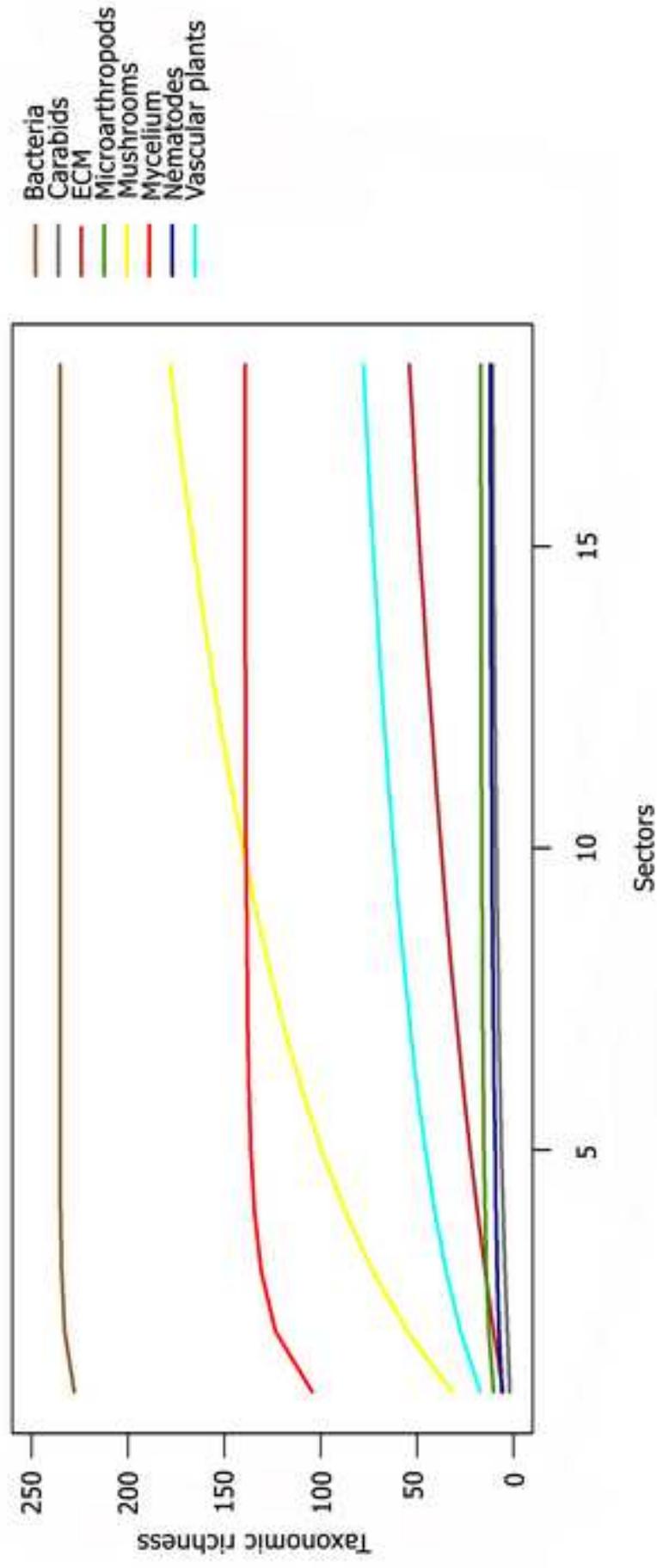


Figure 2
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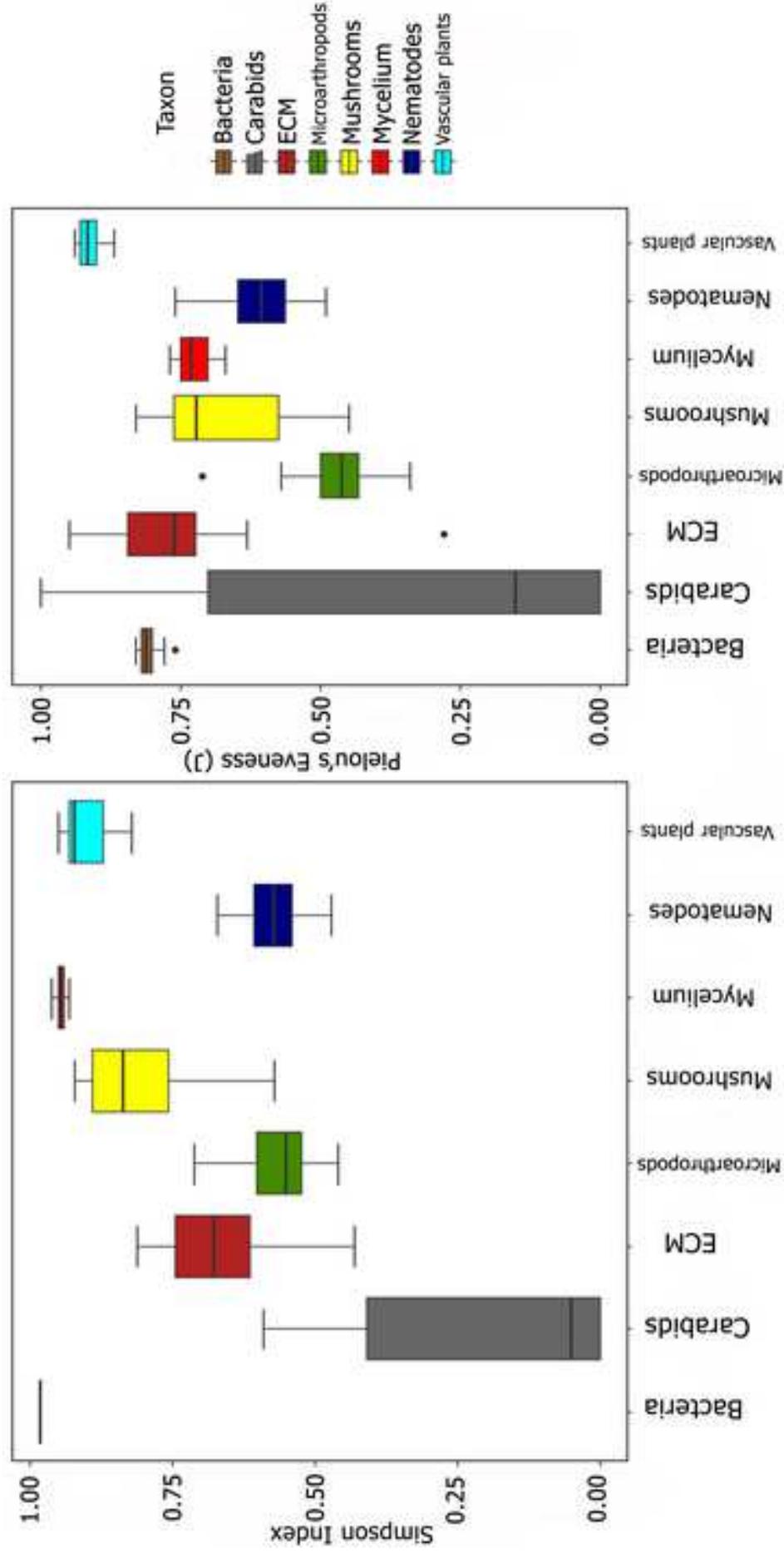


