# Classifying degrees of species commonness: North Sea fish as a case study

Gianpaolo Coro<sup>a,1,2,\*</sup>, Thomas J. Webb<sup>b</sup>, Ward Appeltans<sup>c</sup>, Nicolas Bailly<sup>d</sup>,
 André Cattrijsse<sup>e</sup>, Pasquale Pagano<sup>a</sup>

 <sup>a</sup>Istituto di Scienza e Tecnologie dell'Informazione "Alessandro Faedo" – CNR, Pisa, Italy
 <sup>b</sup>Department of Animal & Plant Sciences, University of Sheffield, Sheffield S10 2TN, UK
 <sup>c</sup>Intergovernmental Oceanographic Commission (IOC) of UNESCO, Oostende, Belgium
 <sup>d</sup>WorldFish, Penang, Malaysia
 <sup>e</sup>Vlaams Instituut voor de Zee (VLIZ), Oostende, Belgium

## 11 Abstract

Species commonness is often related to abundance and species conservation status. Intuitively, a "common species" is a species that is abundant in a certain area, widespread and at low risk of extinction. Analysing and classifying species commonness can help discovering indicators of ecosystem status and can prevent sudden changes in biodiversity. However, it is challenging to quantitatively define this concept. This paper presents a procedure to automatically characterize species commonness from biological surveys. Our approach uses clustering analysis techniques and is based on a number of numerical parameters extracted from an authoritative source of biodiversity data, i.e. the Ocean Biogeographic Information System. The analysis takes into account abundance, geographical and temporal aspects of species distributions. We apply our model to North Sea fish species and show that the classification agrees with independent expert opinion although sampling

\*Corresponding author. Preprint sabmitted to Ecological Modelling Email addresses: coro@isti.cnr.it (Gianpaolo Coro),

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t.j.webb@sheffield.ac.uk (Thomas J. Webb), w.appeltans@unesco.org (Ward Appeltans), n.bailly@cgiar.org (Nicolas Bailly), andre.cattrijsse@vliz.be (André Cattrijsse), pagano@isti.cnr.it (Pasquale Pagano)

<sup>&</sup>lt;sup>1</sup>Telephone Number: +39 050 315 2978

<sup>&</sup>lt;sup>2</sup>Fax Number:  $+39\ 050\ 621\ 3464$ 

biases affect the data. Furthermore, we show that our approach is robust to noise in the data and is promising in classifying new species. Our method can be used in conservation biology, especially to reduce the effects of the sampling biases which affect large biodiversity collections.

- <sup>12</sup> Keywords: Species Commonness, OBIS, Conservation biology, North Sea,
- <sup>13</sup> Clustering, D4Science

## 14 1. Introduction

The term "common species" refers intuitively to a species that is abun-15 dant in a certain area, widespread and at low risk of extinction. By con-16 sequence, "rare species" are less abundant and possibly threatened. Auto-17 matically detecting common and rare species, and how their status changes 18 through time, is an important step in understanding the consequences of en-19 vironmental change for ecosystem functioning. In particular, the abundance 20 of a species in a community or ecosystem is a key indicator of its ecological 21 role and ecosystem function therefore depends on the identities and relative 22 numbers of common and rare species [1]. For instance, rare species may have 23 unique functional traits [2] and make particular contributions to diversity 24 [3]. On the other hand, common species may underpin ecosystem function 25 where they dominate in terms of biomass [4, 5, 6]. Both human activity 26 and natural environmental change typically affect the relative abundances 27 of species [7]. Monitoring changes in the relative abundance of species is 28 straightforward when working on individual, well-monitored systems. How-29

ever, anthropogenic-driven environmental change is affecting entire ecosys-30 tems, requiring large-scale ecological efforts [8]. One approach to monitor 31 species commonness at large scale and in a certain time frame, is to perform 32 meta-analyses on studies of multiple individual communities. This is useful 33 for extracting general trends across multiple taxa [9]. An alternative is to 34 take advantage of the increasing availability of large-scale compilations of 35 biodiversity data, such as the UK's National Biodiversity Network (NBN) 36 [10], the Global Biodiversity Information Facility (GBIF) [11], or the Ocean 37 Biogeographic Information System (OBIS) [12]. These compilations include 38 millions of opportunistic records of the distributions of very large numbers 39 of species, often across multiple decades. This temporal dimension offers 40 significant potential to track the relative commonness of species through 41 time. However, it is difficult to extract robust estimates that are insensitive 42 to changes and biases in sampling effort, from those heterogeneous and un-43 structured data sources [13]. The major issue is that it is hard to separate the 44 signal of the actual relative commonness of a species in the system from the 45 noise of sampling effort that varies in time and space, and in its taxonomic 46 focus. For instance, a species may appear common across a given decade in 47 a large dataset because there was at that time an intensive sampling pro-48 gramme targeting it. Its subsequent reduction in apparent abundance may 49 simply reflect the end of that programme, rather than anything of ecological 50 significance. 51

