



## Impact of airborne iron oxide nanoparticles on *Tillandsia usneoides* as a model plant to assess pollution in heavy traffic areas

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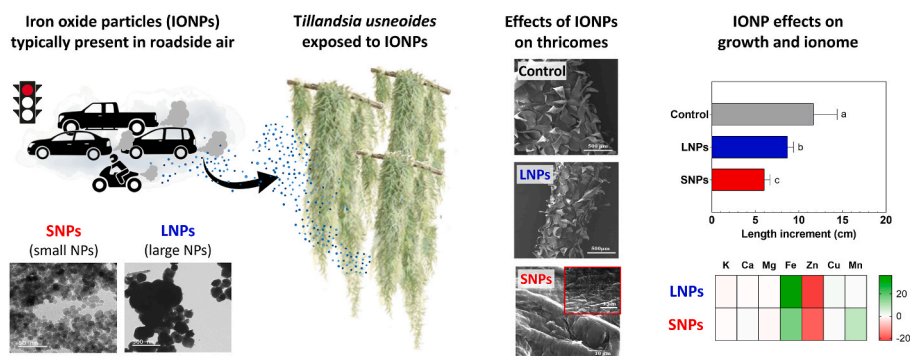
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### HIGHLIGHTS

- Toxicity of airborne magnetite nanoparticles was tested on *Tillandsia usneoides*.
- Biometry and ionome, but not photosynthesis, were altered by the treatment.
- Small nanoparticles were localised on the trichome wings.
- Small nanoparticles were more toxic than large ones for *T. usneoides*.

### GRAPHICAL ABSTRACT



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### ABSTRACT

Due to the increasing evidence of widespread sub-micron pollutants in the atmosphere, the impact of airborne nanoparticles is a subject of great relevance. In particular, the smallest particles are considered the most active and dangerous, having a higher surface/volume ratio. Here we tested the effect of iron oxide ( $\text{Fe}_3\text{O}_4$ ) nanoparticles (IONPs) with different mean diameter and size distribution on the model plant *Tillandsia usneoides*.

Strands were placed in home-built closed boxes and exposed to levels of airborne IONPs reported for the roadside air, i.e. in the order of  $10^7 - 10^8$  items  $\text{m}^{-2}$ . Plant growth and other morpho-physiological parameters were monitored for two weeks, showing that exposure to IONPs significantly reduced the length increment of the treated strands with respect to controls. A dose-dependence of this impairing effect was found only for particles with mean size of a few tens of nanometers. These were also proved to be the most toxic at the highest concentration tested. The IONP-induced hamper in growth was correlated with altered concentration of macro- and micronutrients in the plant, while no significant variation in photosynthetic activity was detected in treated

**Abbreviations:** IONPs, iron oxide nanoparticles; LNPs, large nanoparticles; SNPs, small nanoparticles; LC, low concentration; HC, high concentration.

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samples. Microscopy investigation showed that IONPs could adhere to the plant surface and were preferentially located on the trichome wings.

Our results report, for the first time, evidence of the negative effects of airborne IONP pollution on plant health, thus raising concerns about related environmental risks.

Future research should be devoted to other plant species and pollutants to assess the impact of airborne pollution on plants and devise suitable attenuation practices.

## 1. Introduction

Iron oxide particles (IONPs) in the form of magnetite ( $\text{Fe}_3\text{O}_4$ ) are found as finely dispersed pollutants in the atmosphere (Li et al., 2021), especially in regions where anthropogenic activities are intensive (Zhang et al., 2020). These particles are mostly originated in dense traffic areas and industrial districts, such as cement factories, power industries and exhausted vehicle deposits. Air pollution caused by IONPs has been reported as a major problem either in open spaces or in the lab (Abdul-Razzaq and Gautam, 2001; Verma et al., 2016), thus attracting the attention of public institutions and the scientific community. As other polluting particles (i.e., nano- and microplastics) (Kumar et al., 2021), IONPs can be taken up and internalized by living organisms (Oberdörster et al., 2004). In humans, they are accumulated in various tissues, such as the brain and the pleura, or are absorbed by the serum (Singh and Sanjeeb, 2014). In particular, it has been shown that, when the diameter of these particles is less than approximately 200 nm (Wohlfart et al., 2012; Maher et al., 2016), they can cause severe neurodegenerative diseases since they are able to cross the blood-brain barrier through the olfactory neural pathway (Hamdy et al., 2022a). Reports indicate the impact of airborne IONP pollution as negative and harmful for humans (Zhu et al., 2019) and animals (Kaloyianni et al., 2020), though plant-based green synthesis of IONPs with reduced toxicity has been proposed for their medical applications (Hamdy et al., 2022b, 2022c). In any case, their effect on plants has not been investigated yet, albeit these organisms have a universally acknowledged role as primary producers and climate regulators (Fankhauser et al., 2022; Matzka and Maher, 1999). Actually, leaves are among the principal targets for airborne pollution agents, since their large (smooth or faceted) surfaces can absorb many contaminants, either in the form of molecules or particulates (Omasa et al., 2002), including magnetic particles derived from vehicle emission (Prajapati et al., 2006). As for other environmental matrices, such as soil and water, information about the relationship between IONPs and plants is found with rather variable outcomes, according to administration modality and particle concentration. Until now, studies on soil have been mainly focused on plant nutrition, possibly overlooking the problem of toxicity by IONP high levels, even though these particles can be found at considerable concentrations in some backgrounds, such as roadside terrains (Liu et al., 2022). On the other hand, researchers dealing with Fe deficiency and low availability in plants (Vose, 1982) have proposed the administration of magnetite nanoparticles as a strategy for promoting plant growth in nano-enabled agriculture (Singh et al., 2021). According to some studies, this strategy would also contribute to decrease environmental pollution caused by classical fertilizers (Dola et al., 2022; Ahmed et al., 2023). Concerning water environments, IONP pollution is reported to have a negative impact on plant growth and development (Jafarirad et al., 2019; Giorgetti et al., 2011; Li et al., 2015). For example, the excess of IONPs in hydroponics has been shown to reduce plant biomass production, through induction of oxidative stress and decrease of photosynthetic pigment contents (Wang et al., 2011; Li et al., 2020). Conversely, low concentrations of Fe-containing particles in hydroponics and their administration as foliar spray are reported to induce positive effects on plants (Tombuloglu et al., 2019; Iannone et al., 2016) included the reduction of heavy metal toxicity (Emamveridian et al., 2023). In any case, IONPs used in controlled conditions have been proposed for wastewater treatment (Yadav et al., 2020; Rajendran et al.,

2023).