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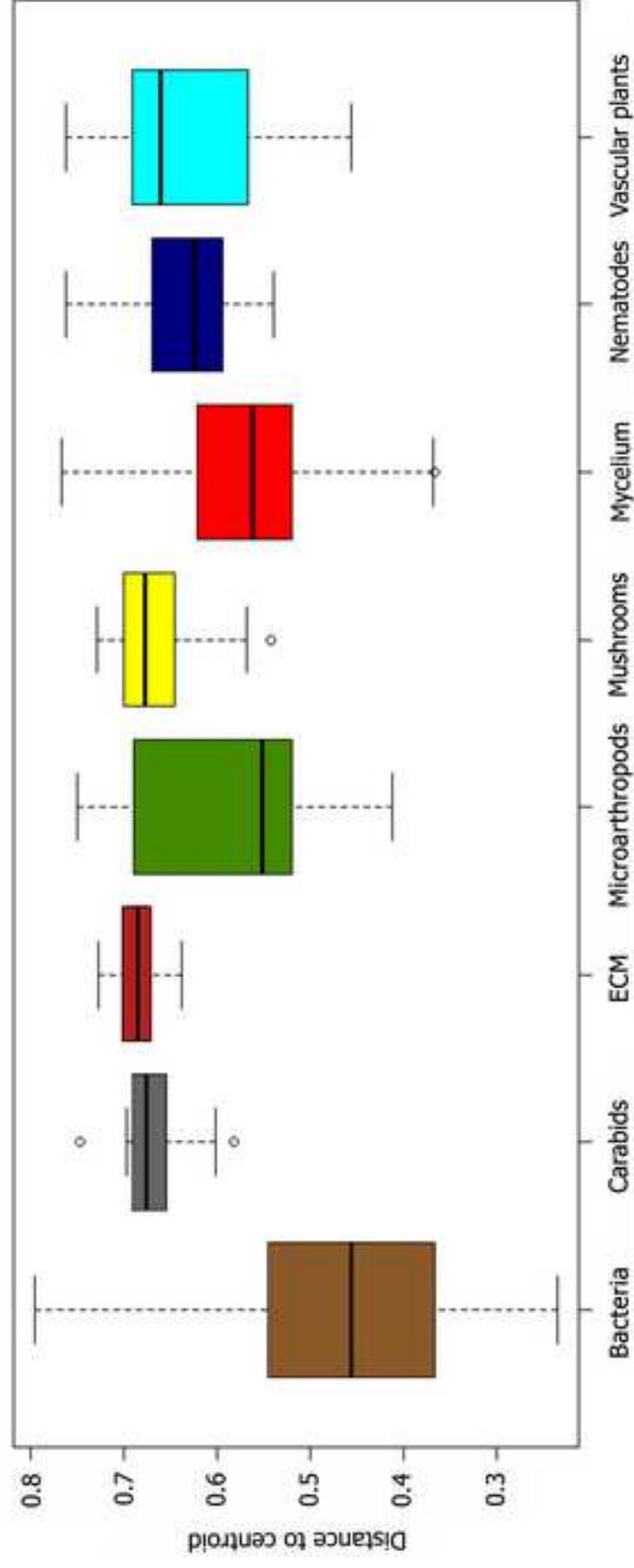


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