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In this paper, we present a method to classify the degree of commonness

of marine fish species in a certain area and time frame, using a large data 53 collection of biodiversity data. In particular, we rely on the OBIS data col-54 lection and, for the purposes of methodological development, we focus on 55 fish from the North Sea, a subset of 70 well-studied but unevenly-sampled 56 species. We use clustering analysis to automatically extract commonness 57 classes from unstructured data and compare these classes with expert opin-58 ion. Reliable concordance between our method and experts, suggests that 59 classifying commonness for less well-studied taxa or regions from data col-60 lections such as OBIS may be possible. We also assess the performance of 61 our method in terms of (i) accuracy (using cross-validation), (ii) robustness 62 to random noise in the data, (iii) dependency on the variables we chose to 63 represent species commonness and (iv) dependency on our definition of these 64 variables. 65

The paper is structured as follows: section 2 gives an overview on tech-66 niques for identifying species commonness. Section 3 describes the survey 67 data we used. Section 4 reports the variables we defined to model the prob-68 lem and describes our modelling approach. Section 5 reports an evaluation 69 of the robustness of our method. It includes a comparison between our auto-70 matic classification and the classifications produced by two experts. Section 71 6 discusses the results, suggests possible usages of our technique and includes 72 conclusive remarks. 73

#### 74 2. Overview

Species commonness and rarity have been investigated in several scientific 75 works. Most approaches derive species commonness from species abundance 76 distributions (SADs) [14, 15]. The intimate connection between abundance 77 and commonness (or rarity) is widely recognized, even if an explicit defini-78 tion of this dependency is unknown [5]. Approaches to model such depen-79 dency and to discover new correlated parameters, range from machine learn-80 ing based approaches to explicit modelling. In this last case, models specify 81 the role that each parameter has in defining species commonness. Search-82 ing for these parameters usually requires analyses by domain experts. For 83 example, Preston [16] analyses how abundance is distributed among species. 84 He recognises the importance of characteristics like (i) the total number of 85 living individuals, (ii) the total number of individuals living at any instant 86 on a given area, (iii) the ratio of the number of individuals with respect to 87 another species, (iv) the number of observed individuals in different data 88 collections. Some authors suggest that common species tend to be common 89 everywhere, as reflected in a general positive relationship between local pop-90 ulation density and regional distribution [17, 18, 19, 20]. These species also 91 tend to remain common through time [21, 22], with major changes in the 92 rank-order of species commonness rather rare. In other studies, common 93 species have been identified with species widely distributed on a territory, 94 whereas rare species have been indicated as those in the Red List for the 95 same territory. For example, using these definitions, Pearman et al. [23] 96

detect spatial patterns for common species in Switzerland. In order to account for this heterogeneity of parameters, other works have promoted using
standard measures and data to compare common and rare species [24].

Unfortunately, no single satisfactory formal definition of species common-100 ness and rarity has been found, especially using explicit modelling. Clustering 101 analysis is a promising approach coming from machine learning techniques 102 that may help to address this. This technique has been widely used for 103 identifying classes of species characteristics. For example, clustering envi-104 ronmental properties has proven to be useful in detecting vegetation types 105 [25], in modelling the coexistence of plants in agro-ecosystems [26] and in 106 detecting new agro-ecosystems [27]. Clustering analysis can also account for 107 the lack of sampling uniformity in data collections, for example to group 108 several species together when few data are available [28]. 109

# 110 3. Data

Our model needs to be trained on species observation data. In order 111 to identify the best training data, we searched for a dataset which was (i) 112 sufficiently large and complex that relative commonness was not straight-113 forward to ascertain but where (ii) the number of species was not too large 114 and (iii) independent estimates of relative commonness were available from 115 expert opinion. Points (ii) and (iii) restricted us to well-known species, with 116 officially accepted scientific names available from the authoritative World 117 Register of Marine Species (WoRMS) [29, 30]. In order to extract data, we 118

consulted the Ocean Biogeographic Information System (OBIS) [31]. OBIS is 119 the world's largest database on the diversity, distribution and abundance of 120 all marine life. OBIS was initiated in 2000 by the Census of Marine Life and 121 now runs under the auspices of UNESCO's Intergovernmental Oceanographic 122 Commission. It currently provides free access to 40 million observations of 123 115,000 marine species, integrated from more than 1,600 datasets provided by 124 nearly 500 institutions worldwide. OBIS is an amalgam of many individual 125 datasets from research projects, national monitoring programmes, museum 126 collections and so on, targeting different taxa in different areas, often using 127 different methods over different years. We limited our analysis on North Sea 128 fish, because fish (Pisces<sup>3</sup>) represents 50% of all data in OBIS and the North 129 Sea has relatively the highest amount of observations of all areas in the world. 130 Thus, we extracted observation records from OBIS and defined the spatial 131 boundaries of North Sea according to the International Hydrographic Orga-132 nization (IHO) indications. Furthermore, we selected only species observed 133 between 2000-2009, as OBIS is particularly rich of datasets and occurrence 134 records for the North Sea in this period. This selection produced a list of 135 247 scientific species names, 70 of which had distinct and accepted species 136 names according to WoRMS. We used this subset of 70 species from OBIS 137 as a benchmark to develop and evaluate our method. 138

<sup>&</sup>lt;sup>3</sup>LSID: urn:lsid:marinespecies.org:taxname:11676

#### 139 4. Method

Starting from the dataset described in the section 3, we used clustering analysis to automatically derive classes of commonness. The aim was also to search for a classification robust enough to account for sampling biases. Clustering analysis requires defining variables on the data. This section reports the steps of our analysis from the definition of these variables to the selection and application of the clustering model.