Due to the assessed presence of IONPs in the atmosphere and their predictable increasing levels in the future, these particles are to be considered a possible environmental threat. Investigating whether airborne IONP contamination can affect plant growth and development becomes of fundamental importance for the scientific community in view of actions to preserve the overall functioning of the biosphere. Here, the problem of airborne magnetite nanoparticle pollution on plants was addressed by using the model species *Tillandsia usneoides* (L.) L. The neotropical genus *Tillandsia* offers the possibility to estimate contaminant levels in the atmosphere, as these plants take up water and nutrients directly from the air due to the lack of a root apparatus (Mosti et al., 2008). *Tillandsia* species present leaf trichomes with an external wing of dead cells and a stem able to absorb water and nutrients (Papini et al., 2010), thus making this plant a valuable tool for biomonitoring atmospheric pollutants (Brighigna et al., 2002). Moreover, *Tillandsia usneoides* is present in a wide range of environments, from Florida to Argentina (Sun et al., 2021), and has already been used as a sensor of airborne pollution (Schreck et al., 2020; Falsini et al., 2022).

Two types of magnetite nanoparticles with different mean size and polydispersity were chosen as models of IONP pollution and were tested on *T. usneoides* cultivated in controlled conditions. This choice stems from the evidence that smaller particles have a higher impact due to their extended surface on equal conditions of chemical composition and mass. On the other hand, particles with wide size distribution represent a realistic model to study pollution by airborne finely dispersed materials which can be generated by different fractioning processes, thus resulting in inhomogeneous size and shape. The aims of this work were: i) assessing if *T. usneoides* is sensible to realistic concentrations of airborne IONPs by evaluating changes in growth, chlorophyll fluorescence and element concentration, ii) investigating possible IONP adhesion to the plant surface by SEM microscopy, iii) evaluating whether IONP dimension can differently affect their toxicity to our plant system.

This report is the first to study the consequences of airborne magnetite pollution on plants and opens new questions on the already serious issue that the excess of nanoparticles can impact living organisms, claiming awareness for its control and reduction.

## 2. Materials and methods

### 2.1. Magnetic nanoparticles, plant material and experimental conditions

Magnetite bare nanoparticles were purchased in powder by Merck and Co., Inc (CAS N° 1317-61-9). Interestingly, these particles have a size distribution that closely resemble the one occurring at a short distance, i.e. 20 m, from a roadside soil (Prajapati et al., 2006) and were considered as large nanoparticles (LNPs). Small magnetite nanoparticles (SNPs) were synthesized by the well-known Massart process by alkaline co-precipitation of  $\text{Fe}^{3+}$  and  $\text{Fe}^{2+}$  precursors (Martens et al., 2019; Bertuit et al., 2022). Briefly, 0.9 mol of  $\text{FeCl}_2 \cdot 4\text{H}_2\text{O}$  and 1.6 mol of  $\text{FeCl}_3 \cdot 6\text{H}_2\text{O}$  were co-precipitated by 7 mol of alkaline ammonium solution. After magnetic decantation, the obtained magnetic particles were isolated and washed several times with nitric acid, acetone, and diethyl ether to remove excess precursors and by-products. At the end of this process, the SNPs were re-suspended in water. SNPs and LNPs were diluted in Milli-Q water at a final iron concentration of about  $0.1 \text{ g mL}^{-1}$  and a drop of the solutions was deposited on carbon-coated grids and

imaged by transmission electron microscopy using a JEOL-1011 at 100 kV. The size distribution histograms of the two different sizes of nanoparticles were analysed from the TEM images using ImageJ software (<https://imagej.nih.gov/>). The histograms were fitted by a LogNormal function.

To administer IONPs to plants, a home-built setup was used as described by Falsini et al. (2022). In each box, nine *T. usneoides* strands with a length of approximately 30 cm were hung to the lid by the hooks. The air circulation was guaranteed by two fans located at the opposite side of the box, directed toward the bottom to create a turbulent motion. These fans were programmed to switch on during the night for a total time of 2h when CAM species like *T. usneoides* open the stomata (Martin and Sedow, 1981). Two beakers filled with demineralized water were placed in the box to maintain humidity; the exchange of air with the external environment was promoted by two circular openings present on the opposite side of the box. Finally, all the boxes were placed in a growth chamber with the following conditions: 24/16 °C day/night, light intensity 300  $\mu\text{mol m}^{-2} \text{s}^{-1}$ , 16-h (day) photoperiod and relative humidity 60–65%.

*Tillandsia usneoides* plants were purchased at a local nursery (Giulio Celandroni Orchidee, San Giuliano Terme, Pisa) and maintained in the growth chamber with the conditions indicated above. Single strand length was measured and samples of homogeneous size and with a similar number of ramifications, not more than 5, were chosen for the trial. The length measurements were performed by analysing the images of the plants with ImageJ, adding all the ramifications length. Thus, the cumulative length of the plants was calculated. The stem diameter was measured as well and resulted constant.

*Tillandsia usneoides* strands were exposed to IONPs having size and concentrations typically present in polluted air environments (Magiera et al., 2011), specifically, in the roadside air i.e. in the order of  $4 \times 10^7$  items  $\text{m}^{-2}$  and  $4 \times 10^8$  items  $\text{m}^{-2}$  (Maher et al., 2016). The calculation of the IONPs amount to be allocated in each prototype chamber was based on the volume of the particles, considering that particles showed a spherical shape of known mean diameter, whose value was extracted as the mean of the corresponding size distribution. The highest polluted atmosphere was replicated in a box of 18  $\text{dm}^3$  by administering a total of  $3.1 \times 10^{-3} \mu\text{g}$  and  $1.04 \times 10^{-4} \mu\text{g}$  for LNPs and SNPs, respectively while the lowest polluted air was obtained by resuspending  $3.1 \times 10^{-4} \mu\text{g}$  and  $1.04 \times 10^{-5} \mu\text{g}$  for LNPs and SNPs respectively. The acronyms of the different samples are listed in Table 1.