#### 146 4.1. Variables definition

The choice of the variables to use in a data mining experiment is very difficult when there is no formal definition of the phenomenon to model. Clustering analysis requires that each element to cluster is associated with a numeric vector. Thus, in our case we had to associate a vector of real numbers to each species, where the numbers were correlated with species commonness. Furthermore, such numbers had to be as independent as possible from each other. This was necessary to reduce noise during the clustering process.

The works reported in section 2, suggest that factors related to abundance and extent are correlated with species commonness. On the other hand, we know that collections of observations can contain biases. In particular, nonuniform sampling in time of the observations affects the estimation of species extents. We decided to classify the degree of commonness of each species in our benchmark dataset on the time frame of one decade (2000-2009), and to produce one classification per species for the decade. The main reason is that we wanted to explore the robustness of the classification rather than producing an analysis of commonness trends. Thus, we took into account the rate of species observations in the decade. In particular, we considered the monthly observations of the species. This rate depends also on the datasets contained in the OBIS collection. A species that is contained in several datasets (each with a different survey scope) is likely to be often encountered in that area.

This process resulted in the following variables, whose definition was guided by a cycle of interactions with domain experts. They refer only to records from the North Sea, extracted with proper geo-spatial queries:

Abundance (A): average number of reported individuals per observation. This quantity takes into account the number of individuals reported each time a species is observed:

$$A = \frac{n. of individuals reported in the record}{n. of observation records}$$

*Intra-Dataset Observations (IntraDO)*: average number of observations per dataset. These datasets come from different OBIS contributors, e.g. Fish-Base and NOAA. This parameter accounts for the frequency of presence of a species in each dataset. If the quantity is high, then the species is often reported by the OBIS contributors:

$$IntraDO = \frac{\sum_{D} n. \ of \ observations \ in \ dataset \ D}{n. \ of \ datasets \ in \ OBIS}$$

Inter-Dataset Observations (InterDO): fraction of datasets containing observation records for a species. This parameter accounts for the observation
frequency of a species among the OBIS contributors:

 $InterDO = \frac{n. of \ datasets \ with \ at \ least \ one \ observation \ for \ the \ species}{n. \ of \ datasets \ in \ OBIS}$ 

*Extension (E)*: fraction of 0.1 degree cells in the North Sea, for which at least
one observation was reported. This measure accounts for the distributional
extent of the species:

$$E = \frac{n. of \ 0.1 \ degree \ cells \ containing \ observations \ for \ the \ species \ in \ North \ Sea}{n \ of \ 0.1 \ degree \ cells \ in \ North \ Sea}$$

Time Rate (TR): fraction of months containing at least one observation
record. This measure accounts for the time rate of the species observations:

$$TR = rac{n. of months containing species observations between 2000 and 2009}{n. of months between 2000 and 2009}$$

Time Rate of Many Observations (TRMO): fraction of months containing a significant number of observations. This is an alternative measure of the observation rate, which accounts for the months in which it was frequent to observe the species. Based on the values of species known to be common or rare, we calculated that 10 observations were a significant threshold in the <sup>192</sup> 2000-2009 decade.

# $TRMO = \frac{n \text{ months containing at least 10 species observations}}{n. \text{ of months between 2000 and 2009}}$

Extracting the values of these variables from our benchmark generated 193 a set of 70 vectors of 6 Real numbers, each referring to one species between 194 2000 and 2009. The values of the variables would need to be recalculated 195 if the focus area and time range change. Applying the same calculations 196 to other data collectors than OBIS, would require finding correspondence in 197 the new collection for the elements constituting the above formulae. These 198 elements can be reconstructed from (i) geo-localized observation records, (ii) 199 the number of individuals per observation, (iii) the identity of the datasets 200 containing the observations, (iv) observation dates. Most data collectors (e.g. 201 GBIF and FishBase) support such information, which reassures us of the 202 potential generality of this approach. Nevertheless, the OBIS Postgres-based 203 collection provides very easy and fast access to retrieve the above values. 204

#### 205 4.2. Clustering

Clustering analysis is a data mining technique which is able to group together numeric vectors, according to a certain similarity criterion. In the case of real valued vectors, similarity is usually measured in terms either of density or of euclidean distances. In our case, we wanted to verify if clustering could extract classes of similarity related to species commonness. To this end, we selected two alternative clustering techniques, named X-Means [32] and DBScan [33]. The former uses a distance based approach, while the latter uses a density-based approach. We selected such algorithms because they automatically find the best number of clusters from the data.

DBScan is a density-based clustering algorithm. It searches for an optimal 215 number of clusters on the basis of two parameters: epsilon and min points. 216 The former is a distance threshold that defines the neighbourhood of a point 217 (epsilon-neighbourhood), while the latter is the minimum number of points 218 required to form a dense region. The DBSCAN algorithm starts selecting an 219 arbitrary point. Then it takes the epsilon-neighbourhood of the point and, 220 if this contains at least *min points* elements, it aggregates the points into a 221 cluster. Otherwise, it assumes that this point could be later found in the 222 epsilon-neighbourhood of another point (and thus added to the cluster of 223 that point), and moves to another point. The process analyses all the points 224 and creates density-connected clusters. For further details see Ester et al. 225 [33]. 226

X-Means is a variant of the popular K-Means algorithm [34], which intro-227 duces several efficiency enhancements. An important difference with respect 228 to K-Means is that the number of optimal clusters to search for is not speci-229 fied a priori. Instead, it requires to set a minimum and a maximum number 230 of clusters  $(K_{min} \text{ and } K_{max})$  to search for. The X-Means algorithm starts 231 from  $K_{min}$  and adds centroids as far as  $K_{max}$  is reached. At each step, the 232 K-Means algorithm is run, which finds the best assignment of the vectors to 233 the indicated number of clusters. K-Means indicates a score for this assign-234

ment, based on the distortion measure, i.e. the average squared distance of 235 the points to their clusters centroids. The X–Means algorithm outputs the 236 result of the K–Means that gave the best score, and consequently the best 237 number of clusters. X–Means also adds efficiency enhancements to K–Means. 238 using kd-trees [35] and *blacklisting* to support processing. Furthermore, at 239 each step of the computation, the location of the centroids of the additional 240 clusters is decided using the Bayesian Information Criterion (BIC) [36]. For 241 further details see Pelleg and Moore [32]. 242

We applied clustering analysis to our North Sea species benchmark. In our 243 experiment, we searched for the clustering analysis detecting the lowest num-244 ber of clusters and presenting a uniform distribution of the vectors in these 245 clusters. We used the implementations running on the D4Science Statistical 246 Manager Service [37, 38], which hosts such procedures as-a-Service. We used 247 several configurations for both the algorithms. Eventually, the best configu-248 ration for DBScan was obtained by setting epsilon = 100 and minpoints = 2. 249 Unfortunately, this ended in 38 clusters and was not practical to use. On 250 the other hand, the X-Means algorithm was executed by asking to search 251 for a number of clusters between 1 and 50. Although the interval was large, 252 the algorithm ended in only four clusters. The algorithm found an optimal 253 separation of the vectors according to their relative euclidean distance. Fur-254 thermore, we noticed that such clusters could be given an interpretation. 255 The dataset and the results are available as supplementary material of this 256 paper. 257

The normalized distribution of the mean values of the variables is re-258 ported in Table 1 for each X–Means cluster. Table 2 reports examples of 259 vectors associated to the clusters and Figure 1 displays the distribution of 260 the values of the clustering variables over the clusters. Table 3 reports the 261 interpretation we gave to these clusters, based on the distributions of their 262 centroids and of the variables values. Cluster number 1, interpreted as the 263 class of "Common" species, contains 12 vectors (corresponding to 12 species), 264 and is characterized by very high values of almost each variable. This means 265 that the species in this cluster are frequent, widespread and with high in-266 dividual density. Cluster 2 ("Moderate Commonness") contains 21 vectors 267 with lower individual density with respect to cluster 1. The most evident 268 characteristics are moderate distributional extent and moderate frequency of 269 observation. Cluster 3 ("Moderate-Low Commonness") contains 23 vectors 270 presenting a low individual density and only moderate reporting frequency 271 by several datasets. Finally, cluster 4 ("Low Commonness", which includes 272 rare species) contains 14 species which are very localized and with low indi-273 vidual density. In this case, we use the term *widespread* to indicate that the 274 species has a large geographical range, in which it is likely to be observed. 275 The term *localized* means that the species lives in highly localized zones, but 276 there could be a certain distance between such zones. Finally, individual 277 density is defined *high* if a large number of individuals are encountered each 278 time the species is observed. 279

#### 280 5. Evaluation

#### 281 5.1. Agreement with experts

In this section, we evaluate the performance of the classification produced 282 by X–Means with respect to expert opinion. In order to create a comparison 283 reference, two of us (Bailly and Cattrijsse) performed independent classifi-284 cation assignments on the 70 benchmark species of North Sea fish, based 285 on expert opinion. Each expert separately assigned the appropriate cluster 286 to each species, selecting among those in Table 3. The experts did not be-287 long to the same institute: Expert 1 (Cattrijsse) is a researcher in Coastal 288 Marine Biology working for the Vlaams Instituut voor de Zee (VLIZ), while 289 Expert 2 (Bailly) is a biologist working in the biodiversity informatics field 290 for the World Fish Center. The result of this classification is available as 291 supplementary material attached to this paper. 292

We estimated the agreement between all the classifications using the ab-293 solute percentage of agreement, defined as the percentage of matching assign-294 ments. Furthermore, we also calculated Cohen's Kappa [39], which estimates 295 the agreement between two evaluators with respect to purely random assign-296 ments. Cohen's Kappa allows comparing complex classification tasks (e.g. 297 with many classes) with simpler ones (e.g. dichotomous scenarios) where 298 high agreement could have occurred by chance. Table 4 reports the Cohen's 299 Kappa values of the agreements, along with two different interpretations 300 commonly used in literature [40, 41]. It is notable that in this experiment 301 the absolute percentage agreement reflects the Kappa values. The values are 302

<sup>303</sup> symmetric, thus we report them once per pair of evaluators.