To obtain these amounts of particles a series of dilutions was operated starting from a stock solution of 2  $\text{g mL}^{-1}$ . Droplets of 31  $\mu\text{L}$  of both types of NPs were placed on a plastic disk under fans before switching on. The fans, being directed downward, were able to induce a circular movement of the IONPs into the box after drying the droplets. During exposure, each *T. usneoides* strand was hydrated with 5 mL of sprayed tap water three times a week. In this same period, increment in plant growth was calculated by subtracting length values at the beginning of the treatment to values at 3, 7, 11 and 14 days of treatment. At each step in which the chambers were opened to make measurements, IONPs were newly administered to maintain the initial amount of particles.

**Table 1**

List of acronyms referring to the different samples used in *T. usneoides* treatments where H and L mean high and low, respectively.

Sample Name	Sample acronym
Control	CNT
LNPs $4 \times 10^7$ items/ $\text{m}^3$ o $1. \times 10^{-4} \mu\text{g}$	LC-LNPs
LNPs $4 \times 10^8$ items/ $\text{m}^3$ o $3 \times 10^{-3} \mu\text{g}$	HC-LNPs
SNPs $4 \times 10^7$ items/ $\text{m}^3$ o $1 \times 10^{-4} \mu\text{g}$	LC-SNPs
SNPs $4 \times 10^8$ items/ $\text{m}^3$ o $3 \times 10^{-3} \mu\text{g}$	HC-SNPs

LC = Low Concentration; LNPs = Large Nanoparticles; HC= High Concentration; SNPs = Small Nanoparticles.

## 2.2. Fluorescence parameters

Chlorophyll fluorescence was measured using a portable fluorimeter (Plant Efficiency Analyzer – Handy PEA, Hansatech Instruments Ltd) on 15-min dark-adapted leaf samples flashed for 1 s with a saturated ( $1800 \mu\text{mol m}^{-2} \text{s}^{-1}$ ) LED red light pulse (650 nm). The potential quantum efficiency of photosystem II Fv/Fm (Fv if the variable fluorescence and indicates the difference between the maximal Fm and the minimal F0 fluorescence) and the general indicator of photosystem I and II efficiency Pindex (Performance index) were recorded.

## 2.3. Element determination

After the treatment, strands were dried at 70 °C for two days for the measurement of the concentration of micro- and macro-elements. 100 mg of oven-dried plant material was digested using 10 mL of 69%  $\text{HNO}_3$  in a microwave digestion system (Mars 6, CEM) with a maximum temperature of 200 °C for 10 min (Bettarini et al., 2019). The amount of K, Ca, Mg, Fe, Zn, Cu and Mn was determined by atomic absorption spectroscopy (PinAAcle 500, PerkinElmer) with certified reference materials (grade BCR, Fluka Analytical, Sigma-Aldrich, method reliability and accuracy with values < 10% and <5% RDS respectively).

## 2.4. Scanning electron microscopy (SEM)

Scanning electron microscopy (SEM) measurements were performed at the CEME-Centro di Microscopia Elettronica “Laura Bonzi”, CNR Research Area (Florence, Italy). A Gaia 3 (Tescan s.r.o, Brno, Czech Republic) focused ion beam-scanning electron microscope electron beam used for SEM imaging had a voltage of 10 kV and was operating in a high vacuum and with a secondary electron detector. Gaia 3 was equipped with an EDS-X-ray microanalysis system (EDAX, AMETEK, USA) TEAM EDS Basic Software Suite TEAM™.

Samples were deposited on a stub, dried in a vacuum and then coated with an ultrathin coating of silver to enhance the contrast thanks to the presence of an electrically conducting material. Microanalysis was performed to identify the presence of nanoparticles on leaf surfaces.

To assess the dimension of trichomes from IONP-exposed samples, for each SEM image three trichomes with the whole surface horizontally visible were selected. Length and width of the wing and diameter of the shield were measured, and values were pooled together to obtain a mean value for each image. Three SEM images were analysed for each plant to obtain a mean value to be considered as the final value of the replicate.

## 2.5. Statistics

The experiment comprised five groups of plants (control plants and plants treated with the two kinds of IONPs at low and high concentrations) and for each group there were nine biological replicates. When measurements were performed over time (plant length increment and chlorophyll fluorescence), the same plant sample was investigated at different times. Linear regression was used to analyse experimental data of plant growth (GraphPad Prism 7, GraphPad Software, San Diego, CA), where the dependent variable was plant length increment while the independent one was sampling time. The angular coefficient (slope) of the linear regression was considered as a descriptor of the plant growth rate (Falsini et al., 2022). The significance of differences ( $p < 0.05$ ) among means at the same experimental times (and among the slope values) was checked by one-way ANOVA using GraphPad Prism 7 (GraphPad Software, San Diego, CA). A HSD-Tukey test was run for post-hoc comparisons. The Shapiro-Wilk test was used to check data normality. Plant length increment, Fv/Fm and Pindex data were analysed by fitting linear mixed-effects models (LMMs) in a repeated measurement ANOVA design, considering plant identity as a random effect factor to account for the temporal correlation of observations. Time was used as an ordinal variable. The physiological effects of the treatment

were investigated using plant length increment, Fv/Fm and Pindex values as response variables and the treatments with IONPs as explanatory variables in a full factorial design. ANOVA type III table, with Satterthwaite's method, was used to prove the significance of the fixed effects and associated interaction factors. The analysis was run with the software R and LMM computations were performed using the lmer function of the lme4 package version 1.1–12 for model fitting.