In order to give insight about the differences between the classifications 304 assignments, we report the example of the lesser pipefish Syngnathus rostel-305 latus<sup>4</sup>, which Expert 2 and X-Means assign to Moderate-Commonness, and 306 Expert 1 to Common. This species presents an Abundance (A) parameter 307 value equal to 17.16, quite far from the 325.27 of the common dab Limanda 308  $limanda^5$ , which is "Common" according to all the assignments. A signifi-309 cant difference is recorded also for the IntraDO values, which is 101.75 for 310 the lesser pipefish and 24521.14 for the common dab. Indeed, Syngnathus 311 rostellatus has a lower number of observation records for (407 records) with 312 respect to *Limanda limanda* (171648 records). This influences the behaviour 313 of X–Means, but its classification can be still considered viable because it 314 agrees with one of the two experts. Figure 2 depicts the distribution of the 315 observation records of the above species, aggregated at 0.5 degrees resolution. 316 One interesting consideration is that, even if the classification classes were 317 automatically detected by the X-Means algorithm, the overall agreement 318 with both the experts is good. On the other hand, the agreement between 319 the two experts is poor. This indicates that the problem is objectively hard. 320 but clustering seems able to reconcile the divergent opinions in some way. 321

The disagreement between experts could be due to their different interpretation of the clusters descriptions. Thus, we investigated this aspect by

<sup>&</sup>lt;sup>4</sup>LSID: urn:lsid:marinespecies.org:taxname:127389

<sup>&</sup>lt;sup>5</sup>LSID: urn:lsid:marinespecies.org:taxname:127139

aggregating the not Common clusters into a generic Non-Common cluster.
Table 5 reports the evaluation in this case. The agreement between Expert
2 and clustering is excellent, while the aggregation introduces misalignment
between Expert 1 and clustering. This is due to a general tendency by Expert
1 to classify more in the Moderate-Commonness class.

We repeated the same evaluation aggregating the *Common* and the *Moderate*-329 Commonness clusters into one cluster, and the Moderate-Low and Low-330 *Commonness* clusters into another cluster. Table 6 reports the agreement 331 in this case. With this aggregation, the agreement by both the experts with 332 the clustering analysis is good, and highest agreement is still with Expert 2. 333 These experiments highlight that even changing the definition of the clus-334 ters, there is a sensible agreement between experts and clustering. This 335 indicates reliability of the automatic classification. It is notable that the 336 variables used by the clustering analysis are likely to be affected by biases, 337 especially when the species is poorly reported in time and is rarely reported 338 by the OBIS contributors. Clustering accounts for the lack of information of 339 some variables, because it compensates with information from the other vari-340 ables. This comes out from the variables combination made by the euclidean 341 distances and by the subsequent optimization process. Furthermore, produc-342 ing classes of commonness (instead of commonness scores) hides fine-grain 343 differences between the vectors. 344

#### 345 5.2. Performance evaluation

We measured the robustness of our method in terms of (i) classifying new 346 species, (ii) dependency on noise, (iii) dependency on the clustering variables 347 and (iv) on their definitions. In particular, we calculated the performance on 348 classifying species that were not included in the training set. To this aim, we 349 used cross-validation. We randomly selected 90% of the species to produce 350 clusters. We checked if the clusters coincided with the ones extracted using 351 100% of the species (complete set), and then we used the other 10% of the 352 species to check if their associated vectors were assigned to the same clusters 353 as in the complete set. We used only 10% of the species as test set because 354 our benchmark dataset had small size. In each experiment, we calculated the 355 accuracy of the classification as the ratio between correct assignments and 356 overall assignments. In the end, we averaged the accuracies of ten executions. 357 In all the experiments the clusters coincided with the ones of the complete set. 358 The overall (averaged) accuracy was 98.57%. This means that for the North 359 Sea case our clusters are stable and the model is promising in classifying new 360 species. 361

As further step, we checked the robustness of our classification to noise. As explained before, the data we extracted from OBIS contain sampling biases. The good agreement of our method with expert opinion already suggests that our approach can manage these biases. Nevertheless, we explored this aspect further by adding an increasing amount of white noise to our data and checking if the clusters remained stable, i.e. if the newly identified

clusters were still the ones of Table 3. We added white noise directly to our 368 variables and Table 7 reports the results: a 10% noise level means that we 369 randomly added or subtracted up to the 10% of a variable value. Referring 370 to Table 7, up to 1% of noise there is no change in the clustering and even 371 at 5% the clusters are very similar to the ones without noise, because most 372 of the species in the original ("clean" data) clusters are found in the corre-373 sponding newly found clusters. The number of clusters changes when 10%374 of noise is reached, but at this level the newly found clusters have still corre-375 spondence with the original clusters. For example, the species belonging to 376 the original cluster 1 are largely included in the newly found cluster 1. The 377 original cluster 2 corresponds to both the new clusters 1 and 2, whereas the 378 original cluster 3 and 4 correspond to the new clusters 2 and 3 respectively. 379 Over 10% of noise the original clusters are no more recognizable. It is our 380 opinion that this limit is a reasonable indicator of robustness to noise. It 381 is remarkable, in fact, that our data are already biased and the white noise 382 only adds more bias. 383

As additional step, we evaluated the influence of each variable on the clustering analysis. Table 8 reports the results of the clustering analysis when we exclude one variable at time. The number of clusters changes and the identity of the original clusters is lost in most of the cases. It is notable that when *InterDO* is missing, the number of clusters is overestimated. In the other cases, the clustering is very simplistic and does not allow easy semantic interpretations. In particular, clusters 1, 3 and 4 are merged together, which means that common and uncommon species are mixed up. These changes indicate that all the variables have an important role (i.e. carry a remarkable amount of information) in the definition of the clusters of Table 3. Our definitions are related to indicators taken from other studies and come from expert opinion (see section 4.1). This analysis confirms that they all have a key role in producing species commonness classes that agree with expert opinion.

As final step, we checked if the commonness classes depend on our defini-398 tions of the variables (see section 4.1). Table 9 reports how the results of the 399 clustering analysis change when the variables definitions are slightly altered. 400 The new definitions in Table 9 still include information that is correlated to 401 the original definitions. For example, in one of the experiments we redefined 402 A as the number of recorded individuals, without dividing for the number of 403 observations. In another case, we defined one time variable as the ratio be-404 tween the two time variables TRMO and TR. The last row of Table 9 reports 405 the case in which all the variables definitions are altered. In all the cases, the 406 clustering analysis identifies four clusters. Furthermore, the original clusters 407 are recognizable in all the cases and sometimes the output coincides with the 408 one of the original model. This means that the clustering analysis is flexible 409 enough to exploit the information associated to the variables, even when the 410 variables definitions change. 411

#### 412 6. Discussion and conclusions

In this paper we have presented an approach to classify species common-413 ness. We have trained our models on a dataset extracted from the OBIS data 414 collection and focusing on North Sea fishes. The performance has been eval-415 uated by comparing automatic assessments with the opinions of two experts. 416 We have demonstrated that our process has good agreement with expert 417 opinion although our analysed dataset contains sampling biases. We have 418 further explored this robustness, by evaluating the effects that random noise 419 in the data has on the classification. The results indicate that the model 420 is reasonably robust in managing noise. Furthermore, we have used cross-421 validation to calculate the performance of our model in classifying species 422 that had not been included in the training set. The performance indicates 423 that the identified clusters are stable for the North Sea species. This gives 424 suggestions about the possible generalisation of our method. In fact, our 425 clustering analysis is also applicable to other areas and large biodiversity 426 data collections. Applying our method to other regions than North Sea re-427 quires the model to be trained on new data. Indeed, we conducted the same 428 analysis on 222 species from OBIS at global scale. Also in this case, we 429 found an optimal separation into four clusters<sup>6</sup> having the same percentage 430 distributions as in Table 1. This result indicates that our classification could 43 be valid for other areas too, but validating this hypothesis requires further 432

 $<sup>^{6}{\</sup>rm The}$  complete classification is available on the D4S cience e-Infrastructure for consultation: http://goo.gl/TYuD6P

<sup>433</sup> investigation and much more effort in terms of experts' reviews. We will
<sup>434</sup> address this issue in future experiments.

We have demonstrated that our process is more dependent on the information included in the variables than to their definition. This is useful when applying our analysis to other biodiversity data collections that report information in a different way from OBIS.

Finally, we have demonstrated also that our set of variables contains a 439 sufficient amount of information to identify four reliable commonness clas-440 sifications. Using a lower number of variables would produce less refined 441 classifications and less clusters (see Table 8). This is a remarkable property, 442 since we defined the variables based on interactions with ecology and data 443 experts (i.e. not using automatic data selection [42]). This may suggest that 444 our variables are ecologically meaningful, i.e. they are really correlated to 445 species commonness. 446

From our analysis, new biodiversity and ecosystem indicators could be 447 identified and this will be part of our future investigations. For example, 448 using our method a species could be found, today, to be "less common" in 449 a certain area with respect to a previous time period. This could indicate a 450 change of the ecosystem in that area or that the species has been overfished. 451 Our method could be also a way to reconcile the opinions of different experts 452 about the commonness of a set of species. For example, it could be used as a 453 supporting tool for biologists, who would rely on an "external" opinion when 454 discussing about species commonness. Furthermore, classifying commonness 455

for fishes in a well-studied region is a first step towards working on less known
taxa in other regions.

Our experiments highlight the intrinsic difficulty of the problem, but the 458 proposed technique represents a step forward in classifying species common-459 ness and in understanding which factors are related to this concept. A data 460 provider like OBIS could embed such method to alert a user about the pos-461 sible commonness of a species in a certain area. In this context, we are 462 planning to build an interface allowing a user to select an IHO area and 463 a time rage, and to retrieve the species possibly classified as *Common* or 464 Moderately-Common. Currently, our clustering technique is released as soft-465 ware [43, 44] inside the i-Marine e-infrastructure [45], which grants free access 466 to statistics about the OBIS database and allows sharing datasets, biological 467 analyses and experimental results. 468

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	А	IntraDO	InterDO	Е	TR	TRMO
Cluster 1	85.3%	85.4%	33.9%	64.3%	35.4%	47.1%
Cluster 2	9.5%	12.4%	26.6%	26.4%	31.5%	37.5%
Cluster 3	4.8%	2.1%	21.4%	8.3%	23.4%	14.7%
Cluster 4	0.4%	0.1%	18.1%	1.0%	9.6%	0.6%

Table 1: Normalized distributions of the mean values of the variables in the X–Means clusters.