### 3. Results

#### 3.1. IONPs characterization

TEM images of magnetite SNPs and LNPs and their size distribution fitted by a Log-normal law are reported in Fig. 1A and B, respectively. As shown from these histograms, the two sets of nanoparticles were both monomodal and therefore suited to report on the effects induced by pollutants within a defined size range. Moreover, the average diameters of the two distributions were representative of the most commonly found for inorganic nanoparticles of metal oxides, with the largest particles having wider polydispersity. TEM images showed that the magnetite SNPs were predominantly spherical in shape with average diameter  $d_0 = 13$  nm and  $\sigma = 0.24$ . By comparison, the LNPs showed marked variations in their morphology. In this case, the best-fit curve gave a (mean) diameter distribution centered at  $d_0 = 230$  nm and  $\sigma = 0.38$ .

#### 3.2. Impact of MNPs on plant growth and photosynthetic parameters

Concerning growth, control plants showed an increment in length over time (Fig. S1), whereas exposure to IONPs caused a significant reduction of plant length at any of the investigated exposure times ( $265.9 < F < 703.7$  and  $P < 0.0001$ ). At the end of the treatment, the increment in length was significantly different also among particle-exposed specimens, with HC-SNPs plants showing the lowest values (Fig. 2A). The increment variation depended significantly on the IONPs used (treatment\*time significant with  $P < 0.001$ , Table S1). As the regression analysis gave significant results for the correlation between increment in length and sampling time (Fig. 2B–S1), the angular coefficient (slope) of this regression line was used to estimate the growth rate itself as in Falsini et al. (2022). This coefficient indicated significant differences among treatments ( $F = 216.5$ ,  $P < 0.0001$ ). All the values obtained for exposed specimens were lower than for control plants, having similar values in the case of LC-LNP, HC-LNP and LC-SNP samples and the lowest one for HC-SNP samples.

Concerning chlorophyll fluorescence, the Fv/Fm and Pindex values in control and treated samples over the whole exposure time are reported in Figs. S2 and S3, respectively. One-way ANOVA did not show any significant change among the different treatments at any time, despite an incipient lowering of Fv/Fm values occurring in treated plants toward the end of the experiment. The interaction treatment\*time was not significant either (Tables S2 and S2).

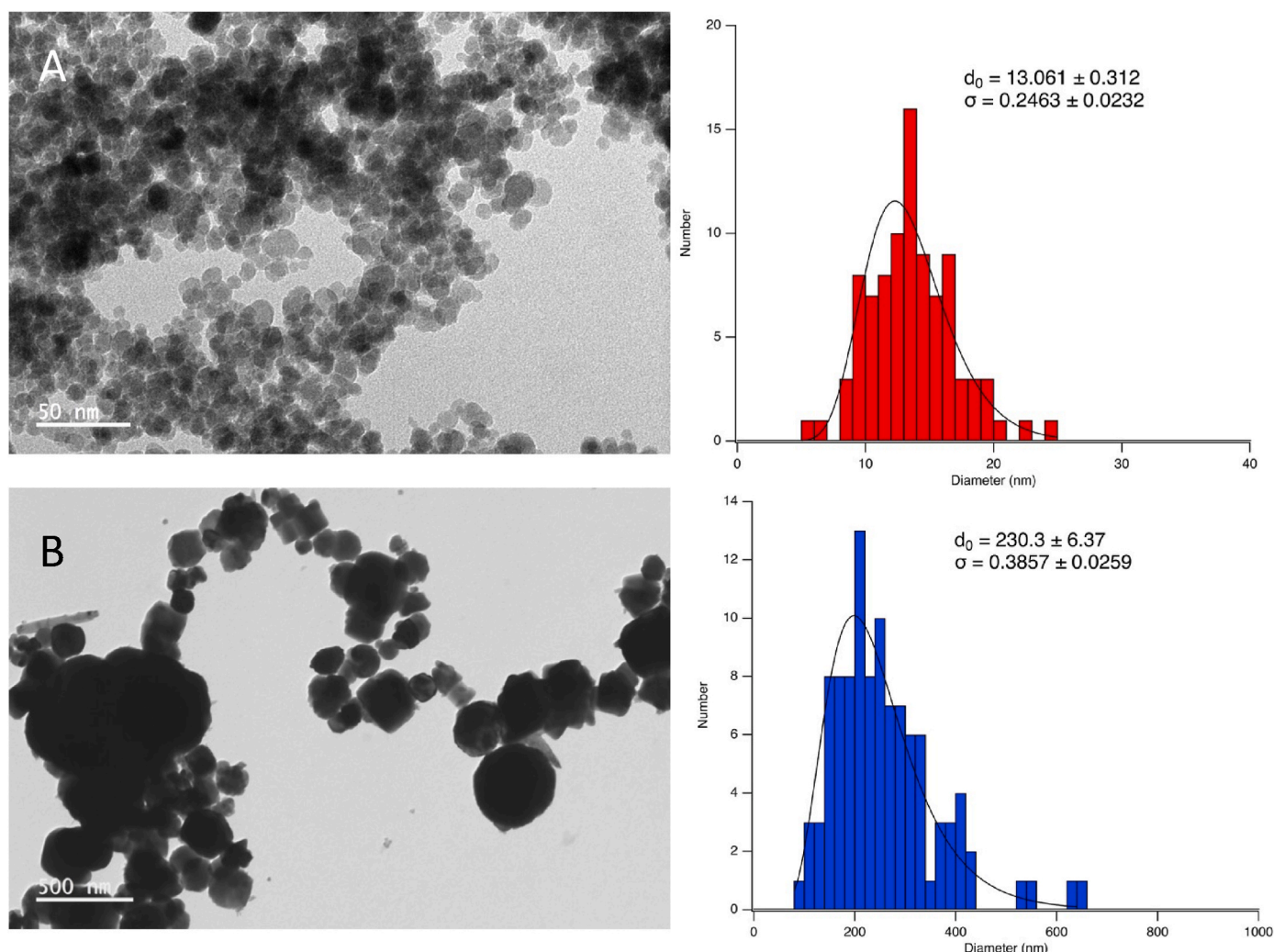
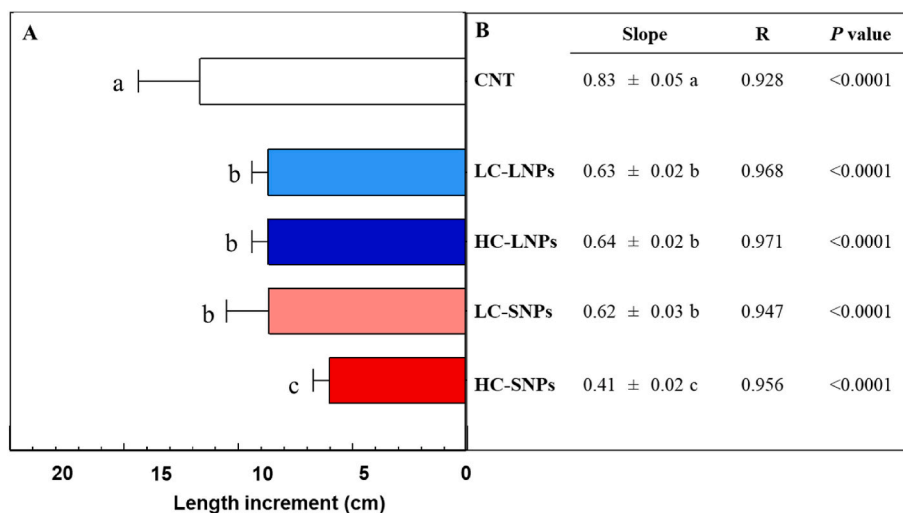