Sp. scientific name	А	IntraDO	InterDO	Е	TR	TRMO	Cluster
Sprattus sprattus	7921.81	2779.67	0.44	0.031	0.44	0.39	1
Trisopterus esmarkii	5477.46	2502.11	0.44	0.027	0.45	0.44	1
Gadus aeglefinus	1680.20	8869.78	0.67	0.039	0.49	0.48	1
Trachurus trachurus	2067.49	1294.33	0.56	0.035	0.45	0.42	2
Pollachius virens	250.39	1433	0.44	0.013	0.43	0.37	2
Platichthys flesus	11.02	647.89	0.56	0.013	0.59	0.5	2
Ammodytes lancea	663.20	49.22	0.67	0.0036	0.26	0.1	3
Mustelus asterias	16.52	96.89	0.33	0.0046	0.38	0.21	3
Scophthalmus rhombus	2.58	82.33	0.56	0.010	0.4	0.17	3
$Pomatoschistus\ pictus$	38.17	2.67	0.33	0.00032	0.083	0	4
$Ciliata\ septentrional is$	5.75	6.22	0.33	0.00076	0.1	0.0083	4
Labrus bergylta	0.07	6.56	0.33	0.00044	0.13	0.017	4

Table 2: Examples of vectors of parameters (with related clusters) for some of the species included in our benchmark dataset.

Cluster		
Number	$\operatorname{Label}$	Definition
		Frequent,
Cluster $1$	Common	$\operatorname{widespread},$
		high individual density
		Moderately frequent,
Cluster $2$	Moderate Commonness	moderately widespread,
		medium individual density
		Poorly widespread,
Cluster 3	Moderate-Low Commonness	poorly-moderately frequent,
		low individual density
		Localized,
Cluster 4	Low Commonness	not frequent,
		very low individual density

Table 3: Interpretation of the X–Means clusters as classes of species commonness.

Kappa values on 4 Clusters					
	Expert 2	Clustering			
Expert 1	0.24	0.57			
Expert 2		0.48			
Kappa ir	nterpretation	Fleiss/Landis–Koch			
	Expert 2	Clustering			
Expert 1	Poor/Slight	Good/Moderate			
Expert 2		$\operatorname{Good}/\operatorname{Moderate}$			
Abso	lute Percenta	age of Agreement			
	Expert 2	Clustering			
Expert 1	46.5%	67.4%			
Expert 2		61.4%			

Table 4: Agreement with Kappa statistic and absolute percentage of agreement on the classification of species in four clusters: *Common, Moderate-Commonness, Moderate-Low Commonness, Low-Commonness.* The table in the middle reports interpretations for the Kappa values.

Kappa values on Comm./Non-Comm. classes					
	Expert 2	Clustering			
Expert 1	0.34	0.39			
Expert 2		0.78			
Kappa	interpretation	${ m Fleiss}/{ m Landis}-{ m Koch}$			
	Expert 2	Clustering			
Expert 1	Marginal/Fair	Marginal/Fair			
Expert 2		$\mathbf{Excellent}/$			
		Substantial			
Abs	solute Percenta	age of Agreement			
	Expert 2	Clustering			
Expert 1	67.4%	69.8%			
Expert 2		92.9%			

Table 5: Agreement with Kappa statistic and absolute percentage of agreement on the classification of species in two clusters: *Common, Non–Common.* The table in the middle reports interpretations for the Kappa values.

Kappa values on 2 aggregated Clusters						
	Expert 2	Clustering				
Expert 1	0.26	0.67				
Expert 2		0.52				
Kappa ii	nterpretation <b>H</b>	leiss/Landis–Koch				
	Expert 2	Clustering				
Expert 1	Marginal/Fair	Good/Substantial				
Expert 2		$\operatorname{Good}/\operatorname{Moderate}$				
Abso	lute Percentag	e of Agreement				
	Expert 2	Clustering				
Expert 1	67.4%	83.7%				
Expert 2		75.7%				

Table 6: Agreement with Kappa statistic and absolute percentage of agreement on the classification of species in two aggregated clusters: *Common and Moderate-Common vs. Moderate-Low and Low-Commonness.* The table in the middle reports interpretations for the Kappa values.

Response to Noise								
	Distribution of the original clusters							
		on the newly found clusters						
	Found							
Added noise	Clusters	Cluster 1	Cluster 2	Cluster 3	Cluster 4			
	(C1, C2,,Cn)							
		100% C1	0% C1	0% C1	0% C1			
0.1%	4	0% C2	100% C2	0% C2	0% C2			
0.170	4	0% C3	0% C3	100% C3	0% C3			
		$\begin{array}{c ccccccccccccccccccccccccccccccccccc$	0% C4	100% C4				
	4	100% C1	0% C1	0% C1	0% C1			
1 0%		0% C2	100% C2	0% C2	0% C2			
170		0% C3	0% C3	100% C3	0% C3			
		0% C4	0% C4	0% C4	$100\%~{\rm C4}$			
		100% C1	4% C1	0% C1	0% C1			
50%	4	0% C2	96% C2	0% C2	0% C2			
570	4	0% C3	0% C3	91% C3	0% C3			
		0% C4	0% C4	9% C4	$100\%~{\rm C4}$			
		70% C1	43% C1	17% C1	0% C1			
10%	3	30% C2	$48\%~{\rm C2}$	66% C2	14% C2			
		0% C3	9% C3	18% C3	86% C3			
50%	1	100% C1	100% C1	100% C1	100% C1			

Table 7: Output of our clustering analysis in response to random noise added to the data. The results are reported with respect to an increasing percentage of added noise. The percentages indicate the distribution of the clusters associated to the clean data over the clusters found for the noisy data.