Fig. 1. Representative TEM micrographs of IONPs and their relative size distribution fitted by a Log-normal law with  $d_0$  and  $\sigma$  as parameters. (A) SNPs (B) LNPs.



**Fig. 2.** A) Final increment in length (cm) of *Tillandsia usneoides* strands exposed to four types of NPs for 14 days. The increment was calculated by subtracting length values at the beginning of the treatment to the values measured at the end of the treatment. Letters close to the histograms indicate significant differences among treatments, according to Tukey's test (at least  $p < 0.01$ ). The reported values are means of 9 replicates + standard deviation. B) Growth rate of *T. usneoides* strands exposed to four types of NPs for 14 days calculated as the slope of the linear regression between plant length increment and sampling time.

### 3.3. IONP effect on the concentration of macro- and microelements

Table 2 reports the effects of the type and dose of IONPs on the concentration of some macro- and micronutrients in plant specimens. Differences among the treatments were highlighted by one-way ANOVA for half of the elements tested ( $0.5925 < F < 8.122$  and  $p < 0.0001$ ). In particular, IONPs exposure did not significantly alter the plant levels of Potassium (K), Calcium (Ca) and Copper (Cu), whereas all the treatments induced a significant decrease in Zinc (Zn) and Magnesium (Mg) accumulation, with the exception of the lowest MNP concentration. A general increase was found in the case of Iron (Fe), significantly for HC-LNPs exposure, and in the case of Manganese (Mn), significantly for HC-SNPs exposure.

The amplitude of the treatment-induced changes in element concentration depended both on the IONP-administered and the element under consideration, as evidenced by the heatmaps (Fig. 3), where the red colour indicates samples with decreased element concentration with respect to the control and the green colour indicates samples with increased element concentration with respect to the corresponding control. In general, the variations were similar for both kinds of IONPs, with no evident dose-dependency and with remarkable intensity only for Fe and Zn.

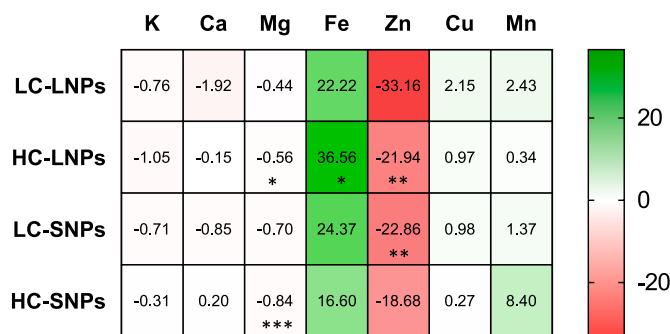
### 3.4. Microscopy on *Tillandsia usneoides*

The investigation performed by Scanning Electron Microscopy (SEM) showed that the trichome wings covered completely the epidermis of *T. usneoides* stem (Fig. 4A, B and C) in all the samples. Trichomes displayed the shape typical for *Tillandsia*, with a central shield and a wing forming a rhombus, or better a kite, in which an axis is longer than the other. The main axis was indicated here as "length" and the minor axis as