Variables influence on the clustering analysis							
	Distribution of the original clusters						
		on the	e newly four	nd clusters			
Excluded	Found						
variable	Clusters	Cluster 1	Cluster 2	Cluster 3	Cluster 4		
variable	(C1, C2,,Cn)						
Δ	9	100% C1	78% C1	100% C1	100% C1		
Π	2	0% C2	22% C2	0% C2	0% C2		
IntroDO	2	100% C1	78% C1	100% C1	100% C1		
IntraDO		0% C2	22% C2	0% C2	0% C2		
	5	100% C1	13% C1	0% C1	0% C1		
		0% C2	87% C2	0% C2	0% C2		
InterDO		0% C3	$0\% \ \mathrm{C3}$	61% C3	$0\% \ \mathrm{C3}$		
		0% C4	$0\% { m C4}$	39% C4	29% C4		
		0% C5	$0\% \mathrm{C5}$	$0\% \mathrm{C5}$	71% C5		
Е	1	100% C1	100% C1	100% C1	100% C1		
TP	n	100% C1	70% C1	100% C1	100% C1		
110	Δ	0% C2	30% C2	0% C2	0% C2		
TRMO	9	100% C1	30% C1	100% C1	100% C1		
INMO	2	0% C2	70% C2	0% C2	0% C2		

Table 8: Modifications in the species clustering when one variable at time is excluded. The percentages indicate the distribution of the original clusters over the newly calculated clusters.

Influence of variables redefinitions on the clustering analysis							
		Distribution of the original clusters					
		on the	e newly four	nd clusters			
Dedefined	Found						
	Clusters	Cluster 1	Cluster 2	Cluster 3	Cluster 4		
variable	(C1, C2,,Cn)						
		100% C1	0% C1	0% C1	0% C1		
	4	0% C2	100% C2	0% C2	0% C2		
A = n. of individuals	4	0% C3	0% C3	100% C3	0% C3		
		0% C4	0% C4	0% C4	100% C4		
		100% C1	0% C1	0% C1	0% C1		
<i>All</i> C 1	4	0% C2	96% C2	0% C2	0% C2		
$A^{\prime\prime} =$ n. of obs.	4	0% C3	4% C3	91% C3	0% C3		
		0% C4	0% C4	9% C4	100% C4		
		100% C1	9% C1	0% C1	0% C1		
IntraDO' = avg. n. of obs.	4	0% C2	91% C2	0% C2	0% C2		
in datasets containing	4	0% C3	0% C3	100% C3	0% C3		
species obs.		0% C4	0% C4	0% C4	100% C4		
		100% C1	0% C1	0% C1	0% C1		
InterDO' = n. of datasets		0% C2	100% C2	0% C2	0% C2		
containing species obs.	4	0% C3	0% C3	100% C3	0% C3		
		0% C4	0% C4	0% C4	100% C4		
		100% C1	30% C1	0% C1	0% C1		
		0% C2	70% C2	40% C2	0% C2		
TR' = n. of months with obs.	4	0% C3	0% C3	60% C3	0% C3		
		0% C4	0% C4	0% C4	100% C4		
		100% C1	35% C1	0% C1	0% C1		
TRMO' = n. of months	4	0% C2	65% C2	0% C2	0% C2		
with at least $10$ obs.	4	0% C3	0% C3	100% C3	0% C3		
		0% C4	0% C4	0% C4	100% C4		
		100% C1	30% C1	0% C1	0% C1		
T=TRMO/TR	4	0% C2	70% C2	0% C2	0% C2		
(subst. to TR and TRMO)	4	0% C3	0% C3	61% C3	0% C3		
· · · · · · · · · · · · · · · · · · ·		0% C4	0% C4	39% C4	100% C4		
		100% C1	8% C1	0% C1	0% C1		
$A^{\prime}, IntraDO^{\prime},$	4	0% C2	92% C2	0% C2	0% C2		
InterDO', $TR'$ ,	4	0% C3	0% C3	100% C3	0% C3		
TRMO'		0% C4	0% C4	0% C4	100% C4		

Table 9: Modifications in the species clustering when variables are redefined in a slightly different way from our default definitions. The percentages indicate the distribution of the original clusters over the newly calculated clusters.



Figure 1: Distribution of the values of our variables over the four clusters identified by our model.



Figure 2: a. Representation of observation records from OBIS for *Syngnathus rostellatus*, aggregated at 0.5 degrees b. Representation of observation records from OBIS for *Limanda limanda*, aggregated at 0.5 degrees.