**Table 2**

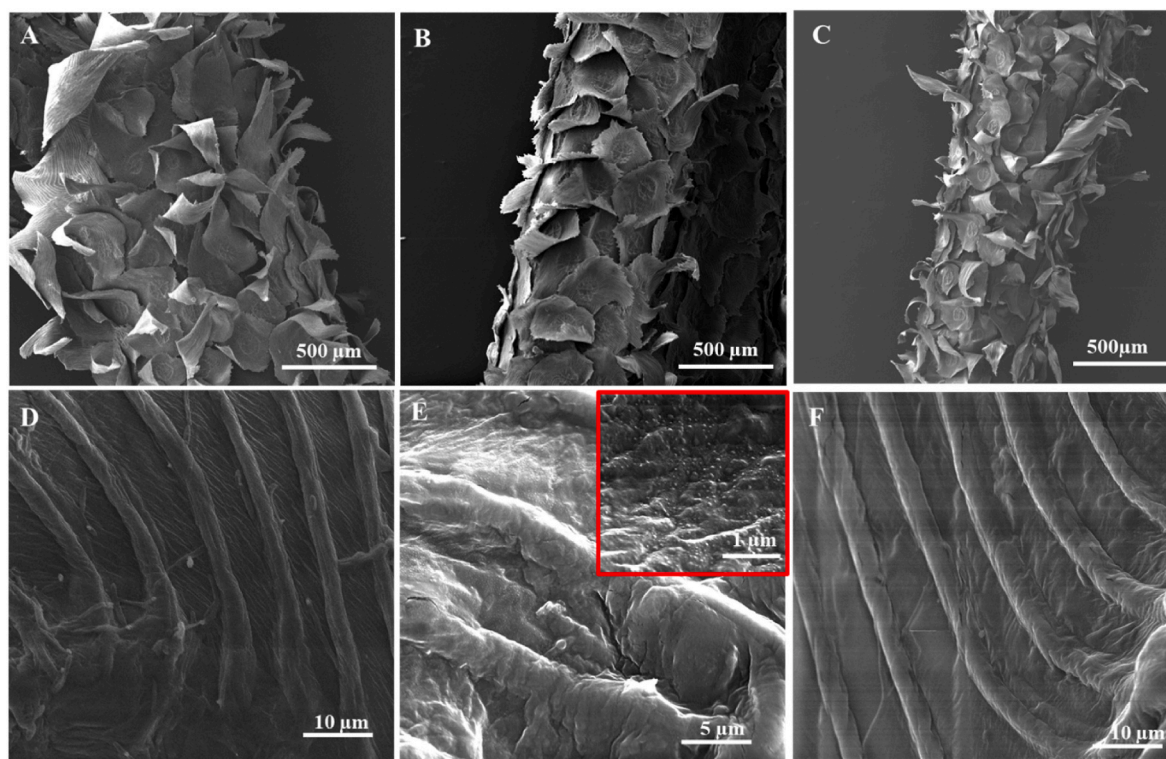
Element concentration ( $\mu\text{g g}^{-1}$  dry weight) in *Tillandsia usneoides* plants after 14 days of growth in the presence of pristine and aged MNPs. Values are mean of 9 replicates ± standard deviation. Asterisks indicate significant differences between treated samples and control samples (\* =  $p < 0.05$ , \*\* =  $p < 0.01$ , \*\*\* =  $p < 0.001$ ).

	K	Ca	Mg	Fe	Zn	Cu	Mn
CNT	11.27 ± 1.10 a	9.49 ± 1.97 a	1.76 ± 0.66 a	64.55 ± 16.07 b	93.75 ± 16.24 a	6.91 ± 1.54 a	31.67 ± 4.17 b
LC-LNPs	10.96 ± 2.01 a	9.28 ± 1.91 a	1.32 ± 0.37 ab	86.77 ± 25.33 ab	60.59 ± 6.51 b	9.07 ± 2.02 a	34.10 ± 4.16 ab
HC-LNPs	10.56 ± 1.14 a	9.13 ± 1.84 a	1.2 ± 0.27 b	101.11 ± 27.87 a	71.81 ± 11.70 b	7.89 ± 1.58 a	32.01 ± 4.20 b
LC-SNPs	10.5 ± 1.29 a	8.44 ± 1.28 a	1.06 ± 0.25 b	88.92 ± 28.45 ab	70.89 ± 13.77 b	7.89 ± 0.88 a	33.04 ± 5.53 b
HC-SNPs	10.22 ± 2.17 a	7.37 ± 2.2 a	0.92 ± 0.32 b	81.15 ± 21.83 ab	75.07 ± 13.30 b	7.18 ± 2.06 a	40.07 ± 6.50 a



**Fig. 3.** Heatmaps of ionome variation in strands *Tillandsia usneoides* plants after 14 days of growth in the presence of LC-LNPs, HC-LNPs, LC-SNPs and HC-SNPs. Colour scale indicates increased (green), unchanged (white) or decreased (red) element concentration in respect to the corresponding control plants. Asterisks indicate significant differences between treated samples and control samples referred to Table 1 (\* =  $p < 0.05$ , \*\* =  $p < 0.01$ , \*\*\* =  $p < 0.001$ ). (For interpretation of the references to colour in this figure legend, the reader is referred to the Web version of this article.)

"width". The most distal part of the wing was recurved outwards (Fig. 4A, B and C). Table 3 reports the average dimensions of the trichomes after exposure to the highest concentration of the IONPs. Control plants showed a trichome size of about  $500 \times 250 \mu\text{m}$ , with a mean shield diameter of around  $85 \mu\text{m}$ . The treated samples showed a significant reduction of hair length ( $F = 66.64$ ,  $p < 0.0001$ ) for both kinds of IONPs, while the other parameters were not altered with respect to the control values. In control samples, no IONPs were detected (Fig. 4D), while many tiny dots smaller than  $100 \text{ nm}$  appeared on the trichome wings of plants treated with SNPs and could be interpreted as magnetite



**Fig. 4.** SEM micrographs of *Tillandsia usneoides* morphology (new leaves) after 14 days of exposure to SNPs and LNPs. A) Control plants: trichomes covering completely the epidermis of the leaf. The trichomes are formed by a central shield (arrows) of 4 cells surrounded by 2 circles of cells and a large external wing (arrowheads) formed by dead cells. The wings overlap one with the other. B) plants exposed to SNPs. C) plants exposed to LNPs. D) Control plants: detail of the trichome wing. E) Detail of the trichome wing exposed to SNPs. Tiny dots smaller than 100 nm appear on the surface of the trichome. F) Detail of the trichome wing exposed to LNPs. Particles are not visible, but the surface of the wing appears corrugated.

**Table 3**

Dimension of *T. usneoides* trichomes in plants exposed for 14 days to small and large IONPs at the highest concentration. Values are mean of 9 replicates  $\pm$  standard deviation. Lowercase letters indicate the significant differences among the treatments according to Tukey's test (at least  $p < 0.05$ ).

	Hair length ( $\mu\text{m}$ )		Hair width ( $\mu\text{m}$ )		Shield diameter ( $\mu\text{m}$ )	
Control	523	$\pm$ 36b	259	$\pm$ 33 ab	85	$\pm$ 11a
HC-SNPs	335	$\pm$ 39a	267	$\pm$ 36b	89	$\pm$ 13a
HC-LNPs	362	$\pm$ 37a	219	$\pm$ 35a	81	$\pm$ 7a

particles (Fig. 4E). Even though the surface of plants exposed to LNPs did not present any particle evidence, the exposure to these contaminants made the surface of the trichome wings appear peculiarly corrugated (Fig. 4F).

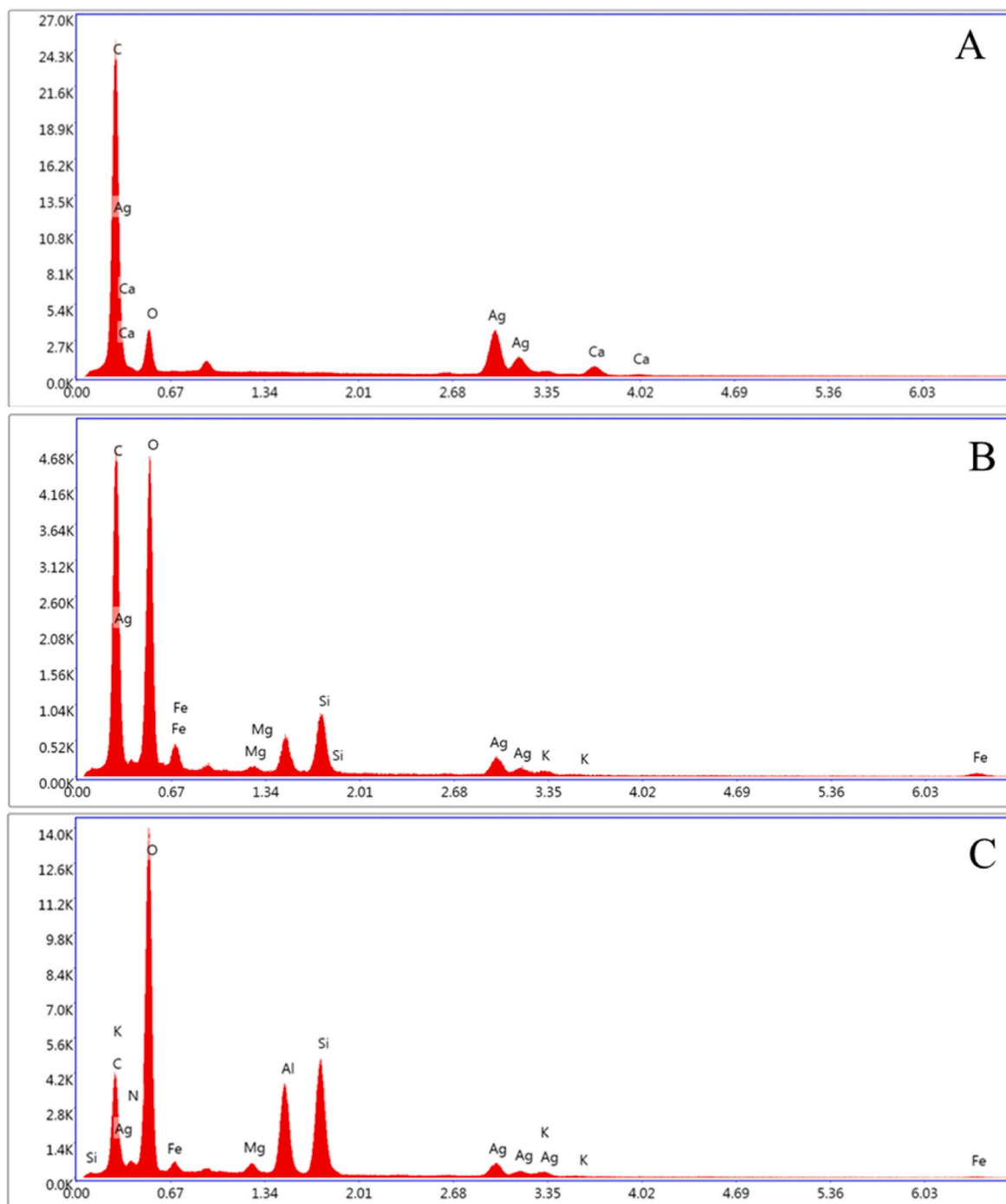
Microanalysis was performed on the trichome wing surface. As expected, while it did not record Fe in the control plants, both SNPs and LNPs treated samples showed a peak of iron (Fig. 5).

#### 4. Discussion

All the experimental evidence reported in this work showed that airborne IONPs at environmentally relevant concentrations impaired *T. usneoides* growth, significantly already from the first exposure time for both types of particles tested and with a dose-dependent effect only for the smaller ones. These also appeared more toxic than their larger counterparts, almost halving both plant length increment and growth rate when present at high concentrations. This marked toxicity shown by small magnetic particles and their concentration effect can be explained by a higher ability of small IONPs to enter the plant tissues and catalyse processes with negative influence on the plant health. The dimension of *Tillandsia* stomata around 50  $\mu\text{m}$  (Martin and Peters, 1984) would allow

the entry of both the nanoparticles here tested, but the thick layer of overlapping trichomes could represent a barrier. More probably, our  $\sim 13$  nm IONPs were small enough to directly penetrate the plant body through possible discontinuities of the plant surface. Similar results, showing higher detrimental effects of small IONPs, supposedly due to their absorption, were also observed by Alkhatib et al. (2019) on tobacco plants exposed to particles in the same size range, even though administered to roots in hydroponics and therefore not directly comparable. Considering that in many contexts the main source of IONPs contamination comes directly from the atmosphere through road traffic, with a minor contribution from other anthropogenic activities (Magiera et al., 2011), our results point out the most critical matrix for studying magnetite nanoparticle administration to plants.

Regarding possible IONP toxicity mechanisms via limitations of photosynthesis, our data proved that at least for the concentrations and exposure times here tested, the reduction in plant growth could not be explained by a negative interference of airborne magnetite pollution with the photosynthetic process. Conversely, a negative effect of excessive amounts of iron oxide particles on several photosynthetic parameters has been reported when they are administered to roots in hydroponics (Alkhatib et al., 2019; Ursache-Oprisan et al., 2011), thus presenting again the difficulty of direct comparison among experiments with IONPs supplied through different routes. The IONPs impact on *T. usneoides* growth could be better explained if considering the particle-induced changes in the plant element concentration. Both types of IONPs caused the increase of Fe concentration and the decrease of Mg and, especially, Zn concentration. Even though the element levels did not fall outside the range commonly reported for plant shoots (Marschner, 1995; Kabata-Pendias and Pendias, 2001), the cumulative effect of the IONP-induced changes present in treated samples could have resulted in unbalanced development of *T. usneoides* plants. The observed Fe accumulation could derive both from the adhesion and/or



**Fig. 5.** Energy Dispersive X-ray spectroscopy microanalysis on control plants (a), plants exposed to SNPs (b) and plants exposed to LNPs (c).

entering of the IONPs in the plant or from the direct release of the metal from the IONPs, as proposed for trials in hydroponics and soil systems (Iannone et al., 2016; Yan et al., 2020; Deng et al., 2022). In our case, during the contact with the plant surface, the magnetic particles could leak Fe ions in the thin film of water that moves by capillarity among the overlapping trichomes of *Tillandsia* and is taken up by the plant together with the nutrients (Herppich et al., 2019). A consequence of this Fe-enrichment in the water entering the plant could be the observed decrease in Zn accumulation, since these two micronutrients share some common transporters (Colangelo and Guerinot, 2006) and therefore can compete for the same uptake systems. At the same time, Fe excess is known to negatively affect Mg availability to plants (Kirk et al., 2022), thus causing the decrease recorded in the concentration of this other

nutrient in IONP-treated *T. usneoides* strands. Following another study, the changes in the element profile could be explained considering that the IONPs generate a magnetic field which may alter the conductivity of the ion channels (Hughes et al., 2010). In fact, though the magnetic field generated by a nanoparticle is expected to be low, the effect of sharpness and fractal-like shape can play a role, as it is well established in the case of the electric field enhancement exploited by SERS (Boyacka and Le Ru, 2009).

SEM images provided information about the effect of IONPs on macro-morphological changes of the plant surface and their possible presence on the strands. Both kinds of magnetic particles were able to provoke a significant decrease in the length of *T. usneoides* trichomes. Since these structures are involved in water and ion uptake (Brighigna

et al., 2002; Mosti et al., 2008), the IONP-mediated reduction in their dimension could concur to explain the above-reported changes in the plant element profile, even though the mechanism by which the magnetic particles can affect the shape of the trichomes remains to be elucidated. In addition to the reduction in size, LNPs appeared to affect the distension of the wing itself, resulting in a corrugated surface that could have impaired the functioning of the trichomes. Regarding the presence of the particles on the strands, SEM images showed that apparently, only the smaller IONPs were able to adhere to the trichome surface, or at least to remain adhered thereafter the sample preparation, contrarily to the larger IONPs. This supposed stronger adherence of the SNPs to the trichomes might affect their functioning, contributing to determine their higher toxicity with respect to the LNPs. Furthermore, trichomes could represent a route for SNP entrance inside the mesophyll, thus concurring to increase their impact on *T. usneoides*. In any case, the preferential IONP localization was on the trichome wing and not on the central shield, as it was found for particles of plastic materials (Falsini et al., 2022). In addition, the microanalysis of the SEM images revealed the presence of Fe on the surface of the treated samples, even for LNP-treated plants where the detection of adhering particles was not possible. Probably, the enrichment of this element on the plant surface was due to the release of Fe from the particles themselves during their contact with the plant and, as regards SNPs, to the direct presence of the particles on the surface. In any case, even if the total amount of Fe did not exceed the toxicity limits, the enrichment of this element in the cells of the plant surface could have led to their impaired growth thus limiting the development of the whole strand. Actually, high concentrations of Fe are known to induce reductions in plant cell growth, with many toxicity mechanisms including ferroptosis (Distéfano et al., 2021).

## 5. Conclusions

Our results showed that airborne pollution at environmentally relevant IONP concentrations can impair the development of the model plant *Tillandsia usneoides*. Small IONPs were more toxic with respect to large ones; this effect can be traced back to their higher surface/volume ratio, which renders nanomaterials more reactive or prone to release metal ions, as well as to the enhanced ability of tiny objects to penetrate plant tissues. The observed hamper in plant growth was related to changes in the micro- and macro-elemental profile, though not in photosynthetic efficiency. IONPs were found to localize on the leaf trichomes and preferentially on their wings. While previous studies were mainly focused on the effects of IONPs administered by root exposure, thus taking into account contamination from water and soil, here we showed that also airborne magnetic particles can represent a threat to plant life.

This study draws attention to the possible harmful effects of airborne contaminants on plants, raising concerns about additional risks for the whole ecosystem. Further research should be undertaken not only to widen the investigated pollutants and species to complete the picture of this environmental issue, but also to devise strategies for its attenuation.

## CRediT authorship contribution statement

**Sara Falsini:** Writing – review & editing, Methodology, Investigation, Formal analysis, Data curation, Conceptualization. **Ilaria Colzi:** Writing – review & editing, Supervision, Investigation, Formal analysis, Data curation, Conceptualization. **Marco Dainelli:** Writing – review & editing, Investigation, Formal analysis. **Elia Parigi:** Writing – review & editing, Methodology, Formal analysis. **Maria Cristina Salvatici:** Methodology, Investigation. **Alessio Papini:** Writing – review & editing, Resources, Data curation. **Delphine Talbot:** Writing – review & editing, Resources, Methodology, Formal analysis. **Ali Abou-Hassan:** Writing – review & editing, Resources, Methodology, Formal analysis. **Cristina Gonnelli:** Writing – review & editing, Writing – original draft, Supervision, Resources, Project administration, Methodology, Funding

acquisition, Conceptualization. **Sandra Ristori:** Writing – review & editing, Supervision, Resources, Project administration, Methodology, Funding acquisition, Conceptualization.

## Declaration of generative AI and AI-assisted technologies in the writing process

The authors declare no use of AI and AI-assisted technologies in the writing process.

## Declaration of competing interest

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: Sara Falsini, Ilaria Colzi, Marco Dainelli reports financial support was provided by Ministry of Education and Merit. If there are other authors, they declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

## Data availability

Data will be made available on request.

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## Appendix A. Supplementary data

Supplementary data to this article can be found online at <https://doi.org/10.1016/j.chemosphere.2024.141765>.

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